POLITECNICO DI TORINO
Repository ISTITUZIONALE

Experimental and modelling optimisation of sustainable techniques for the pre-treatment of the organic fraction municipal solid waste to improve anaerobic digestion

Original
Experimental and modelling optimisation of sustainable techniques for the pre-treatment of the organic fraction municipal solid waste to improve anaerobic digestion / Demichelis, F., Deorsola, F.A., Robotti, E., Cravotto, G., Marengo, E., Tommasi, T., Grillo, G., Fino, D.. - In: JOURNAL OF CLEANER PRODUCTION. - ISSN 0959-6526. - 399:(2023). [10.1016/j.jclepro.2023.136594]

Availability:
This version is available at: 11583/2981606 since: 2023-09-05T08:13:03Z

Publisher:
ELSEVIER SCIENCE

Published
DOI:10.1016/j.jclepro.2023.136594

Terms of use:

This article is made available under terms and conditions as specified in the corresponding bibliographic description in the repository

Publisher copyright

(Article begins on next page)

16 September 2026

Experimental and modelling optimisation of sustainable techniques for the pre-treatment of the organic fraction municipal solid waste to improve anaerobic digestion

Francesca Demichelis[1*], Fabio Alessandro Deorsola[1], Elisa Robotti[2]*, Giancarlo Cravotto[3], Emilio Marengo[2], Tonia Tommasi[1], Giorgio Grillo[3], Debora Fino[1].

[1]Department of Applied Science and Technology (DISAT), Politecnico di Torino, Corso Duca degli Abruzzi 24, 10129 Torino, Italy

[2]Department of Sciences and Technological Innovation, University of Piemonte Orientale, Viale Michel 11, 15121 Alessandria, Italy

[3]Department of Drug Science and Technology, University of Turin, via Pietro Giuria 9, 10125 Turin, Italy.

Abstract

The study investigated and compared the anaerobic digestion (AD) of real organic fraction municipal solid waste (OFMSW) prior pre-treated with four types of pre-treatments: mechanical, thermal, hydrodynamic-cavitation (HC), and ultrasound (US). The tested pre-treatments and AD configurations were selected through Design of Experiments and then regression models were built to find the most promising configurations in terms of biogas production and energetic sustainability of the whole process. The novelty of the research is the simultaneously study of the working conditions of the pre-treatments; and AD parameters like the two origins of the inoculum, its incubation time, and the substrate: inoculum ratio (SI).

The results demonstrated that the best configurations of pre-treatments and AD were the ones performed with thermal pre-treatment at 120 °C for 45 min (with inoculum incubation of 10 d at substrate: inoculum (SI) ratio of 2:1) and HC at 55 °C (with inoculum incubation of 10 d at SI of 3:1). The thermal, and to some extent the mechanical pre-treatment, evidenced as significant the interaction between the pre-treatment time and the inoculum incubation time. AD of US-OFMSW achieved the lowest performances since inhibition occurred, probably due to the lignocellulosic inhibitors release after ultrasound pre-treatment.

Keywords: anaerobic digestion, pre-treatments, kinetic, energy sustainability, design of experiments, surface response

Abbreviation

AD: anaerobic digestion

CAS: mesophilic digestate of cow-agriculture sludge

COD: Chemical Oxygen Demand

DoE: design of experiment

DR: disintegration rate

HC: hydrodynamic-cavitation

OFMSW: organic fraction municipal solid waste

TOC: Total Organic Carbon

TS: total solids

VS: volatile solids

WAS: mesophilic digestate of wastewater activated sludge.

Highlights

- Evaluation of technical and energetic sustainability of AD of pre-treated OFMSW.
- DoE provided the pre-treatment and AD configurations to test.
- Regression model detected the feasible and energetic sustainable AD processes.
- The interaction of pre-treatment and AD enhanced the methane production.
- Thermal and HC pre-treatments reached the highest energetic performances.

