POLITECNICO DI TORINO Repository ISTITUZIONALE

Modelling of technical, environmental, and economic evaluations of the effect of the organic loading rate in semi-continuous anaerobic digestion of pre-treated organic fraction

Original Modelling of technical, environmental, and economic evaluations of the effect of the organic loading rate in semi- continuous anaerobic digestion of pre-treated organic fraction municipal solid waste / Demichelis, Francesca; Robotti, Elisa; Deorsola, FABIO ALESSANDRO; Marengo, Emilio; Tommasi, Tonia; Fino, Debora In: ENVIRONMENTAL POLLUTION ISSN 0269-7491 344:(2024). [10.1016/j.envpol.2024.123417]
Availability: This version is available at: 11583/2989124 since: 2024-05-29T17:47:07Z
Publisher: ELSEVIER SCI LTD
Published DOI:10.1016/j.envpol.2024.123417
Terms of use:
This article is made available under terms and conditions as specified in the corresponding bibliographic description in the repository
Publisher copyright

(Article begins on next page)

Modelling of technical, environmental, and economic evaluations of the effect of the organic loading rate in semi-continuous anaerobic digestion of pre-treated organic fraction municipal solid waste

Demichelis Francesca, Robotti Elisa, Deorsola Fabio Alessandro, Marengo Emilio, Tommasi Tonia, Fino Debora

PII: S0269-7491(24)00131-3

DOI: https://doi.org/10.1016/j.envpol.2024.123417

Reference: ENPO 123417

To appear in: Environmental Pollution

Received Date: 25 October 2023

Revised Date: 18 December 2023

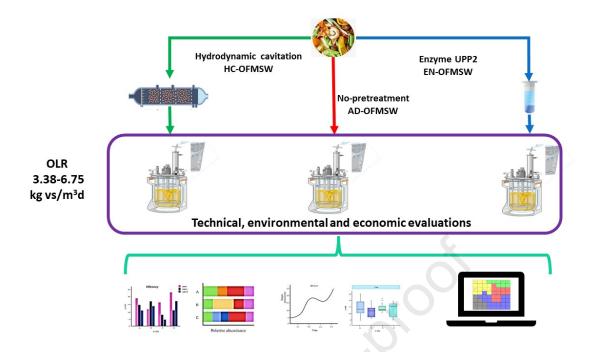
Accepted Date: 19 January 2024

Please cite this article as: Francesca, D., Elisa, R., Alessandro, D.F., Emilio, M., Tonia, T., Debora, F., Modelling of technical, environmental, and economic evaluations of the effect of the organic loading rate in semi-continuous anaerobic digestion of pre-treated organic fraction municipal solid waste, *Environmental Pollution* (2024), doi: https://doi.org/10.1016/j.envpol.2024.123417.

This is a PDF file of an article that has undergone enhancements after acceptance, such as the addition of a cover page and metadata, and formatting for readability, but it is not yet the definitive version of record. This version will undergo additional copyediting, typesetting and review before it is published in its final form, but we are providing this version to give early visibility of the article. Please note that, during the production process, errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

© 2024 Published by Elsevier Ltd.





- 1 Modelling of technical, environmental, and economic evaluations of the effect of the organic
- 2 loading rate in semi-continuous anaerobic digestion of pre-treated organic fraction municipal
- 3 solid waste.
- 4 Demichelis Francesca ^a, Robotti Elisa ^b, Deorsola Fabio Alessandro ^a, Marengo Emilio ^b, Tommasi
- 5 Tonia^a, Fino Debora ^a.
- 6 ^aDepartment of Applied Science and Technology (DISAT), Politecnico di Torino, Corso Duca degli
- 7 Abruzzi 24, 10129 Torino, Italy
- 8 bDepartment of Sciences and Technological Innovation, University of Piemonte Orientale, Viale
- 9 Michel 11, 15121 Alessandria, Italy
- 10 corresponding author: Demichelis Francesca francesca.demichelis@polito.it
- 11 Keywords: semicontinuous anaerobic digestion, pre-treatments, carbon footprint, economic
- evaluation, Principal Component Analysis, Kohonen neural networks

13 Abstract

- 14 The study concerned technical feasibility, economic profitability, and carbon footprint (CF) analysis
- of semi-continuous anaerobic digestion (sAD) of organic fraction of municipal solid waste
- 16 (OFMSW). The research assessed the pre-treatment effect on sAD by varying organic loading rates
- 17 (OLR) from 3.38 to 6.75 kgvs/m³d. Three sAD configurations were investigated: hydrodynamic-
- 18 cavitated (HC-OFMSW), enzymatically pre-treated (EN-OFMSW), and non-pre-treated (AD-
- 19 OFMSW). Principal Component Analysis and Supervised Kohonen's Self-Organizing Maps
- 20 combined the experimental, economic, and environmental evaluations.
- 21 The sAD configurations were grouped predominantly according to the OLR however, within each
- OLR group the configurations were clustered according to the pre-treatments.
- The finding highlighted that pre-treatments offset inhibition in sAD of OFMSW due to the OLR
- increase, being economically profitable and CF negative up to 4.50 kgvs/m³d for EN-OFMSW and
- 25 to 5.40 kgvs/m³d for HC-OFMSW. Whereas sAD-OFMSW remained economically and
- environmentally viable only up to 3.87 kgvs/m³d.

- 27 HC-OFMSW reached the highest performance. In detail, for HC-OFMSW the NPV and CF ranged
- 28 from 17679.30 to 43827.12 euros and from -51.08 to -407.210 kg CO₂eq/1 MWh daily produced, by
- 29 decreasing the OLR from 5.40 to 3.87 kgvs/m³d.
- 30 These results are fundamental since pre-treatment is usually expensive due to additional energy or
- 31 chemical requirements.

32

33

34

35

36

37

38

39

40

41

42

43

44

45

46

47

48

49

50

51

52

1. Introduction

The European biogas plant market is expected to grow from 1.87 billion euros in 2021 to 3.47 billion euros in 2028 (European biogas plant market, 2023). According to the European Biogas Association, in Europe 27, the use of biogas represents more than 11.5 GW of installed power generation capacity (Power and renewables, 2022). Nowadays, at the industrial scale, biogas is produced from organic waste like agro-waste and organic fraction of municipal solid waste (OFMSW) through anaerobic digestion (AD) (Aslanzadeh et al., 2014). The performance of AD of organic waste provides two benefits: 1) reduction of organic waste, and 2) production of renewable energy vector, and fertilizer. At the industrial scale, anaerobic digestion is performed in continuous feeding mode (sAD), which means a system fed continuously or semi-continuously with an average residence time of the substrate in the reactor, expressed by the parameter hydraulic residence time (HRT). The HRT refers to the time that the substrate remains in a digester, and it is calculated as the ratio between the volume of the AD reactor (V) and the flow rate of a digester (Q). The parameter HRT is inversely related to the parameter organic loading rate (OLR), which measures the quantity of substrate (kg of volatile solids) used to feed the digester for a unit volume (V) in a day. In sAD, the OLR and HRT play a key role, and their choice is critical. Currently, at the industrial scale, the working OLR represents a problem in the management of the process. In general, high HRT corresponds to low OLR and vice versa, but it is necessary to find a balance between the OLR and the HRT to optimise the sAD efficiency in terms of biogas production, and volatile solid removal (Aslanzadeh et al., 2014), and to optimise the reactor volume (Demirer and Chen, 2005).

The main problem related to the choice of OLR is the fact that an increase in OLR can promote higher 53 biogas production rates, but excessive OLR inhibits the process by decreasing biogas production and 54 methane content. Inhibition of the process caused by OLR consists of over-acidification and foaming. 55 Acidification in AD is a fundamental phase since AD is divided into four phases (hydrolysis, 56 acidification, acetogenesis, and methanogenesis) but over-acidification for excessive OLR leads to a 57 pH drop due to volatile fatty acid accumulation, that reduces the methane production and the quality 58 of the digestate. OLR depends on the type and concentration of the substrate, AD temperature, and 59 biogas digester. 60 To date, three macro-types of studies have been performed to investigate the effect of OLR on AD on 61 62 different substrates. The first type concerns the effect of the OLR variation on the efficiency of the biogas production from different substrates. Specifically, Babaee and Shayegan, (2016) studied the 63 OLR between 1.4 -2.75 kgvs/m³d on vegetable waste identifying an optimal OLR at 10.80 kgvs/m³d. 64 Zhou et al., (2022) investigated the variation of OLR from 1.00 to 13.80 kgvs/ m³d on food waste 65 with an optimal OLR of 10.80 kgvs/m³d. Moreover, Liu et al.,(2018) varied OLR from 1 to 2.5 66 kgvs/m³d on spirulina waste with an optimal OLR of 2.25 kgvs/m³d. The second type of study 67 concerns the effect of mixing on the OLR variation and on biogas production. In detail, Rog et al., 68 (2023) studied the OLR from 0.38 to 2.31 kgvs/m³d on activated sludge identifying the optimal OLR 69 70 of 1.98 kgvs/m³d. Nges and Liu, 2(010) at 250 rpm varied OLR from 1.6 to 20.5 kgvs/m³d on dewatered sludge with an optimal OLR corresponding to HRT of 30 d. Furthermore Leite et al., 71 (2017) studied the activated sludge at optimal OLR of 1.90 kgvs/m³d withholding mixing 2 h before 72 73 feeding. The third type of study manages the increase of the OLR by performing the two-stage AD. In detail, two-stage AD was investigated by Aslanzadeh et al., (2014) on OFMSW with OLR from 2 74 to 14 kgvs/m³d with optimal OLR of 12 kgvs/m³d. Two-stage AD was investigated by Dareioti and 75 Kornaros, (2014) on olive mill wastewater, cheese whey and liquid cow manure with HRT from 5 to 76 0.75 d, with an optimal OLR of 12.70 kgvs/m³d. Another study concerning the two-stage AD was the 77 study of Wijekoon et al., (2011) on molasses-based synthetic wastewater with OLR from 5 to 12 kg 78

cod/ m³d with an optimal OLR of 8 kg cod/ m³d. All the above-mentioned studies only investigate OLR from a technical point of view. The present study aims to investigate the possibility of managing the increase of the OLR of sAD through the performance of pre-treatment on OFMSW before sAD in one single stage. In detail, this study analyses the effect of enzymatic and hydrodynamic-cavitation pre-treatments performed in (Demichelis, F. et al., n.d.) and (Demichelis et al., 2023), respectively, by increasing the OLR from 3.38 to 6.75 kgvs/m³d corresponding to HRT from 16 to 8 d. The novelty of this study is in not limiting the research to the identification of the most technically high-performing sAD configuration but combining it with the sustainability analysis through statistical elaborations. The three-fold metrics are adopted by including technical feasibility analysis through laboratory tests, economic viability through cost-benefit analysis, and environmental sustainability by carbon footprint analysis (CF) according to ISO 14067. The results of the three-fold metrics are elaborated with Principal Component Analysis and Supervised Kohonen's Self-Organizing Maps. To the best of the author's knowledge, no studies are currently available regarding the adoption of pre-treatment to address the increased OLR in sAD of OFMSW. Additionally, there is a lack of technical, economic, and environmental studies in the literature about the sAD of OFMSW considering the effect of the OLR. Moreover, a standardised assessment of the three-fold sustainability metrics is not available yet but is necessary as stated by (Rajendran and Murthy, 2019).