Introduction

Anaerobic digestion (AD) is a mature technology adopted to stabilise the organic matter and convert it into biogas. Nowadays, the global warming change challenge is promoting the application of renewable alternative sources of energy, and AD can enhance biogas production, which is a renewable energy employed to produce heat-electricity and transport fuel. The above-mentioned applications of biogas are not still widely implemented due to the higher costs of biogas production rather than other renewable energy sources such as wind or photovoltaic. Nevertheless, biogas is an energy which can be stored and used directly without conversions and is able to face the problem of peak requirement and power failure (Scherzinger and Kaltschmitt, 2021). Due to these properties, AD is a mature technology to face climate change and resource depletion and produce clean energy (Hai et al., 2022) according to the Sustainable Development Goals.

The optimisation and improvement of AD consists in the study of the reactor design, the process conditions (as temperature, pH, mixing, etc.), the feedstocks employed and its pre-treatment (Kainthola et al., 2019). Based on the current scientific literature, pre-treatments may significantly improve AD, considering both the efficiency of the pre-treatment and its effect on the overall AD. The efficiency of the pre-treatment depends on the technique as well as on the type and composition of substrate employed. The performance of AD depends on several factors: the kinetics and the hydrolysis, the rate-limiting step, which depends on process conditions, the substrate composition, and the biodegradability (Carmona-Cabello et al., 2018). In the literature, referring to AD, the term biodegradability expresses the amount of material biologically convertible in methane. Complex substrates can range from no biodegradable, as lignin, to complete biodegradable, as starch and sugar (Cheah et al., 2019). Nevertheless, biodegradable compounds could have low degree of bioavailability since part of the bio-matter can be incorporated into complex and barely biodegradable lignocellulosic structures. In addition, substrates constituted by large particles will be slowly degraded due to the actual limited surface area (Harun et al., 2019).

Currently, the pre-treatments are investigated for substrates as wastewaters from treatment plants, crops-harvesting residues, wastes from the food industry, animal manure and organic fraction from households. The Organic Fraction Municipal Solid Waste (OFMSW), considered in the present study is a heterogeneous substrate, coming from the collection of municipal wastes including food industry and households, with an average composition including easily biodegradable compounds (as 17.5 % lipids, 17.7 % proteins, 17.1 % starch, 10.5 % free sugars), and harder biodegradable components (as 18.6 % cellulose, 9.7 % lignin and 8.6 % hemicellulose) (Pleissner and Peinemann, 2020).

In the last 40 years the most studied pre-treatments of organic wastes from the food industry and households reported in the scientific literature available in Science Direct, concerning the laboratory scale, are mechanical (35 %), thermal (26 %), ultrasonic (18 %), chemical (12 %) and microwaves (9 %).

Mechanical pre-treatments have been evaluated because the particle size significantly affects the kinetics and stability of AD, determining the success or failure of the process.

Thermal pre-treatments have been performed at mild temperature because most part of OFMSW consists of starch and hemicellulose, which can be hydrolysed at relative low temperatures (from 90 to 180 °C) with longer residence time (Li et al., 2017). Hydrothermal, steam explosion and vapour-thermal are the most adopted thermal pre-treatments; among them, the vapour-thermal pre-treatment requires a lower energy and a lower heating time compared to the hydrothermal and it could be done with jacket reactors for a wide range of substrates, without water addition (Scherzinger and Kaltschmitt, 2021).

Hydrodynamic cavitation (HC) prior to AD has provided a faster disintegration and the solubilization of larger organic molecules, so that it could be easily digested by the microbial inoculum, promoting a lower incubation time, higher degradation rates, and higher COD reduction with higher biogas generation (Saxena et al., 2019).

According to (Demichelis et al., 2022), for AD in batch feeding mode, inoculum plays a key role and the most important parameters for the inoculum are the incubation time (Zhang et al., 2019a), and source, and the substrate: inoculum (SI) ratio (Zhang et al., 2019b).

For the best of author's knowledge, the available scientific studies about pre-treatments and AD focused on them separately as in the studies from (Cesaro et al., 2014) and (Karthikeyan et al., 2018), who reviewed the available pre-treatment methods for organic wastes before the AD process. Other studies focused on: i) the evaluation of the pre-treatment performances through the disintegration rate, as in (Demichelis et al., 2018); ii) on the effects of pre-treatments (use of acid or alkaline reagents, and effect of temperature) on the AD of waste activated sludge from the perspectives of organic matter composition, thermodynamics, and multi-omics, as in (Chen et al., 2022); iii) by considering the improvement of AD after the pre-treatments, by an approach based on Anaerobic Digestion Model 1 (ADM1), as in (Huang et al., 2021) and (Wang et al., 2020). All these studies revealed that pre-treatment is a fundamental step with recalcitrant feedstocks to improve the methane yield; but, they did not investigate the interaction between pre-treatments and AD, which can be accomplished only if the conditions of both phases are changed contemporarily. The present study evaluated the AD of real OFMSW after pre-treatment with four types of pre-treatments: mechanical, thermal, HC, and US. The study concerned experimental tests and modelling. The experimentally tested pre-treatments and AD configurations were selected through DoE, then the enhancement of AD process (pre-treatment and AD) was evaluated in terms of biogas and methane productions, VS removal, kinetic study, and ESI.

The novelty of the study was the optimization of AD considering the combined effect and the interaction between pre-treatment and AD: this was possible by varying simultaneously the conditions of the two phases according to DoE, instead of simply adopting these pre-treatments before AD. The results of the experimentation, were used to build ~~a~~ regression model correlating the factors involved in the study, their interactions, and their quadratic effects, to the production of biogas and ESI, to obtain predictive models for the identification of the best running conditions for pre-treatment

and AD. In the results section, the best running condition will be detected for each type of pre-treatment by separately considering the biogas, the kinetic and the ESI results. In the conclusion section the best overall running conditions will be identified, among all the tested pre-treatments, by combining the results obtained for biogas production, kinetics, and ESI.

These results are of fundamental importance since there is a lack of information to understand the simultaneous effect of different pre-treatment techniques on AD performances, that should be urgently studied to improve the whole AD process according to (Atelge et al., 2020) and (Abraham et al., 2020).

Material and methods

2.1 Substrate and inoculum characterisation

AD of OFMSW, provided by San Carlo S.p.A (Fossano, Italy), was performed with two inocula: the mesophilic digestate of wastewater activated sludge (WAS), according to (Kumar Biswal et al., 2020), provided by SMAT (Torino, Italy), and the mesophilic digestate of cow-agriculture sludge (CAS), based on (Gu et al., 2020), supplied by "Cascina La Speranza" (Fossano, Cuneo, Italy).

The OFMSW, WAS and CAS were characterised in Table 1. VS/TS and TOC contents of OFMSW were higher than 90 % and 8,000 mg/kg, respectively; and this abundance of organic matter proved the suitability of OFMSW to be employed as feedstock for AD (L.Zhang et al., 2019). OFMSW had acid pH (5.6 ± 0.2), but the addition of WAS and CAS inocula increased the buffer capacity, since their pH were7.1 ± 0.1 (for WAS) and 7.7 ± 0.1 (for CAS). The physical-chemical properties of WAS agreed with (Suksong et al., 2019) and the ones of CAS agreed with mixtures of inocula from dairy manure and agricultural residues as reported in (Chen et al., 2008).

The C:N ratio of CAS was more suitable for AD than that of WAS, due to the higher carbon to nitrogen balance. CAS (a mix of cow manure and agricultural residues) could improve AD for its C:N ratio because the inhibition effect of nitrogen and ammonia from manure was limited by the carbon deriving from agricultural residues.

Table 1: Physical and chemical properties of OFMSW and inocula.

	TS (%)	VS (%)	pH (-)	C (%)	H (%)	N (%)	S (%)	C/N (-)	TOC (g/kg)
OFMSW (mean)	19.32	96.76	5.31	48.42	6.76	2.97	0.20	16.3	24,995.82
OFMSW (dev.st)	0.61	0.53	0.22	0.51	0.70	0.32	0.12	1.4	114.92
WAS (mean)	5.09	70.7	7.12	35.42	3.04	4.51	0.01	7.92	9.52
WAS (dev.st)	0.11	1.0	0.11	0.51	0.02	0.11	0.01	0.14	0.12
CAS (mean)	5.82	70.3	7.74	40.62	3.09	7.92	0.03	5.12	12.04
CAS (dev.st)	0.12	1.0	0.12	0.61	0.07	0.11	0.01	0.12	0.24

2.2 Physical pre-treatments

In the present study the focus was on AD performed on pre-treated OFMSW. The DoE investigated the pre-treatments and the AD, to evaluate, specifically for mechanical and thermal pre-treatments, the interactions between these two steps (pre-treatment and AD).