97

98

99

79

80

81

82

83

84

85

86

87

88

89

90

91

92

93

94

95

96

2. Materials and methods

2.1 Substrate

The organic fraction of municipal solid waste (OFMSW) was supplied by San Carlo S.p.A (Fossano, Italy). To start the semicontinuous anaerobic digestion (sAD) the mesophilic digestate of cowagriculture sludge (CAS) was used as inoculum according to the previous study of (F Demichelis et al., 2022). OFMSW and inoculum properties agreed with the study of (Srisowmeya et al., 2020) and (Gu et al., 2020), respectively (Table 1).

105	All the experiments were performed with the same lot of OFMSW and inoculum, to limit the process
106	variability, since OFMSW strictly depends on season variability.
107	The OFMSW was frozen at +4 °C according to (Zeng et al., 2010) and (Gu et al., 2020) to prevent
108	the natural decomposition of the organic matter content and to avoid the variation of its physic-
109	chemical compositionThen OFMSW was defrost down for the daily feed in the anaerobic digestors.

John Marie President Control

Table 1 Physical and chemical properties of OFMSW and inoculum TS= total solids, VS=volatile solids.

	OFMSW	CAS
TS (%)	6,02 ±0.6	5.8±0.1
VS (%)	97.1±0.5	70.3±1.0
pH (-)	5.5 ±0.2	7.7±0.1
C (%)	48.1±0.5	40.6±0.6
H (%)	6.2±0.7	3.0±0.0
N (%)	3.1 ± 0.3	7.9±0.1
S (%)	0.2±0.1	0.0±0.0
C/N (-)	15.5 ±1.4	5.1±0.1
TOC (g/kg)	24914.6± 114.9	12.0±0.2

2.2 Physical and enzymatic pre-treatments 114 Before semicontinuous anaerobic digestion (sAD), OFMSW was pre-treated with two different pre-115 treatments, which were investigated in two previous studies; hydrodynamic cavitation (HC) 116 (Demichelis et al., 2023) and enzymatic pre-treatment (EN) (Demichelis, et al., n.d.). 117 HC was performed with a rotor/stator HC-unit (Rotocav[®], E-PIC srl – Mongrando, Italy) at 55 °C for 118 10 min (Bruni et al., 2010). HC is a pre-treatment that promotes the formation, growth, and implosion 119 of vapor bubbles in a liquid at temperatures lower than the boiling point, which generates 120 microenvironments characterized by locally very high temperatures and intense pressure waves. 121 122 EN was performed with UltraPract® P2 (UPP2), which is a mix of cellulases, hemicellulases, pectinases, and proteases, and it has been designed for biogas plants treating vegetables. UPP2 was 123 used with a dose of 1 mL/100 g TS in the pH range 7.0 - 7.5 (GmbH, 2022) at 45 °C, (Demichelis, et 124 al., n.d.). 125 The significant differences were calculated with Pearson test $\alpha < 0.05$, which is a correlation test to 126 investigate the presence of the linear correlation between pairs of variables, considering the 127 significance of those having p < 0.05. 128 129 2.3 Anaerobic digestion 130 Three sAD configurations were tested: hydrodynamic cavitated OFMSW (HC-OFMSW), 131 enzymatically pre-treated OFMSW (EN-OFMSW), and non-pre-treated OFMSW (AD-OFMSW) as 132 blank. Each sAD configuration was tested in duplicate for a total of 6 reactors. 133 sAD was performed in a 1 L reactor (Duran, Germany) with a working volume of 0.8 L at 6 %w/w 134 total solid contents. The sAD was performed in mesophilic conditions and the temperature was kept 135 at 37 °C with a 55 L water bath (Julabo Corio, C), and the anaerobic condition was guaranteed by 136 purging nitrogen. The top of the reactor had three ports: one as the inlet to feed the pre-treated and 137 non-pre-treated OFMSW in the reactor, the second to remove the digestate, and the third to collect 138 the biogas into a Tedlar (Germany) 2 L gas bag. The tested organic loading rates (OLRs) were selected 139

by changing the hydraulic retention time (HRT) from 16 to 8 d, which means OLR variations between 3.38 and 6.75 kg vs/ m³ d. Each HRT was maintained at least for a time equal to two HRTs to allow the achievement of the pseudo-steady state. The starting HRT and the consequential OLR were selected according to the results obtained in the AD of OFMSW performed in batch feeding mode (F Demichelis et al., 2022). Biogas was measured with the water displacement method and its composition was evaluated through SRA Micro-GC, which includes a Molsieve 5A column (for the analysis of permanent gases like hydrogen, nitrogen, methane, and carbon monoxide) employing argon as a carrier (column temperature: 100 °C) and a TCD detector. The injection temperature was 90 °C and the pressure was 30 psi. To evaluate the quality of the sAD process, the pH was daily measured with a pH340 WTW pH-meter (Mettler Toledo, Germany) according to DIN 38404 C5 methodology.

- 151 At the end of each tested HRT, the volatile solid (VS) was removed and volumetric biogas productions
 152 were measured to evaluate the performances of sAD.
- The VS removed was evaluated with Eq. 1 (Li et al., 2018).

154
$$VS\ removed\ (\%) = 1 - \frac{VS\ output\cdot (1-VS\ input)}{VS\ input\cdot (1-VS\ output)}$$
 (1)

where VS removed was the removed volatile solids (%), VS input and VS output were the volatile solids concentrations in the OFMSW at the beginning and end of sAD.

2.4 Economic analysis

- The economic study was performed based on a cost-benefit analysis to evaluate the cost-effectiveness and to identify the barriers to implementing a cost-effective strategy for the 15 sAD investigated (3 AD configurations x 5 OLR). A study estimate was adopted as a preliminary estimate of capital and operational costs, and possible revenues considering the main equipment included in the process.
- A detailed description of the cost-benefit analysis methodology is reported in Section 2 of SI.
- The cost-benefit analysis as well as the carbon footprint analysis (paragraph 2.5) referred to the sAD plant with a capacity of production of 1 MWh/d of primary energy. The process was based on the data

- obtained at the laboratory scale, but it was scaled up at the industrial level through assumption and
- 167 corrective factors based on (Bruno et al., 2023) (Green and Southhard, 2019) and (Turton et al., 2018).
- The process flow diagram consisted of four phases: the first is the pre-treatment unit (enzymatic and
- hydrodynamic cavitation pre-treatments), the second one is the sAD, the third phase is the CHP unit,
- and the fourth one is the composting unit.
- 171 CH₄ yields of the 15 sAD configurations tested at the laboratory scale were scaled according to (Green
- and Southhard, 2019) and (Kowalczyk et al., 2011).
- 173 (Green and Southhard, 2019) suggested assuming a scale factor equal to 90 % of the lab scale values.
- Moreover, (Kowalczyk et al., 2011) highlighted that the scaling from a laboratory to a larger scale is
- acceptable but variations in deviation depend on the organic loading rate. The corrective factors for
- 176 CH₄ were derived by assuming a 10% reduction with each one-unit increase in the OLR. These factors
- were established in alignment with the correction factor determined from the study of an AD plant
- simulated using Aspen (Aui et al., 2019) and (Alfonso-Cardero et al., 2021). By increasing the OLR
- the corrective factors were more and more restrictive due to the increase of viscosity and high
- difficulties in mixing.
- In the present study, it has been tested; HRT = 16 d (OLR = $3.38 \text{ kgvs/m}^3 \text{d}$), HRT= 14 d (3.87 kgvs)
- $/m^3d$), HRT = 12 d (OLR = 4.50 kgvs $/m^3d$), HRT = 10 d (OLR = 5.40 kgvs $/m^3d$), and HRT = 8 d
- 183 (OLR = $6.75 \text{ kgvs/m}^3 \text{d}$).
- The corrective factors were as follows: 90 % of the OLR from 3.38 to 3.87 kgvs/m³d, 80 % for OLR
- $= 4.50 \text{ kgvs/m}^3 \text{d}$, 70 % for OLR = 5.40 kgvs/m³d, and 60 % for OLR 6.75 kgvs/m³d, based on the
- 186 CH₄ specific production values at the laboratory scale. The CH₄ yields, obtained by adopting the
- scale-corrective factor, were comparable to the ones available for the industrial plants of OFMSW
- 188 (plant available in Piedmont, Italy) and industrial scale co-digestion of agro-waste according to (Naqi
- 189 et al., 2019).
- The turbine was designed to treat 1 MWh/d and it was assumed equal to 50 kW in line with the study
- of (Huiru et al., 2019). The composting unit was designed considering a residence time of 90 d

- according to (Evangelisti et al., 2014). The area of the composting unit is 4 times the height of the
- composting unit, according to (Ennio, 2018) and a working volume of 70 % was assumed.
- The detailed methodology to calculate capital and operational cost (Table S1-S4) and to design the
- dimension of the equipment (Table S5) is provided in Section 1 of SI through equations and tables.
- The cost-benefit analysis was referred to in 2023. The profitability of the tested 15 sAD configurations
- 197 (Table S6) was evaluated with the net present value (NPV) (Eq.3) and payback time (PBT) (Eq.4).
- NPV quantified the profitability of the sAD configurations considering a plant lifetime equal to 20 y
- and considering a 5% discount for the future cash flows referring to the present value (Pleissner et
- 200 al., 2016).
- NPV > 0 means that sAD process is profitable.