The mechanical pre-treatment was performed with the mixer blender (Aigostar Archer 30RKN, China) of 1.8 L, at three maximum speed values (15, 30, 45and 60 min) (Gagić et al., 2018), requiring 0.023 kWh/L.

The thermal pre-treatment was performed in the heating bath (Corio C Julabo, Merck, Germany) at three temperature values 60, 90 and 120 °C (Bruni et al., 2010) and for three time periods (15, 30 and 45 min), settled according to the mechanical pre-treatment. The energies to perform the pre-treatments were 0.040 kWh/L for T = 60 °C, 0.048 kWh/L for T = 90 °C and 0.059 kWh/L for T = 120 °C.

Two hydrodynamic-cavitation (HC) pre-treatments were performed using a rotor/stator HC unit (Rotocav®, E-PIC srl – Mongrando, Italy) at two temperatures (25 and 55 °C) for 10 min (Bruni et al., 2010). The two pre-treatments required 0.022 kWh/L and 0.073 kWh/L.

The ultrasound pre-treatment (US) was performed in a 3 L powerful multiprobe reactor (Weber Ultrasonics AG, Karlsbad - Germany) for 30 min at 22 Hz and 200 W (Lauberte et al., 2021), requiring 0.020 kWh/L.

2.2.1 Pre-treatment evaluation

The evaluation of each pre-treatment was performed through the Disintegration Rate (DR) (Eq.1 -2) (Bougrier et al., 2005).

$$DR_{COD}(\%) = \frac{SCOD_1 - SCOD_0}{TCOD - SCOD_0} \cdot 100 \quad (1)$$

$$DR_N\ (\%) = \frac{SN_1 - SN_0}{TN - SN_0} \cdot 100 \quad (2)$$

where $SCOD_0$ and $SCOD_1$ are the Soluble Chemical Oxygen Demand (SCOD) before and after pre-treatment, respectively; TCOD is the total COD; SN_0 and SN_1 are the soluble nitrogen before and after pre-treatment, respectively, and TN is the total nitrogen.

Total and soluble COD and total nitrogen were detected through a COD LCI 400 and a LCK 338 (HACH LANGE GHB, Germany) and quantified by a spectrophotometer 5000 D, (HACH, Canada).

2.3 Anaerobic digestion set up

AD was performed on OFMSW in 1.0 L Pyrex glass bottles (Duran, Germany) with a working volume of 80 %, at 37 °C, placed in a 55 L thermostatic water-bath (Julabo-Corio-C, Merck, Germany), operating in batch mode with 6 % total solids (TS) of OFMSW. Each bioreactor was manually shaken, and AD ended when biogas production was below 1%v/v of the total volume of biogas produced up to that time (Angelidaki et al., 2009).

Each bioreactor was connected by 6 mm Teflon tubes (PTFE, Germany) to a gasholder, made by 2 L Pyrex glass bottles (Duran, Germany). Biogas was analysed qualitatively by a biogas-analyser (GA5000, GMBH, Germany) and quantitatively by water displacement.

AD on not pre-treated OFMSW was performed as control to detect increase or decrease of the performances with respect to pre-treated OFMSW.

The WAS and CAS inocula, selected considering a previous study (Demichelis et al., 2022), were separately cultivated under anaerobic conditions at 37°C in 2 L Pyrex glass bottles (Duran, Germany), for three different periods (0, 5 and 10 d) and then inoculated in the pre-treated OFMSW considering the Substrate: Inoculum (SI) ratio ranging from 1:2 to 2:1 for mechanical and thermal pre-treatment and from 1:3 to 3:1 for HC and US, based on volatile solids (VS) (Demichelis et al., 2022). Since HC

and US pre-treatments were already optimised in (Lauberte et al., 2021), tests were performed on the effect of the SI ratio.

2.4 Analytical methods

The OFMSW and the two inocula (WAS and CAS), were physically and chemically characterized. TS and VS content were detected according to UNI EN 15216:2021 and elemental analysis (CHNSO) was performed through an Elemental Macro Cube system (Vario, Germany).