202
$$NPV (euro) = \sum_{t=1}^{T} \frac{C_t}{(1+d)^t} - C_0$$
 (3)

- where, C_0 was the initial capital investment, C_t was the net cash flow during period t, d was the
- 204 discount rate, and t was the sAD plant lifetime.
- The PBT referred to the amount of time it took to recover the cost of an investment. It was calculated
- 206 with (Eq.4)

$$PBT(y) = \frac{c_0}{\text{Net cash flow per period}} \tag{4}$$

208

209

2.5 Environmental analysis

- 210 The carbon footprint (CF) analysis was performed according to ISO 14067, with the database
- 211 Ecoinvent 3.5 and the software SimaPro 9.5.02. The CF analysis compared the global warming
- potential (GWP) of the 15 (3 pre-treatments x 5 OLR) tested sAD configurations focusing both on
- 213 the pre-treatment and variations of OLR. To compare these 15 sAD configurations, the functional unit
- 214 (FU) was assumed equal to 1 MWh/d of produced primary energy according to (Bruno et al., 2023).
- 215 The adopted approach was from grave to gate according to (Ugwu et al., 2022), which means from
- 216 the pre-treatment of OFMSW (grave) to bioenergy production (gate) (as reported in Figure S1). The

217	collection and transport of OFMSW to the AD plant were not considered since were the same for all
218	the 15 sAD configurations. The CF-analysis (ISO 14067) is geographical, and time referred. The
219	present study was geo-referred to Italy and in detail in the northwest of Italy, where San Carlo SpA
220	(Fossano, Piedmont, Italy) is located. San Carlo SpA is the waste treatment plant that supplied the
221	OFMSW for the present study. The study was time-referred to in 2023, as well as the cost-benefit
222	analysis. A detailed description of CF-analysis methodology is reported in Section 2 of SI.
223	In the present study, only the direct consequences of sAD of OFMSW were considered, whereas the
224	environmental impacts of the infrastructures and capital goods were excluded (Thushari et al., 2020),
225	because they were less important to the overall results.
226	The sAD included the foreground and background systems in agreement with (Clift et al., 2000). The
227	foreground system was directly involved with the reference flow management and the background
228	system considered energy production and chemical supply (Thushari et al., 2020).
229	According to (Piccinno et al., 2016), the laboratory process provides only limited indication of the
230	possible environmental impacts. Moreover, (Carlqvist et al., 2022) stated that to better understand the
231	environmental impact of the future system it is needed to consider an upscaled. The Life Cycle
232	Inventory (LCI) was based on primary data obtained at the laboratory scale but then these data were
233	scaled according to the scaling up explained for the cost-benefit analysis (paragraph 2.4 and S2 in
234	SI).
235	The same boundary conditions were assumed for the cost-benefit and CF analysis to make consistent
236	the considerations derived from the study of economic and environmental sustainability. The CF
237	analysis included Scope 1 and Scope 2. Scope 1 concerns the direct emissions from the pre-treatments
238	and sAD units that occurred from fuel combustion for the digester start-up, biogas combustion in the
239	CHP section to provide heat and electricity, and movement of the OFMSW and digestate in the plant.
240	Biogas combustion is generally considered carbon neutral since the emitted carbon was previously
241	absorbed from the atmosphere. However, to avoid double counting biogas combustion is considered
242	a positive emission since the emitted carbon was previously absorbed from the atmosphere (Aui et

al., 2019). Moreover, the fugitive CH₄ emissions during AD and CHP stages are considered equal to 243 1.5% of the CH₄ produced (Aui et al., 2019) 244 The electricity and heat generated by CHP unit and the compost were considered avoided emissions 245 aligned with (Carlsson et al., 2015).. The flows and the inventory data are reported in Table S7 and 246 S8, respectively. 247 OFMSW was considered a zero burden according to (Lamnatou et al., 2019). The low heating value 248 of the CH₄ was assumed to be 9.94 kWh/Nm³ according to (Rillo et al., 2020). 249 Life cycle impact assessment was performed with IPCC 2021 GWP 100 V0.1 which contains the 250 global warming potential climate with change factors of IPCC with a timeframe of 100 y. The 251 consistency of the results was proven through a sensitivity analysis by varying the CH₄ yields about 252 \pm 5% volume according to (F. Demichelis et al., 2022). 253

254

255

2.6. Principal Components Analysis

Multivariate data analysis was carried out by two tools: pattern recognition through Principal 256 Component Analysis (PCA) (Massart, 1988) and classification using Supervised Kohonen's Self-257 Organizing maps (SKSOMs) (Melssen et al., 2006) (Brandi et al., 2021). The detailed methodology 258 of PCA and SOMs is reported in section 3 of SI. PCA provides the scores and the loadings. From the 259 score plot, the existence of groups of samples with similar or different behaviour is derived while 260 from the loadings plot, the correlations between the variables are highlights. Here, PCA was applied 261 after autoscaling (mean centering and normalization to unit variance) for a preliminary exploration 262 of the dataset. 263 Kohonen's Self-Organizing Maps (KSOMs) are artificial neural networks (i.e., mathematical 264 algorithms) for solving complex problems by simulating the human brain functioning. Supervised 265 Kohonen networks (SKN) are supervised methods for classification purposes (Melssen et al., 2006) 266 (Brandi et al., 2021). Here, SKNs were run with the following settings: non-toroidal boundary, batch 267 algorithm, squared topology, random initialization of weights, and learning rate decreasing linearly 268

from 0.5 to 0.01, a top map of 6 x 6 neurons, and 300 training epochs. To highlight the differences 269 between the HRT values (Y variable), the data were first centered according to the different pre-270 treatments and then range-scaled. The calculations were performed in cross-validation with Venetian 271 blind with 6 cancellation groups. PCA was carried out by MATLAB R2014a (The Mathworks, Natick, 272 MA, USA) using in-house-developed routines; Kohonen SOMs were built with the Kohonen and 273 CPANN toolbox for MATLAB from Milano Chemometrics (Ballabio et al., 2009). Graphical 274 representations were carried out by MATLAB, Statistica v.7 (Statsoft Inc., Tulsa, OK, USA), and 275 Excel 2016 (Microsoft Corporation, Redmond, WA, USA). 276

277

278

279

3. Results and discussion

- 3.1 Anaerobic digestion in semicontinuous feeding mode
- 280 The semi-continuous anaerobic digestion (sAD) of organic fraction of municipal solid waste
- 281 (OFMSW) was performed in 1 L reactors, to detect the optimal working condition by varying the
- OLR. Three sAD configurations were tested: hydrodynamic cavitated OFMSW (HC-OFMSW),
- enzymatically pre-treated OFMSW (EN-OFMSW), and non-pre-treated OFMSW (AD-OFMSW) as
- reference. Figure 1 depicts the methane yields, the pH trends, and the volatile solids (vs) removal of
- 285 the three sAD configurations at different OLRs.
- Five OLRs were tested, by changing the HRT. In detail it has been tested; HRT = 16 d (OLR = 3.38
- 287 kgvs/m³d), HRT= 14 d ($3.87 \text{ kgvs/m}^3\text{d}$), HRT = 12 d (OLR = $4.50 \text{ kgvs/m}^3\text{d}$), HRT = 10 d (OLR =
- 288 5.40 kgvs/m³d), and HRT = 8 d (OLR = 6.75 kgvs/m^3 d).
- Considering the results of Figure 1, two observations can be made: the first about the variation of
- OLR and the second one about the effect of pre-treatments.
- Increasing OLR from 3.38 to 4.50 kgvs/m³d (HRT from 16 to 12 d) enhanced methane productions
- and VS removals, whereas the increase of OLR from 5.40 to 6.75 kgvs/m³d inhibited the sAD
- configurations as demonstrated by the decrease of methane productions and the drop of pH values.

- 294 HC- and EN-OFMSW reached higher methane production and VS removal values rather than sAD-
- OFMSW. Despite the increase in the OLR, pre-treatments can keep the anaerobic digestion process
- stable and efficient since they increase the digestibility rate of the substrate by preventing the
- inhibition effects (Wei et al., 2022).
- Among the pre-treated sAD configurations, HC-OFMSW reached the highest methane production
- and process stability.
- Considering the variation of OLR from 3.38 to 4.50 kgvs/m³d, methane production of HC-OFMSW
- rose from 0.29 to 0.38 Nm³/ kgvs d, whereas for EN-OFMSW and AD-OFMSW varied only from
- 302 0.24 to $0.32 \text{ Nm}^3/\text{ kgvs d}$ and from 0.19 to $0.25 \text{ Nm}^3/\text{ kgvs d}$, respectively.
- The increase of OLR from 5.4 to 6.75 kgvs/m³d (HRT from 10 to 8 d) decreased methane production
- for all three sAD configurations, but HC and EN pre-treatments buffered the inhibitory effects.
- The effects of the variations of OLR were demonstrated by the pH and VS removal trends (in Figure
- 306 1). The pH provides information about the stability of the reaction medium since its variation is
- related both to the buffering capacity of the reaction system and to the variation in the equilibrium
- between the species, which participate in the trophic chain of the microorganisms involved in the
- 309 sAD process.
- For pH values between 6.5 and 7.5, the AD process is generally considered stable (Morales-Polo et
- al., 2018), however, the pH can indicate unbalanced conditions of the sAD only with a certain delay
- compared to the evolution of the buffer effect of the substrate employed in the system.
- For HC, EN, and AD-OFMSW, at OLR from 3.38 to 4.50 kgvs/m³d (HRT from 16 to 12 d), the pH
- was between 7 -7.5, which is the perfect range for AD (Chen et al., 2008), whereas at OLR from 5.4
- to 6.75 kgvs/m³d (HRT from 10 to 8 d), the pH deeply decreased from 6.46 to 5.18, which proved the
- possible volatile fatty acid (VFA) accumulation (Morales-Polo et al., 2018).
- 317 High OLR can cause overloading which boosts faster hydrolysis and over-acidification of the
- medium, by promoting the over-accumulation of VFAs, which inhibits the methanogenesis and
- 319 consequently stops AD process (Meegoda et al., 2018).

320	(Villa and Ferguson, 2016) investigated the effect of the overloaded grease-waste in the AD system,
321	by finding that the quick shocks in the OLR variation could be able to cause shifts in microbial
322	populations, and methane yields returning to normal levels after developing a tolerance to higher
323	OLR. Hence, the study of (Villa and Ferguson, 2016) suggested that by improving the AD resistance
324	to overloading, after the initial overloading, the system can develop diversified methanogenic
325	microorganisms able to improve AD.
326	In the present study, the overloading equal to $6.75~kg_{vs}/m^3d$ inhibited the sAD, but the pre-treatment,
327	especially of HC-OFMSW buffered the system and maintained the methane production and VS
328	removal equal to 0.24 $Nm^3/kg_{vs}d$ and 55.05 %w/w, respectively. These results agreed with the study
329	of (Garuti et al., 2018) which tested OLR around 6 $kg_{vs}\!/m^3d$ and measured, through rheological
330	analysis, the capacity of HC to improve the mixing of the substrate without forming floating matter
331	by increasing the OLR.
332	The VS removal trend was the same for all the three sAD configurations; it increased from OLR 3.38
333	to $3.87 \text{ kgvs/m}^3\text{d}$ (from HRT=16 d to 14 d), then it decreased from OLR = 4.50 to $6.75 \text{ kgvs/m}^3\text{d}$
334	(from HRT=12 d to 8 d).
335	Volatile solid (VS) is a fundamental parameter because it represents the measurement of the organic
336	fraction of total solids and VS removal measures the digester efficiency and the quality of the
337	digestate for further applications (Mei et al., 2016). Considering the sAD configurations, the highest
338	VS removal was achieved by HC-OFMSW, followed by EN-OFMSW, and last AD-OFMSW.
339	In HC pre-treatment, the temperature (55°C) and high pressure prevent evaporation and boost
340	hydrolysis by forming a substrate that is more biodegradable and more stable for AD also at high
341	OLR (Barber, 2016). In HC, the implosion of the bubbles at a temperature of around 50 $^{\circ}\text{C}$ could
342	boost the removal of VS and increase methane production (Calcio et al., 2018).
343	EN-OFMSW was performed with UPP2, which is a cocktail of cellulases, hemicellulases, pectinases,
344	and proteases. UPP2 can boost the AD process by degrading the lignocellulosic wastes(Nabi et al.,
345	2019). The advantage of UPP2, compared to other industrial enzymes is the shorter hydrolysis time

(2 h), moreover, the methane productions and VS removals obtained for OLR from 3.40 to 5.40
kgvs/m³d agreed with the ones described in the research of (Khanh Nguyen et al., 2021).
Considering the methane productions, pH, and VS removals, the most promising OLRs were in the
range of 3.38 and 3.87 kgvs/m ³ d, (HRT 16 and 14 d) for sAD-OFMSW, between 3.87 and 4.50
kgvs/m ³ d (HRT 16 and 12 d) for EN-OFMSW, and in the range of 3.87- 5.40 kgvs/m ³ d, (HRT 16 and
10) for HC-OFMSW. The HC-OFMSW reached the highest methane production and process stability
rather than the EN-OFMSW and AD-OFMSW because thermo-physic pre-treatment at T $<$ 100 °C
can destroy the lignocellulosic barrier without creating recalcitrant molecule (Balasundaram et al.,
2022) and increasing the substrate availability to biochemical reaction (Mohammad Rahmani et al.,
2022).

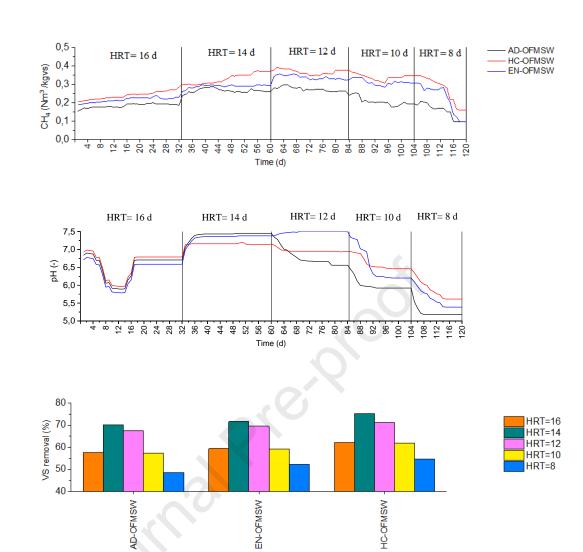


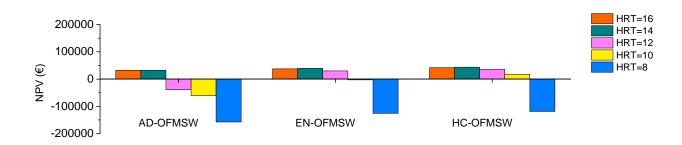
Fig. 1. Methane and pH daily productions and VS removal trends according to the hydraulic retention time (HRT)

363 3.2 Cost-benefit analysis

- 364 The cost-benefit analysis aimed to evaluate the cost-effectiveness and to identify the barriers to
- implementing a cost-effective strategy for the investigated 15 sAD.
- From an economic perspective, the choice of OLR value is difficult to choose because high OLR
- 367 compared to low OLR implies smaller reactor volumes, but larger amounts of digestate to daily
- handle. Fig. 2 depicts the net present value (NPV) of the three sAD configurations in the plant life of
- 20 y that produced 1 MWh/d primary energy. The payback time (PBT) is reported in Table S6 of
- supplementary information. The results achieved by the cost-benefit analysis can be studied based on
- 371 the effect of the OLR and pre-treatments.
- 372 Considering the OLR variation, for all three sAD configurations (AD-OFMSW, EN-OFMSW, and
- 373 HC-OFMSW) the NPV increased from OLR 3.38 to 3.87 kgvs/m³d.
- For all the sAD configurations, the highest NPV and lowest PBT were observed for OLR of 3.87
- kgvs/m 3 d (HRT = 14 d) according to (Choudhary et al., 2020).
- 376 In detail, for OLR equal to 3.87 kgvs/m³d, NPV was 43827.12 € for HC-OFMSW, 39316.77 € for
- 377 EN-OFMSW, and 32587.11 € for sAD-OFMSW.
- 378 The decrease in NPV occurred from 4.50 to 6.75 kgvs/m³d. In detail, at OLR = $6.75 \text{ kgvs/m}^3\text{d}$ all
- three sAD configurations achieved a negative NPV and PBT longer than the chosen plant life (20 y).
- Considering the effect of pre-treatment, the highest NPV and shortest PBT were achieved by HC-
- OFMSW, followed by EN-OFMSW, and last AD-OFMSW.
- For AD-OFMSW, the NPV was positive, and the PBT ranged between 5-6 years, for OLR from 3.38
- 383 to $3.87 \text{ kgvs/m}^3 \text{d}$ (HRT= 16-14 days).
- For EN-OFMSW, the NPV was positive, and the PBT varied from 3-7 years, for OLR from 3.38 to
- 385 $4.50 \text{ kgvs/m}^3 \text{d}$ (HRT= 16-12 days).
- In the case of HC-OFMSW, the NPV was positive, and the PBT ranged between 2-18 years, for OLR
- from 3.38 to 5.40 kgvs/m³d (HRT= 16-10 days).

This trend proved that the pre-treatments could be economic profitability offsetting the inhibition due 388 to the increase of OLR. 389 The possibility of pre-treated AD being economically profitable even at higher OLR was due to the 390 possibility of reducing the working volume of the sAD reactor and consequently, the capital costs 391 decreased. 392 Indeed, the capital costs for EN-OFMSW at an OLR range of 3.87-4.50 kgvs/m³d were lower than 393 those at an OLR of 3.38 kgvs/m³d. Similarly, for HC-OFMSW, the capital costs at an OLR range of 394 3.87-5.40 kgvs/m³d were lower than those at an OLR of 3.38 kgvs/m³d. 395 However, it is important to underline that even for pre-treatments, the increase in OLR reduced the 396 NPV, but pre-treatments were able to slow down plant impairment. 397 These results are of fundamental importance since usually, capital and operational costs of sAD with 398 pre-treatment are critical and expensive due to the application of additional energy or chemicals 399 (Michalsk and Ledakowicz, 2014). 400 Usually, the main economic issue related to pre-treatments is the possible economic unprofitability 401 since the enhancement of methane yields could not offset the efforts of the item and energy required 402 by pre-treatment (Fu et al., 2018). Whereas, in the present study, the pre-treatments were 403 economically profitable up to OLR =4.50 kgvs/m³d (HRT = 12 d) for EN-OFMSW and OLR=5.40 404 kgvs/m³d (HRT = 10 d) for HC-OFMSW. In both HC- and EN-OFMSW configurations, the pre-405 treatment unit represented less than 15 % of capital costs and 10 % of the operational ones. As stated 406 by the technical performances, the pre-treatment allowed the process to be stable to produce enough 407 408 methane to overcome the energy required to manage the process. HC-OFMSW reached higher economic profitability than EN-OFMSW because the bottleneck of EN-409 OFMSW was the cost of the enzyme UPP2, the regulation of pH for UPP2 through the addition of 410 NaOH and the duration of pre-treatment (EN-OFMSW= 2 h vs HC-OFMSW= 10 min). Hence, HC-411 OFMSW could be economically profitable until OLR =5.40 kgvs/m³d, whereas EN-OFMSW could 412 be until OLR= $4.50 \text{ kgvs/m}^3 \text{d}$. 413

414	Usually, enzyme-assisted AD is not economically profitable due to the cost of the enzyme (Khanh
415	Nguyen et al., 2021). This cost could be overcome by the implementation of a cocktail of enzymes
416	like UPP2, which promoted a synergistic effect during the floc structure disintegration of OFMSW
417	(Salihu and Alam, 2016). However, also a cocktail of enzymes has drawbacks as the request of time
418	for the hydrolysis and correction of pH (Meegoda et al., 2018) because UPP2 required 2 h and pH =
419	7-7.5 obtained with NaOH addition.
420	The studies of (Jiang et al., 2020) reported that microaerobic pre-treatment is also an alternative pre-
421	treatment, both for solid substrate as cellulosic substrate and liquid substrate as food waste and brown
422	water, through which aerobic microorganisms could be enriched and produce more extracellular
423	enzymes like cellulase to boost the hydrolysis rate. Microaerobic pre-treatment could be more
424	economical than other biological pre-treatments since it does not require an addition of chemicals.
425	To conclude the cost-benefit analysis, HC-OFMSW reached the highest values of NPV and shortest
426	PBT for all the tested OLRs. The economic profitability of HC-OFMSW until OLR = $5.40 \text{ kgvs/m}^3\text{d}$
427	disagreed with the study of (Passos et al., 2017). (Passos et al., 2017) investigated the thermochemical
428	pre-treatment (i.e. HC-OFMSW) of dairy cow manure at OLR= 5.25 kgvs/m ³ d, resulting in more
429	expensive than AD performed without pre-treatment.
430	EN-OFMSW was less advantageous than HC-OFMSW, because the marginal biogas increase may
431	not be able to justify the expensive cost of enzymes according to (Montgomery and Bochmann, 2014).
432	However, both pre-treatments HC- and EN-OFMSW reached economic performance higher than the
433	ones of AD-OFMSW.



434

435

Fig. 2. Net present value (NPV) of the sAD configurations gathered per type of pre-treatments according to hydraulic retention time (HRT).