The VS removed at the end of AD was evaluated through Eq. 3 according to (Li et al., 2018):

$$VS\ removed\ (\%) = \frac{VS\ input - VS\ output}{VS\ input - (VS\ input \cdot VS\ output)} \cdot 100 \quad (3)$$

where VS removed is the percentage of removed volatile solids, VS input and VS output are the volatile solids concentrations in the feed substrate before and after AD.

The pH was measured according to DIN 38404 C5 methodology with a pH340 WTW pH-meter (Mettler Toledo, Germany).

2.5 Design of experiments

The adopted Design of Experiment (DoE) identified the role played by the factors, their interactions, and quadratic effects, and accomplished the optimization of the system with the final identification of the best conditions for process running.

For mechanical and thermal pre-treatments, the DoE involved the simultaneous study of factors related to pre-treatment and to AD to identify the effect eventually played by the interaction between these two phases.

For HC and US pre-treatments, the DoE involved only the optimization of AD since the pre-treatments were previously optimized (Calcio et al., 2018) (Lauberte et al., 2021).

Different DoEs were adopted to optimize the four pre-treatments since they differed from the number of factors to be studied; moreover, some practical constraints needed to be taken into account: i) the maximum number of experiments that could be run simultaneously, due to the number of available AD reactors; ii) the necessity to simultaneously run all the experiments related to a single pre-treatment to guarantee a lower experimental error; iii) the necessity to add some replications of the

experiments to evaluate the experimental error. See the supplementary material for the complete list of experiments established by DoE and the types of models investigated.

For AD, three experimental factors were considered:

- inoculum incubation (INOC), set at three levels: 0, 5 and 10 d (Demichelis et al., 2022).
- SI, set at: i) three levels (1:2, 1:1 e 2:1) for mechanical and thermal pre-treatments, according to the ones investigated in (Demichelis et al., 2022), and ii) at five levels (1:3, 1:2, 1:1, 2:1, 3:1) for HC and US according to (Liu et al., 2019) and (Kawai et al., 2014).
- origin of inoculum (ORIG), set at two levels, namely WAS and CAS (Demichelis et al., 2022).

ORIG was a qualitative factor, hence the experiments identified by the DOEs were repeated for CAS and WAS independently.

The DoE investigated the performances of the biodegradation of OFMSW; and specific biogas production (NL/kg vs) and ESI (-) were selected as the experimental responses to be modelled. The response surface methodology provided the best experimental conditions through a grid search algorithm exploring the obtained models in the overall experimental domain (scaled in the range [0,1] for each factor) with a step of 0.1 for each factor included in the model.

2.5.2 DoE for mechanical and thermal pre-treatments

For mechanical and thermal pre-treatments, the DOE included both the pre-treatment and the AD. For the AD, INOC was studied at three levels (0, 5 and 10 d), SI at three levels (1:2, 1:1 and 2:1) and ORIG at two qualitative levels (WAS and CAS) (Demichelis et al., 2022).

For the pre-treatments, the experimental factors were settled according to the study of DR_{COD} and DR_N developed in the present study (section 3.2):

- For the mechanical pre-treatment: the time of pre-treatment ~~was added~~ (PT) was studied at three levels (15, 30 and 45 min)
- For the thermal pre-treatment two factors were added: the temperature (TEMP) at three levels (60, 90 and 120° C) and the time of pre-treatment (PT) at three levels (15, 30 and 45 min).

For the mechanical pre-treatment a central composite design was adopted with two replications of the centre of the domain to evaluate the experimental error. The resulting 16 experiments are reported in Table S1.

For the thermal pre-treatment, a fractional factorial design (FFD) 2^{4-1} was adopted, where the fourth factor (SI) was confused with the interaction between the first two (PT*TEMP), providing a total of 2^3=8 experiments. A star design was added, providing 2p+1 = 2*4+1=9 experiments (p being the number of factors), to evaluate the quadratic effects. Two more replications of the centre of the domain were added to evaluate the experimental error, providing a total of 20 experiments, reported in Table S2.