3.3 Environmental evaluation

437

Carbon footprint analysis (CF) was performed according to ISO 14067 on three sAD configurations 438 by varying five OLRs from 3.38 to 6.75 kgvs/m³d. The results of CF analysis (Fig. 3) were expressed 439 on the base of the functional unit (FU) equal to 1 MWh/d of produced primary energy. The method 440 IPCC 2021 GWP 100 V0.1 was adopted to calculate the climate change impact category. The results 441 achieved through the CF analysis are evaluated according to the effect of the OLR and pre-treatments. 442 Considering the variation in OLR, all three sAD configurations (AD-OFMSW, EN-OFMSW, and 443 HC-OFMSW) exhibited a decrease in GWP impacts within the OLR range of 3.38-3.87 kgvs/m³d, 444 indicating a reduction in emissions, and detail the avoidance of emissions. The most significant 445 reductions in GWP impacts were-at OLR of 3.87 kgvs/m³d (HRT = 14); in detail: -407.21-kg CO₂ 446 eq/FU for HC-OFMSW, -377.18 kg CO₂ eq/FU for EN-OFMSW and -336.44 kg CO₂ eq/FU for sAD-447 OFMSW. At OLR ≥ 4.50 kgvs/m³d for pre-treated sAD, and at OLR > 3.87 kgvs/m³d for AD-448 OFMSW, the GWP increased because the CH₄ production slightly decreased by increasing of the 449 OLR. Additionally, the energy requirements increased with the increase of OLR from 3.87 to 6.75 450 kgvs/m³d, primarily due to a higher daily substrate feed (+ 3.55% for AD-OFMSW, +3.2% for EN-451 OFMSW, and +2.5% for HC-OFMSW in Table S5). 452 Considering the effect of the pre-treatments, the lowest GWP impacts were achieved by HC-OFMSW, 453 followed by EN-OFMSW, and last AD-OFMSW. This rank agreed with the technical and cost-benefit 454 analyses (paragraphs 3.1 and 3.2). In detail, the GWP impacts were negative for AD-OFMSW from 455 $3.38 \text{ to } 3.87 \text{ kg}_{vs}/m^3 \text{d}$ (HRT= 16-14 d), for EN-OFMSW from $3.38 \text{ to } 4.50 \text{ kgvs/m}^3 \text{d}$ (HRT= 16-12 d) 456 and for HC-OFMSW from 3.38 to 5.40 kgvs/m³d (HRT= 16-10 d). This trend proved that the pre-457 treatments could buffer the inhibition occurring with the increase of OLR because they increased the 458 content of organic matter available to be converted into methane (Deepanraj et al., 2017). 459 Most of the available studies about environmental evaluation focusing on GWP referred to existing 460 industrial AD plants. The present study compared the results achieved with studies concerning the 461

climate change impact category of real AD-OFMSW or real AD of agro- waste in the European 462 scenarios, to be closer to the geographical context and feedstock treated in the present study. 463 The results of CF-analysis of sAD-OFMSW agreed with most of the studies performed in batch, feed-464 batch, and continuous AD performed with OFMSW (Kumar and Samadder, 2020), and in detail for 465 OLR = 3.87 kgvs/m³d, the GWP agreed with the study of (Mezzullo et al., 2013), which investigated 466 the same OLR, in which the avoided impacts were mainly due to the replacement of the energy and 467 fertilizer productions. 468 The results obtained in the present study by AD-OFMSW within the range of 3.38 and 4.50 kgvs/m³d 469 aligned with previous studies utilising 1 MWh/d as a functional unit. Examples include the study of 470 (Fusi et al., 2016) and (Bacenetti et al., 2016) reporting -375 and 408 kg CO₂eg/FU for sAD feed with 471 agro-waste and silage, respectively. 472 The GWP impacts reached by EN-OFMSW for HRT = 14-12 d agreed with the one obtained by 473 (Agostinho et al., 2015) under similar working conditions with cellulase enzyme applied on 474 lignocellulosic material before anaerobic digestion at the industrial scale. 475 The avoided emissions achieved by HC-OFMSW in the range 3.87-5.40 kgvs/m³d were higher than 476 those obtained by (Vosooghnia et al., 2021) through ultrasound pre-treated agro-waste before AD. 477 In detail, the avoided emissions for HC-OFMSW at 4.50 kgvs/m³d were around 6 % higher than the 478 479 ones reached by the plant in the UK, performing thermos-mechanical pre-treatment, (Evangelisti et al., 2014) and by the plant in France, performing physical pre-treatment, (Lamnatou et al., 2019) 480 under similar OLR. 481 These results proved that EN and HC performed in the present study were lower impactful than other 482 pre-treatments (already implemented at the industrial scale) since they can solubilise the complex 483 organic substrates improving their biodegradability even at high OLR. 484 The energy cost to carry out HC was offset by the methane surplus produced by HC-OFMSW at 485 OLRs from 3.38 to 5.40 kg_{vs}/m³d, since for these OLRs the HC reached positive energy balance due 486

to lower energy plant and operating costs, higher process yields and energy savings due to shorter
pre-treatment process times (10 min) (Calcio et al., 2018).
The EN-OFMSW avoided emissions for OLR from 3.38 to 4.50 5.40 kgvs/m³d and achieved positive
emissions with OLR from 5.40 to 6.75 kgvs/m ³ d. The GWP impacts related to EN-OFMSW were
higher than the those of HC-OFMSW, since EN lasted 2 h and required the addition of NaOH to keep
pH around 7-7.5 for the UPP2 activity. Furthermore, EN reached lower methane production and pH
stability compared to HC. In the scientific literature, there are no specific studies about the
environmental impacts of sAD performed with a cocktail of enzymes like UPP2. However, the GWP
contribution of enzyme in EN-OFMSW cannot be neglected since the enzyme production represents
10.5-11% of the total GWP impact in enzymatic pre-treatment (Olofsson et al., 2017). In the CF
analysis of the present study, the GWP associated with UPP2 was taken from Ecoinvent 3.5
considering a generic enzyme (secondary data) because no data were available for UPP 2 since it is a
patented commercial cocktail of cellulases, hemicellulases, pectinases, and proteases.
The sensitivity analysis confirmed the GWP results and the trends. HC-OFMSW and EN-OFMSW
played a positive effect on the increase in OLR because they buffered the inhibition and kept the
process stable, which means lower CO ₂ eq emissions (Wei et al., 2022). The results confirmed HC-
OFMSW as an effective pre-treatment (Saxena et al., 2019).

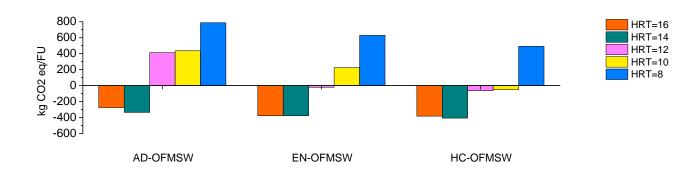


Fig. 3. CF results calculated with the metho IPCC 2021 GWP 100 V0.1

3.4 Principal Component Analysis

508

516

517

518

519

520

521

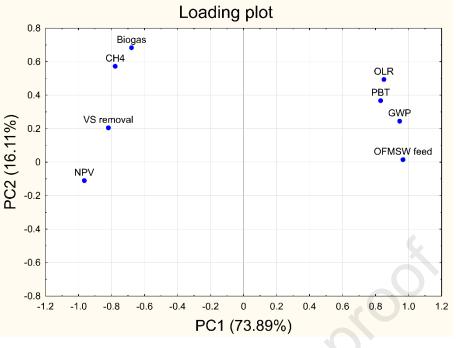
522

- Principal Component Analysis (PCA) was applied to the dataset consisting of the 30 samples (6 HRT values x 3 pre-treatments x 2 replicates) described by 8 variables (biogas and methane productions, VS removal, OFMSW feed, OLR, GWP, NPV, and PBT). Data were auto-scaled (mean centering followed by normalization to unit variance) hence all the variables accounted for the same amount of information. The first two PCs calculated explain about 90% of the overall information (73.89 % PC1 and 16.11 % PC2).
- Considering the loading plot (Fig. 4a), different behaviours can be detected:
 - GWP, OFMSW feed, PBT and OLR show positive weights on PC₁, while biogas, CH₄, VS removal and NPV are at negative weights on the same PC. These two groups of variables are negatively correlated with each other (when one group of variables increases the other one decreases and vice versa)
 - NPV was negatively correlated to OFMSW feed; indeed, when OFMSW feed increased
 NPV decreased, and vice versa.
 - Biogas and CH₄, were strongly correlated with each other, as is the case of PBT, OLR and GWP.
- The score plot proves that the samples were grouped predominantly according to the HRT (or to the corresponding OLR), however, within each HRT group (or corresponding OLR) the samples were clustered according to the pre-treatment. Samples with HRT=8 d (OLR = 6.75 kg _{vs}/m³ d) were at positive scores on PC₁, showing higher values of OFMSW feed, GWP, PBT and OLR. This is more evident for AD-OFMSW, while the trend decreases for EN-OFMSW and HC-OFMSW samples respectively.
- HRT= 8 d and HRT=14 d exhibited opposite trends, indeed HRT=8 d samples reached low values of
- VS removal, NPV, biogas, and CH₄, whereas the samples with HRT= 14 d reached high values of
- these variables and low values of OFMSW feed, GWP, OLR and PBT.

533	At HRT=14 -16 d, all the tested configurations reached high NPV, but the HC configuration depicted
534	higher NPV values, biogas productions and CH ₄ productions, compared to EN-OFMSW and above
535	all AD-OFMSW. This trend was due to the pre-treatment unit able to cover the costs for the additional
536	energy or chemicals and increase the process stability (Michalsk and Ledakowicz, 2014).
537	HRT samples were in groups with a progressive changing of their position in the score plot: from
538	right, counter clockwise, there are HRT= 8, 10, 12, 14, and 16 d. From the initial behaviour of HRT=8
539	d samples, at HRT=10 and 12 d, all the samples are at positive or slightly negative values on PC2, but
540	the behaviour seems more dependent on the pre-treatment applied, with AD samples are in fact at
541	more positive values on PC1 (higher OLR, PBT, GWP, OFMSW feed; lower biogas, NPV, CH4, and
542	VS removal).
542	v S Tellioval).
543	HRT=14 d samples were characterized by the lowest values of OLR, PBT, GWP, and the best balance
543	HRT=14 d samples were characterized by the lowest values of OLR, PBT, GWP, and the best balance
543 544	HRT=14 d samples were characterized by the lowest values of OLR, PBT, GWP, and the best balance between CH4 and biogas productions and VS removal, on one side, and NPV, on the other.
543544545	HRT=14 d samples were characterized by the lowest values of OLR, PBT, GWP, and the best balance between CH ₄ and biogas productions and VS removal, on one side, and NPV, on the other. HRT=16 d samples behaved like HRT=14 d samples but with lower NPV and lower biogas, CH ₄ , and
543544545546	HRT=14 d samples were characterized by the lowest values of OLR, PBT, GWP, and the best balance between CH4 and biogas productions and VS removal, on one side, and NPV, on the other. HRT=16 d samples behaved like HRT=14 d samples but with lower NPV and lower biogas, CH4, and VS removal.
543544545546547	HRT=14 d samples were characterized by the lowest values of OLR, PBT, GWP, and the best balance between CH4 and biogas productions and VS removal, on one side, and NPV, on the other. HRT=16 d samples behaved like HRT=14 d samples but with lower NPV and lower biogas, CH4, and VS removal. The pre-treatments reached the same trend for almost all HRT values, and HC-OFMSW samples

а

b



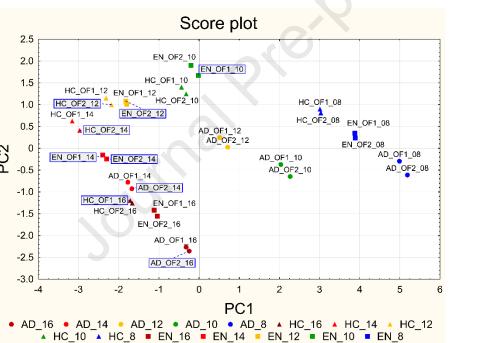


Fig. 4. PCA results are represented by loading plot (a) and score plot (b). where the samples are labelled according to both the treatment applied and the HRT: the treatment is indicated with a different shape of the label (AD as circles, HC as triangles, EN as squares), while the HRT is indicated by a different colour (HRT 16 = brown, 14 = red, 12 = yellow, 10 = green, 8 = blue).

3.5 Supervised Kohonen's Self-Organizing Maps

557

558

559

560

561

562

563

564

565

566

567

568

569

570

571

572

573

574

575

576

577

578

579

580

To highlight more clearly the effects of HRT and OLR, notwithstanding the pre-treatment, data were centred according to the pre-treatments. Supervised Kohonen self-organizing maps were applied as classification method for discriminating the samples according to the HRT value. Calculations were performed with a top map of 6x6 squared neurons and 300 epochs; cross-validation was applied with a Venetian blind procedure with 6 cancellation groups. Fig. 5a reports the top map, where neurons are coloured according to the class assignation; samples in the same neuron or adjacent neurons are more similar. Fig. 5b-i represents the trend of the weights of each original variable concerning the separation of the samples identified on the top map. All the samples were correctly classified the according to HRT, both in fitting and cross-validation, as can be noticed by the top map (Fig. 5a). Replications were in the same neuron or adjacent ones; furthermore, all the samples with HRT=14 d or HRT=16 d were in the same neuron, showing lower variabilities if compared to the other classes. These results confirmed the choice of the corrective factor for CH₄ values (paragraph 2.4). The plots of the weights of each original variable (Fig. 5 b-i) are represented on a colour scale from black (weights=1) to white (weights=0). For a given variable, when the colour of a neuron containing one or more samples is black, those samples showed a high value of the considered variable; whereas when the neuron is white, the samples showed a low value of that variable. Considering the plots of the weights, it was possible to verify the differences existing between the classes.

- OFMSW feed, GWP and OLR exhibited almost identical behaviours, with high values for HRT=8 d, smaller values for HRT=10 d, and the best values for HRT ≥14 d.
- Biogas and CH₄. were high for HRT=10 (especially biogas) and 12 d (especially CH₄), while HRT=14 d reached quite high values of biogas and very high for CH₄. VS removal was high for HRT=14 and 12 d. Biogas, CH₄, and VS removal were low for HRT=8 d.
- PBT was high for all the HRT values except for HRT=14 and 16 d, for which it was very low
- NPV showed very low values for HRT=8 d (for all the sAD configuration it was negative)
 and especially high for HRT ≥14 d.

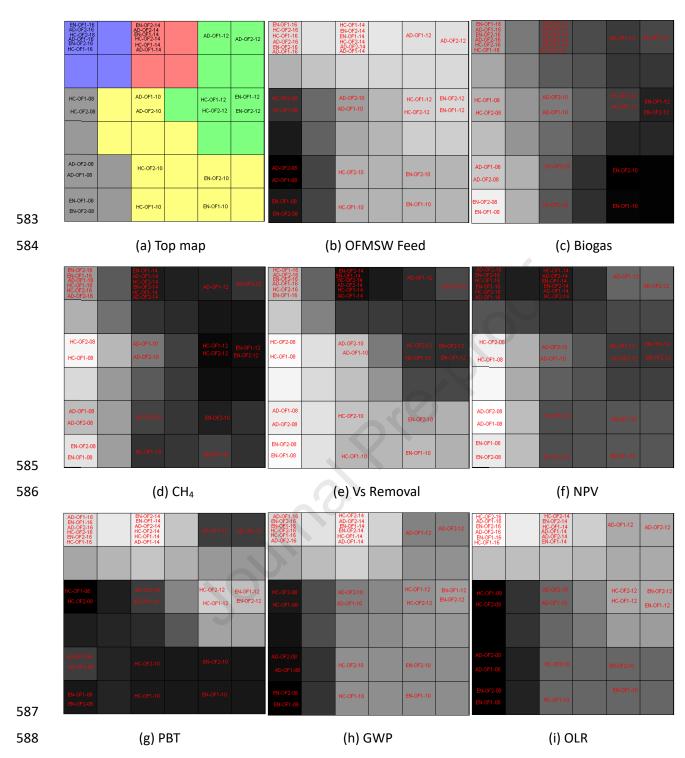


Fig. 5. Supervised Kohonen's self-organizing maps: top map (a); the plot of the weights of each original variable (b-i) – black corresponds to weights = 1, white corresponds to weights = 0.

4 Conclusions

This study investigated the technical feasibility, economic viability, and carbon footprint-analysis
(CF) of semi-continuous anaerobic digestion (sAD) of the organic fraction of municipal solid waste
(OFMSW). The study explored the effect of pre-treatment on sAD by increasing the organic loading
rates (OLR) from 3.38 to 6.75 kgvs/m ³ d. Two pre-treated sAD configurations were investigated:
hydrodynamic-cavitated (HC-OFMSW), enzymatically pre-treated (EN-OFMSW), and one non-pre-
treated (AD-OFMSW) as control. The results of the three-fold sustainability evaluations were
elaborated through Principal Component Analysis and Supervised Kohonen's Self-Organizing Maps.
Statistical analysis highlighted that the samples were grouped predominantly according to the OLR
(or to the corresponding HRT), however, within each HRT group the samples were clustered
according to the pre-treatment.
The main finding pointed out that pre-treated sADs were technically feasible, economically
profitable, and carbon negative up to 4.50 kgvs/m³d for EN-OFMSW and 5.40 kgvs/m³d for HC-
OFMSW. In contrast, the sAD-OFMSW exhibited viability only up to OLR of 3.87 kgvs/m ³ d.
These outcomes were important because pre-treatments, conventionally considered expensive due to
additional energy or chemicals, emerged as viable solutions to be implemented at the industrial scale.
Pre-treated-AD required smaller plants compared to AD-OFMSW by treating a higher amount of
OFMSW per unit of volume and time and by producing higher amount of renewable energy. Hence,
HC-OFMSW can contribute to the decarbonization of the energy-productive system and enhance
waste management. These actions aligned both with the principles of Bioeconomy and contributed to
the target outlined in Mission 2 of the National Recovery Plan.

613	Statements and Declarations
614	Authorship contribution statement
615	The contributions of the authors are detailed in the following. F. Demichelis contributed to the
616	conceptualization of the study, she carried out the experimental tests and wrote part of the paper. F.A.
617	Deorsola contributed to realizing the methodology and reviewed the paper, and T. Tommasi reviewed
618	the manuscript. E. Robotti coordinated and realized the modelling study and contributed to writing
619	and reviewing the manuscript. E. Marengo contributed to the modelling analysis and D. Fino
620	supervised the study.
621	Funding
622	The authors gratefully acknowledge financial support from Region Piemonte (Italy), POR FESR
623	2014/2020, Project BIOENPRO4TO.
624	Competing interests
625	The authors declare no competing interests.
626	Data availability
627	Data are completely reported in the manuscript and the supplementary material.
628	Appendix
629	The appendix provided information about environmental and economic analysis.
630	

- 631 Reference
- Agostinho, F., Bertaglia, A.B.B., Almeida, C.M.V.B., Giannetti, B.F., 2015. Influence of cellulase
- enzyme production on the energetic-environmental performance of lignocellulosic ethanol.
- Ecol. Modell. 315, 46–56. https://doi.org/10.1016/j.ecolmodel.2014.09.005
- 635 Alfonso-Cardero, A., Pagés-Díaz, J., Contino, F., Rajendran, K., Lorenzo-LLanes, J., 2021. Process
- simulation and techno-economic assessment of vinasse-to-biogas in Cuba: Deterministic and
- uncertainty analysis. Chem. Eng. Res. Des. 169, 33–45.
- https://doi.org/10.1016/j.cherd.2021.02.031
- 639 Aslanzadeh, S., Rajendran, K., Taherzadeh, M.J., 2014. International Biodeterioration &
- Biodegradation A comparative study between single- and two-stage anaerobic digestion
- processes: Effects of organic loading rate and hydraulic retention time. Int. Biodeterior.
- Biodegradation 95, 181–188. https://doi.org/10.1016/j.ibiod.2014.06.008
- Aui, A., Li, W., Wright, M.M., 2019. Techno-economic and life cycle analysis of a farm-scale
- anaerobic digestion plant in Iowa. Waste Manag. 89, 154–164.
- https://doi.org/10.1016/j.wasman.2019.04.013
- Babaee, A., Shayegan, J., 2016. Effect of Organic Loading Rates (OLR) on Production of Methane
- from Effect of organic loading rates (OLR) on production of methane from anaerobic
- digestion of vegetables waste. https://doi.org/10.3384/ecp11057411
- Bacenetti, J., Sala, C., Fusi, A., Fiala, M., 2016. Agricultural anaerobic digestion plants: What LCA
- studies pointed out and what can be done to make them more environmentally sustainable.
- Appl. Energy 179, 669–686. https://doi.org/10.1016/j.apenergy.2016.07.029
- Balasundaram, G., Vidyarthi, P.K., Gahlot, P., Arora, P., Kumar, V., Kumar, M., Kazmi, A.A.,
- Tyagi, V.K., 2022. Energy feasibility and life cycle assessment of sludge pretreatment methods
- for advanced anaerobic digestion. Bioresour. Technol. 357, 127345.
- https://doi.org/10.1016/j.biortech.2022.127345
- Ballabio, D., Consonni, V., Todeschini, R., 2009. The Kohonen and CP-ANN toolbox: A collection
- of MATLAB modules for Self Organizing Maps and Counterpropagation Artificial Neural
- Networks. Chemom. Intell. Lab. Syst. 98, 115–122.
- https://doi.org/10.1016/j.chemolab.2009.05.007
- Barber, W.P.F., 2016. Thermal hydrolysis for sewage treatment: A critical review. Water Res. 104,
- 53–71. https://doi.org/10.1016/j.watres.2016.07.069