2.5.3 DoE for hydrodynamic cavitation and ultrasound

Since HC and US pre-treatments were previously optimised (Calcio et al., 2018) (Lauberte et al., 2021), only AD was investigated considering the SI ratio at five levels (1:3, 1:2, 1:1, 2:1, 3:1) according to (Liu et al., 2019) and (Kawai et al., 2014), while INOC was studied at three levels (0, 5 and 10 d). The DoE involved all the possible combinations of the levels for both factors, providing a total of 5*3=15 experiments with one more experiment consisting in a replication of the centre of the domain to evaluate the experimental error. The resulting 16 experiments are reported in Table S3. For HC, the DoE was repeated independently at 25 and 55 °C.

2.6. Calculation of Regression Models

For each DoE, regression models were assessed relating biogas production and ESI to the investigated factors, their interactions, and their quadratic effects, independently for CAS and WAS origins. Only statistically relevant (α-level < 0.05) coefficients, identified by ANOVA (Analysis Of Variance), were included in the final models (Box, G., Hunter, S., 2005). See Supplementary Material for the description of the coefficients included in each evaluated model.

2.7 Kinetic study

The kinetics of AD was studied to evaluate the disintegration rate (k_d) and the biogas volumetric rate. The k_d was calculated by a first-order kinetic model Eq. (4):

$$B(t) = B_{exp}(1 - e - k_d t) \quad (4)$$

where $B(t)$ is the cumulative methane production at given time t (d), B_{exp} represents the ultimate methane potential yield (NL/kg vs) at the 5th day, k_d is the first-order disintegration rate (1/d) and t is the time of the process (1/d).

The biogas volumetric rate was calculated through Eq.5

$$V\ biogas\ rate\ \left(\frac{L}{L \cdot d}\right) = \frac{Biogas\ (L)}{Volume\ of\ reactor\ (L) \cdot time\ (d)} \quad (5)$$

2.9 Energy sustainable index

The energetic sustainability of AD was measured with the energy sustainable index (ESI) calculated according to (Kovalovszki et al., 2020) and Eq. (6):

$$ESI = \frac{Q_{pro}}{Q_s} \quad (6)$$

where Q_{pro} was the energy produced from AD, considering that methane equals to 7.2 kWh/m^3 (Rillo et al., 2020) and Q_s is the system thermal load measured in kWh (Eq. 7), and corresponded to the sum of the thermal power required for heating the substrate (Q_{sub}), the heat loss from the reactor walls (Q_{loss}), the heat loss through the tube (Q_p), according to (Mehr et al., 2017) and the energy consumed to perform the pre-treatments ($Q_{pre-treatment}$) with the specific consumption reported in sections 2.2 for each type of pre-treatment.

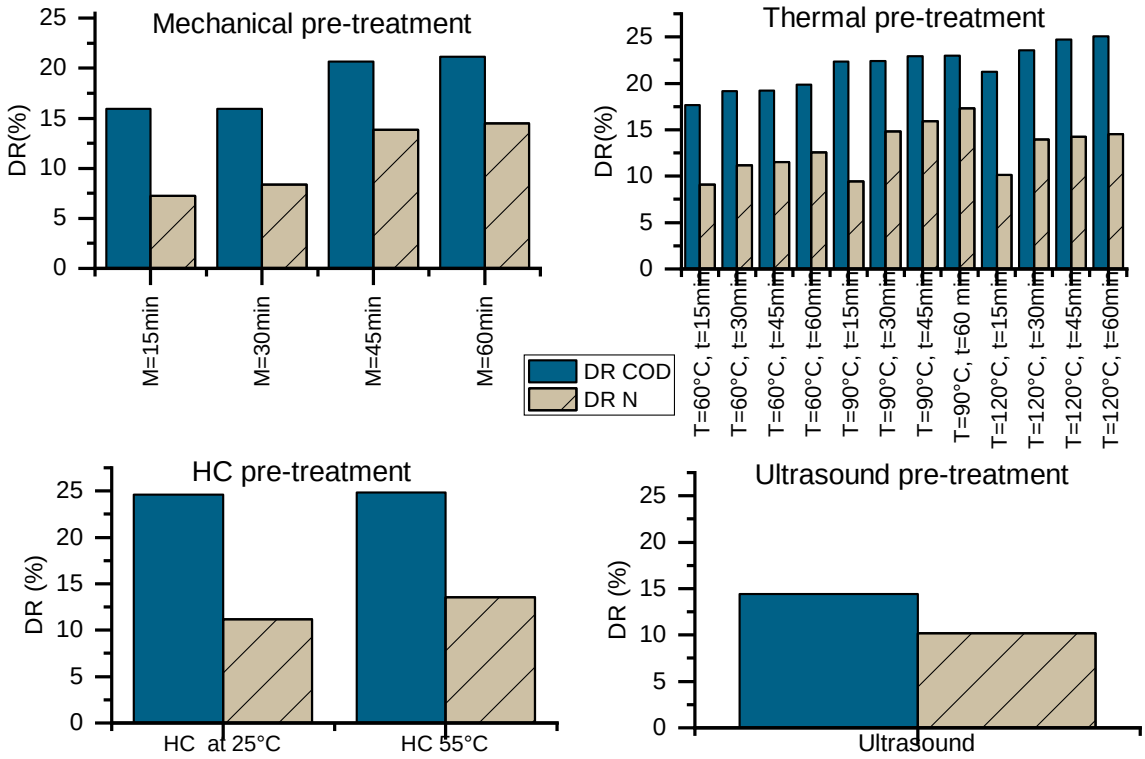
$$Q_s = Q_{sub} + Q_{loss} + Q_p + Q_{pre-treatment} \quad (7)$$

3.Results

3.1 Disintegration rate

The study of the DR, calculated for COD and nitrogen (Figure 1), was performed to establish the suitable experimental domain for the factors included in the study of the four investigated pre-treatments, in terms of improvement of the available degradable matter (increase of the solubilisation of ready-digestible compounds of the OFMSW).

295 For the mechanical pre-treatment, the highest DR values were achieved after 45 min of pre-treatment,
296 whereas a further increase to 60 min did not show any statistically significant improvement of DR
297 and for the AD tests, a pre-treatment time of 60 min was not considered.
298 For the thermal pre-treatment, the highest DR values were achieved at the highest tested temperature
299 (120 °C) after 45 min of pre-treatment. By increasing the thermal pre-treatment time to 60 min, for
300 all the tested temperatures (60, 90 and 120 °C), no statistically significant differences were detected
301 ($\alpha - level$ = 0.05), and a pre-treatment time of 60 min was not investigated in AD tests.
302 The results proved that, for thermal pre-treatments, the temperature played a more significant effect
303 than time, according to (Gagić et al., 2018).
304 No statistical differences were detected in DR calculated for COD and nitrogen, comparing HC at 25
305 and 55 °C; and these two configurations were tested.
306 Notwithstanding the worst results achieved by the US pre-treatment on DR for COD and nitrogen
307 (14.41 % and 10.21 %), compared to the other treatments, it was included in the experimentation
308 since it is considered as promising according to the literature (Lauberte et al., 2021).
309 Among the tested pre-treatments, those reaching the highest DR_{COD}, were: thermal pre-treatment at
310 120 °C for 45 min (27.85 %) and HC at 55 °C (27.92 %) and 25 °C (27.86 %), due to the synergic
311 effect of temperature and time, and for HC pre-treatments, the formation of extremely reactive
312 microenvironments generated inside the bubbles, characterized by intense pressure waves and
313 hydraulic jets, and responsible of a series of chemical and physical transformations in the OFMSW.
314 The DR indirectly describes the efficiency of degradation of complex organic substrates, but it only
315 quantifies the performances of the pre-treatment, neglecting its effect on AD.



316

317 Figure1: Disintegration rate (DR) of pre-treatments calculated for COD and nitrogen.

318 **3.2 Biogas production**

319 The AD performance was investigated through the evaluation of the productions of biogas and
320 methane (Figure 2a), and the VS removal (Figure 2b), according to (Li et al., 2019), and through the
321 response surfaces of the models calculated for biogas production (Figure 3).

322 **3.2.1 Mechanical pre-treatment**

323 In the case of AD of mechanically pre-treated OFMSW performed with WAS and CAS, the
324 productions of biogas and CH_4 (Figure 2a: A1,A2) and the VS removal (Figure 2b: A1, A2) showed
325 similar trends.

326 The surface responses (Figure 3A1 and A2) and the built models (Table S4) proved that the interaction
327 between INOC and SI was relevant; and that the effect of INOC was similar for WAS and CAS, but,
328 the models for biogas production with WAS showed higher R^2 value (R^2= 0.9269, Table ~~5~~S4) than
329 CAS (R^2=0.8161), due to the absence of the SI parameter for the CAS origin ($\alpha < 0.05$). The biogas
330 production (Figure 3A1 and A2) improves by increasing INOC at high SI (2:1). The quadratic effect

of INOC was also evident, while neither the linear nor the quadratic effect of the pre-treatment was relevant.

The best experimental conditions, identified by the grid search algorithm (Table ~~9~~ 2), were, for the two inocula, at SI = 2:1, with INOC = 10 d and pre-treatment time of 15 min. For these configurations, biogas and methane productions and VS removal were in the range: 695.46-699.17 NLbiogas/kg_{VS}, 475.66-482.45 NL CH_4/kg_{VS} (Figure 2a: A1, A2), and 71.99- 73.52 % w/w for VS (Figure 2b: A1, A2), which were higher than the ones reached in (Zhang and Banks, 2013) (VS removal in the range 57-64 % w/w for mesophilic AD of mechanically pre-treated OFMSW) due to the higher INOC (from 5 to 10 d).

The results proved that the particle size reduction did not notably increase the extent of degradation, because the important aspect was the formation of a homogeneous substrate to feed AD without impurities (Jain et al., 2015). The excess of smaller particles could lead to acid accumulation inside the digester as proven by (Panigrahi et al., 2020); indeed, AD of mechanically shredded OFMSW reached negligible different CH_4 content by reducing OFMSW particle size from 4 to 2 mm (respectively 0.34 and 0.31 Nm^3/kg_{VS}).

3.2.2 Thermal pre-treatment

For AD of thermally pre-treated OFMSW with WAS and CAS inocula, the biogas and CH_4 productions (Figure 2a: B1, B2) and VS removal (Figure 2b: B1, B2) showed similar trends.

The models for biogas production exhibited similar R^2 values for CAS (R^2= 0.9630) and WAS (R^2 = 0.9644), and, for both inocula, PT showed a significant interaction with the INOC (Figure 3B1, Table S5), while it was negligible as linear factor. This result was due to the higher buffering capacity of the incubated inoculum, which produced the acclimatised micro-organisms able to biodegrade the OFMSW (Zhang et al., 2019a).

Figures 3B1, B2, and B3 represent only the response surfaces for WAS (the factors not represented are in turn fixed at the central value), since the two origins showed similar response surfaces. The

quadratic effect of INOC was evident, and changes in INOC correspond to the most significant increases of the experimental response.

With the two inoculum origins, the configurations with pre-treatment at the highest temperature (120 °C) and the highest INOC (10 d) showed statistically negligible differences if pre-treatment was performed for 15 or 45 min, because at the highest temperature, the pre-treatment time was negligible due to the solubilisation and degradation effects played by the temperature (Li et al., 2017). This result is in agreement with the literature: the increase of temperature promoted the feedstock conversion degree, and the pre-treatment temperature affected AD performances more than pre-treatment time according to (Gagić et al., 2018). Thermal pre-treatment is usually carried out at a higher temperature (from 150 to 200 °C), but its main drawback is the high energy requirement, which usually cannot be balanced by the high biogas production, leading to the consequential reduction of the economic overall profitability of the process (Rajput et al., 2018). The optimal conditions identified through the grid search algorithm (Table 2) were, for the two inocula, at SI= 2:1, INOC = 10 d, and pre-treatment at 120 °C for 15 min: the biogas production was predicted to be equal to 665 NLbiogas/kg_{VS} in these conditions, which were not included in the DoE.

3.2.3 HC pre-treatment

The AD performances on HC-treated OFMSW at 25 (Figure 3: C1, C2) and 55 °C (Figure 3: D1, D2) were similar.

The model for biogas with CAS reached R^2