- Brandi, J., Robotti, E., Manfredi, M., Barberis, E., Marengo, E., Novelli, E., Cecconi, D., 2021.
- Kohonen Artificial Neural Network and Multivariate Analysis in the Identification of
- Proteome Changes during Early and Long Aging of Bovine Longissimus dorsi Muscle Using
- SWATH Mass Spectrometry. J. Agric. Food Chem. 69, 11512–11522.
- https://doi.org/10.1021/acs.jafc.1c03578
- Bruni, E., Jensen, A.P., Angelidaki, I., 2010. Steam treatment of digested biofibers for increasing
- biogas production. Bioresour. Technol. 101, 7668–7671.
- https://doi.org/10.1016/j.biortech.2010.04.064
- Bruno, M., Marchi, M., Ermini, N., Niccolucci, V., Pulselli, F.M., 2023. Life Cycle Assessment and
- 671 Cost–Benefit Analysis as Combined Economic–Environmental Assessment Tools: Application
- to an Anaerobic Digestion Plant. Energies 16. https://doi.org/10.3390/en16093686
- 673 Calcio, E., Tabasso, S., Grillo, G., Cravotto, G., Dreyer, T., Schories, G., Altenberg, S., Jashina, L.,
- Telysheva, G., 2018. Comptes Rendus Chimie Wheat straw lignin extraction with bio-based
- solvents using enabling technologies. Comptes rendus Chim. 21, 563–571.
- https://doi.org/10.1016/j.crci.2018.01.010
- 677 Carlqvist, K., Wallberg, O., Lidén, G., Börjesson, P., 2022. Life cycle assessment for identification
- of critical aspects in emerging technologies for the extraction of phenolic compounds from
- spruce bark. J. Clean. Prod. 333. https://doi.org/10.1016/j.jclepro.2021.130093
- 680 Carlsson, M., Naroznova, I., Møller, J., Scheutz, C., Lagerkvist, A., 2015. Resources, Conservation
- and Recycling Importance of food waste pre-treatment efficiency for global warming potential
- in life cycle assessment of anaerobic digestion systems. "Resources, Conserv. Recycl. 102, 58–
- 683 66. https://doi.org/10.1016/j.resconrec.2015.06.012
- 684 Chen, Y., Cheng, J.J., Creamer, K.S., 2008. Inhibition of anaerobic digestion process: A review 99,
- 685 4044–4064. https://doi.org/10.1016/j.biortech.2007.01.057
- 686 Choudhary, A., Kumar, A., Kumar, S., 2020. Techno-economic analysis, kinetics, global warming
- potential comparison and optimization of a pilot-scale unheated semi-continuous anaerobic
- reactor in a hilly area: For north Indian hilly states. Renew. Energy 155, 1181–1190.
- https://doi.org/10.1016/j.renene.2020.04.034
- 690 Clift, R., Doig, A., Finnveden, G., 2000. The application of Life Cycle Assessment to Integrated
- Solid Waste Management. Part 1 Methodology. Process Saf. Environ. Prot. 78, 279–287.
- 692 https://doi.org/10.1205/095758200530790

- Dareioti, M.A., Kornaros, M., 2014. Effect of hydraulic retention time (HRT) on the anaerobic co-
- digestion of agro-industrial wastes in a two-stage CSTR system. Bioresour. Technol. 167,
- 695 407–415. https://doi.org/10.1016/j.biortech.2014.06.045
- 696 Deepanraj, B., Sivasubramanian, V., Jayaraj, S., 2017. ScienceDirect Effect of substrate
- 697 pretreatment on biogas production through anaerobic digestion of food waste. Int. J. Hydrogen
- 698 Energy 42, 26522–26528. https://doi.org/10.1016/j.ijhydene.2017.06.178
- Demichelis, F., Robotti, E., Deorsola, F. A., Marengo E., T., T., Fino, D., n.d. Enzymatic hydrolysis
- of organic fraction municipal solid waste to enhance and decarbonize the anaerobic digestion.
- 701 Under Revis.
- Demichelis, F., Deorsola, F.A., Robotti, E., Cravotto, G., Marengo, E., Tommasi, T., Grillo, G.,
- Fino, D., 2023. Experimental and modelling optimisation of sustainable techniques for the pre-
- treatment of the organic fraction municipal solid waste to improve anaerobic digestion. J.
- 705 Clean. Prod. 399, 136594. https://doi.org/10.1016/j.jclepro.2023.136594
- Demichelis, F, Tommasi, T., Deorsola, F.A., Marchisio, D., Fino, D., 2022. Effect of inoculum
- origin and substrate-inoculum ratio to enhance the anaerobic digestion of organic fraction
- municipal solid waste (OFMSW). J. Clean. Prod. 351, 131539.
- 709 https://doi.org/10.1016/j.jclepro.2022.131539
- 710 Demichelis, F., Tommasi, T., Deorsola, F.A., Marchisio, D., Mancini, G., Fino, D., 2022. Life cycle
- assessment and life cycle costing of advanced anaerobic digestion of organic fraction
- municipal solid waste. Chemosphere 289, 133058.
- 713 https://doi.org/10.1016/j.chemosphere.2021.133058
- Demirer, G.N., Chen, S., 2005. Two-phase anaerobic digestion of unscreened dairy manure. Process
- 715 Biochem. 40, 3542–3549. https://doi.org/10.1016/j.procbio.2005.03.062
- 716 Ennio, S., 2018. REALIZZAZIONE DI NUOVO IMPIANTO DI COMPOSTAGGIO IN LOCALE
- 717 CHIUSO.
- Evangelisti, S., Lettieri, P., Borello, D., Clift, R., 2014. Life cycle assessment of energy from waste
- via anaerobic digestion: A UK case study. Waste Manag. 34, 226–237.
- 720 https://doi.org/10.1016/j.wasman.2013.09.013
- Fu, S.F., Chen, K.Q., Zhu, R., Sun, W.X., Zou, H., Guo, R.B., 2018. Improved anaerobic digestion
- performance of Miscanthus floridulus by different pretreatment methods and preliminary
- economic analysis. Energy Convers. Manag. 159, 121–128.

- 724 https://doi.org/10.1016/j.enconman.2018.01.014
- Fusi, A., Bacenetti, J., Fiala, M., Azapagic, A., 2016. Life cycle environmental impacts of
- electricity from biogas produced by anaerobic digestion. Front. Bioeng. Biotechnol. 4.
- 727 https://doi.org/10.3389/fbioe.2016.00026
- Garuti, M., Langone, M., Fabbri, C., Piccinini, S., 2018. Monitoring of full-scale hydrodynamic
- cavitation pretreatment in agricultural biogas plant. Bioresour. Technol. 247, 599–609.
- 730 https://doi.org/10.1016/j.biortech.2017.09.100
- 731 GmbH, B.A., 2022. BIOPRACT [WWW Document]. URL https://www.biopract-
- abt.de/en/products/ultrapract/ultrapract-p2 (accessed 9.20.23).
- Green, D.W., Southhard, M.Z., 2019. Perry's Chemical Engineers, 9 th Editi. ed.
- Gu, Y., Chen, X., Liu, Z., Zhou, X., Zhang, Y., 2020. Bioresource Technology Effect of inoculum
- sources on the anaerobic digestion of rice straw 158, 149–155.
- 736 https://doi.org/10.1016/j.biortech.2014.02.011
- Huiru, Z., Yunjun, Y., Liberti, F., Bartocci, P., Fantozzi, F., 2019. Technical and economic
- feasibility analysis of an anaerobic digestion plant fed with canteen food waste. Energy
- 739 Convers. Manag. 180, 938–948. https://doi.org/10.1016/j.enconman.2018.11.045
- Jiang, J., Li, L., Li, Y., He, Y., Wang, C., Sun, Y., 2020. Bioaugmentation to enhance anaerobic
- 741 digestion of food waste: Dosage, frequency and economic analysis. Bioresour. Technol. 307.
- 742 https://doi.org/10.1016/j.biortech.2020.123256
- Khanh Nguyen, V., Kumar Chaudhary, D., Hari Dahal, R., Hoang Trinh, N., Kim, J., Chang, S.W.,
- Hong, Y., Duc La, D., Nguyen, X.C., Hao Ngo, H., Chung, W.J., Nguyen, D.D., 2021. Review
- on pretreatment techniques to improve anaerobic digestion of sewage sludge. Fuel 285,
- 746 119105. https://doi.org/10.1016/j.fuel.2020.119105
- Kowalczyk, A., Schwede, S., Gerber, M., Span, R., 2011. Scale Up of Laboratory Scale to
- Industrial Scale Biogas Plants. Proc. World Renew. Energy Congr. Sweden, 8–13 May,
- 749 2011, Linköping, Sweden 57, 48–55. https://doi.org/10.3384/ecp1105748
- Kumar, A., Samadder, S.R., 2020. Performance evaluation of anaerobic digestion technology for
- energy recovery from organic fraction of municipal solid waste: A review. Energy 197,
- 752 117253. https://doi.org/10.1016/j.energy.2020.117253
- Lamnatou, C., Nicolaï, R., Chemisana, D., Cristofari, C., Cancellieri, D., 2019. Science of the Total

- Environment Biogas production by means of an anaerobic-digestion plant in France : LCA of
- greenhouse-gas emissions and other environmental indicators. Sci. Total Environ. 670, 1226–
- 756 1239. https://doi.org/10.1016/j.scitotenv.2019.03.211
- Leite, W., Magnus, B.S., Guimarães, L.B., Gottardo, M., Belli Filho, P., 2017. Feasibility of
- 758 thermophilic anaerobic processes for treating waste activated sludge under low HRT and
- intermittent mixing. J. Environ. Manage. 201, 335–344.
- 760 https://doi.org/10.1016/j.jenvman.2017.06.069
- 761 Li, Yangyang, Wang, Y., Yu, Z., Lu, J., Li, D., Wang, G., Li, Yu, Wu, Y., Li, S., Xu, F., Li, G.,
- Gong, X., 2018. Effect of inoculum and substrate / inoculum ratio on the performance and
- methanogenic archaeal community structure in solid state anaerobic co-digestion of tomato
- residues with dairy manure and corn stover 81, 117–127.
- 765 https://doi.org/10.1016/j.wasman.2018.09.042
- Liu, X., Khalid, H., Amin, F.R., Ma, X., Li, X., Chen, C., Liu, G., 2018. Effects of hydraulic
- retention time on anaerobic digestion performance of food waste to produce methane as a
- 768 biofuel. Environ. Technol. Innov. 11, 348–357. https://doi.org/10.1016/j.eti.2018.06.004
- Massart, D., 1988. Chemometrics: A Textbook. Elsevier Amsterdam.
- Meegoda, J.N., Li, B., Patel, K., Wang, L.B., 2018. A Review of the Processes, Parameters, and
- Optimization of Anaerobic Digestion. https://doi.org/10.3390/ijerph15102224
- Mei, R., Narihiro, T., Nobu, M.K., Kuroda, K., Liu, W.T., 2016. Evaluating digestion efficiency in
- full-scale anaerobic digesters by identifying active microbial populations through the lens of
- 774 microbial activity. Sci. Rep. 6, 1–10. https://doi.org/10.1038/srep34090
- Melssen, W., Wehrens, R., Buydens, L., 2006. Supervised Kohonen networks for classification
- problems. Chemom. Intell. Lab. Syst. 83, 99–113.
- 777 https://doi.org/10.1016/j.chemolab.2006.02.003
- Mezzullo, W.G., McManus, M.C., Hammond, G.P., 2013. Life cycle assessment of a small-scale
- anaerobic digestion plant from cattle waste. Appl. Energy 102, 657–664.
- 780 https://doi.org/10.1016/j.apenergy.2012.08.008
- 781 Michalsk, K., Ledakowicz, S., 2014. Alkaline hydrogen peroxide pretreatment of energy crops for
- 782 biogas production. Chem. Pap. 68, 913–922. https://doi.org/10.2478/s11696-013-0531-5
- 783 Mohammad Rahmani, A., Gahlot, P., Moustakas, K., Kazmi, A.A., Shekhar Prasad Ojha, C., Tyagi,

- V.K., 2022. Pretreatment methods to enhance solubilization and anaerobic biodegradability of
- lignocellulosic biomass (wheat straw): Progress and challenges. Fuel 319, 123726.
- 786 https://doi.org/10.1016/j.fuel.2022.123726
- Montgomery, L.F.R., Bochmann, G., 2014. Pretreatment of feedstock for enhanced biogas
- production Pretreatment of feedstock for enhanced biogas production (electronic version) 1–
- 789 24.
- Morales-Polo, C., del Mar Cledera-Castro, M., Yolanda Moratilla Soria, B., 2018. Reviewing the
- anaerobic digestion of food waste: From waste generation and anaerobic process to its
- 792 perspectives. Appl. Sci. 8. https://doi.org/10.3390/app8101804
- 793 Nabi, M., Zhang, G., Zhang, P., Tao, X., Wang, S., Ye, J., Zhang, Q., Zubair, M., Bao, S., Wu, Y.,
- 794 2019. Contribution of solid and liquid fractions of sewage sludge pretreated by high pressure
- homogenization to biogas production. Bioresour. Technol. 286, 121378.
- 796 https://doi.org/10.1016/j.biortech.2019.121378
- Nagi, A., Kuhn, J.N., Joseph, B., 2019. Techno-economic analysis of producing liquid fuels from
- biomass via anaerobic digestion and thermochemical conversion. Biomass and Bioenergy 130,
- 799 105395. https://doi.org/10.1016/j.biombioe.2019.105395
- Nges, I.A., Liu, J., 2010. Effects of solid retention time on anaerobic digestion of dewatered-sewage
- sludge in mesophilic and thermophilic conditions. Renew. Energy 35, 2200–2206.
- https://doi.org/10.1016/j.renene.2010.02.022
- 803 Olofsson, J., Barta, Z., Börjesson, P., Wallberg, O., 2017. Integrating enzyme fermentation in
- lignocellulosic ethanol production: Life-cycle assessment and techno-economic analysis.
- Biotechnol. Biofuels 10, 1–14. https://doi.org/10.1186/s13068-017-0733-0
- Passos, F., Ortega, V., Donoso-Bravo, A., 2017. Thermochemical pretreatment and anaerobic
- digestion of dairy cow manure: Experimental and economic evaluation. Bioresour. Technol.
- 808 227, 239–246. https://doi.org/10.1016/j.biortech.2016.12.034
- Piccinno, F., Hischier, R., Seeger, S., Som, C., 2016. From laboratory to industrial scale: a scale-up
- framework for chemical processes in life cycle assessment studies. J. Clean. Prod. 135, 1085–
- 811 1097. https://doi.org/10.1016/j.jclepro.2016.06.164
- Pleissner, D., Neu, A., Mehlmann, K., Schneider, R., Puerta-quintero, G.I., Venus, J., 2016.
- Bioresource Technology Fermentative lactic acid production from coffee pulp hydrolysate
- using Bacillus coagulans at laboratory and pilot scales. Bioresour. Technol. 218, 167–173.

- Power and renewables, 2022. European biogas plant market [WWW Document]. Eur. biogas plant Mark.
- Rajendran, K., Murthy, G.S., 2019. Biocatalysis and Agricultural Biotechnology Techno-economic and life cycle assessments of anaerobic digestion A review. Biocatal. Agric. Biotechnol. 20,
- 820 101207. https://doi.org/10.1016/j.bcab.2019.101207

https://doi.org/10.1016/j.biortech.2016.06.078

- Rillo, E., Gandiglio, M., Lanzini, A., Bobba, S., Santarelli, M., Blengini, G., 2020. Life Cycle
- Assessment (LCA) of biogas-fed Solid Oxide Fuel Cell (SOFC) plant. Energy 126, 585–
- 823 602. https://doi.org/10.1016/j.energy.2017.03.041
- Rog, W., Scandolara, B., Moraes, B. De, Takayuki, M., Florencio, L., Helena, R., Belli, P., 2023.
- Mesophilic anaerobic digestion of waste activated sludge in an intermittent mixing reactor:
- Effect of hydraulic retention time and organic loading rate 338.
- 827 https://doi.org/10.1016/j.jenvman.2023.117839
- 828 Salihu, A., Alam, M.Z., 2016. Pretreatment Methods of Organic Wastes for Biogas Production. J.
- 829 Appl. Sci. 16, 124–137. https://doi.org/10.3923/jas.2016.124.137
- 830 Saxena, S., Saharan, V.K., George, S., 2019. Modeling & simulation studies on batch anaerobic
- digestion of hydrodynamically cavitated tannery waste effluent for higher biogas yield.
- Ultrason. Sonochem. 58, 104692. https://doi.org/10.1016/j.ultsonch.2019.104692
- 833 Srisowmeya, G., Chakravarthy, M., Devi, G.N., 2020. Critical considerations in two-stage
- anaerobic digestion of food waste A review. Renew. Sustain. Energy Rev. 119, 109587.
- https://doi.org/10.1016/j.rser.2019.109587
- Thushari, I., Vicheanteab, J., Janjaroen, D., 2020. Material flow analysis and life cycle assessment
- of solid waste management in urban green areas, Thailand. Sustain. Environ. Res. 30.
- https://doi.org/10.1186/s42834-020-00057-5
- Turton, R., Shaeiwitz, J.A., Bhattacharyya, D., 2018. Analysis, Synthesis, and Design of chemical
- process, 5th editio. ed.
- Ugwu, S.N., Harding, K., Enweremadu, C.C., 2022. Comparative life cycle assessment of enhanced
- anaerobic digestion of agro-industrial waste for biogas production. J. Clean. Prod. 345,
- 843 131178. https://doi.org/10.1016/j.jclepro.2022.131178
- Villa, R., Ferguson, R.M.W., 2016. Organic loading rate: A promising microbial management tool

845	in anaerobic digestion d e 100, 348-356. https://doi.org/10.1016/j.watres.2016.05.009
846	Vosooghnia, A., Polettini, A., Rossi, A., Vázquez-Rowe, I., Francini, G., 2021. Carbon footprint of
847	anaerobic digestion combined with ultrasonic post-treatment of agro-industrial organic
848	residues. J. Environ. Manage. 278. https://doi.org/10.1016/j.jenvman.2020.111459
849	Wei, Y., Gao, Y., Yuan, H., Chang, Y., Li, X., 2022. Effects of organic loading rate and
850	pretreatments on digestion performance of corn stover and chicken manure in completely
851	stirred tank reactor (CSTR). Sci. Total Environ. 815, 152499.
852	https://doi.org/10.1016/j.scitotenv.2021.152499
853	Wijekoon, K.C., Visvanathan, C., Abeynayaka, A., 2011. Effect of organic loading rate on VFA
854	production, organic matter removal and microbial activity of a two-stage thermophilic
855	anaerobic membrane bioreactor. Bioresour. Technol. 102, 5353-5360.
856	https://doi.org/10.1016/j.biortech.2010.12.081
857	Zeng, S., Yuan, X., Shi, X., Qiu, Y., 2010. Effect of inoculum / substrate ratio on methane yield and
858	orthophosphate release from anaerobic digestion of Microcystis spp . J. Hazard. Mater. 178,
859	89–93. https://doi.org/10.1016/j.jhazmat.2010.01.047
860	Zhou, H., Jiang, J., Zhao, Q., Li, L., Wang, K., Wei, L., 2022. Effects of organic loading rates on
861	high-solids anaerobic digestion of food waste in horizontal flow reactor: Methane production,
862	stability and mechanism. Chemosphere 293, 133650.
863	https://doi.org/10.1016/j.chemosphere.2022.133650
864	

Highlights

- Semi-continuous anaerobic digestion (sAD) evaluated the organic loading rates (OLR).
- Organic fraction of municipal solid waste was physical and biological pre-treated.
- Techno-economic-environmental study on sAD was elaborated with statistical analyses.
- sAD were grouped by OLR and within each group, sAD were clustered by pre-treatment.
- Physical pretreatment was economic and environmentally viable up to 5.40 kgvs/m³d.
- Semi-continuous anaerobic digestion (sAD) investigated five organic loading rates (OLR).
- Organic fraction of municipal solid waste was physical and biological pre-treated.
- Principal component analysis and Kohonen supervised map were calculated.
- The best performances were reached at OLR =3.87 kg/m³d

D	اءد	ara	tion	οf	inter	actc
u	-(.)	ala	11()[1		mer	->1>

oxtimes The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.
\Box The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: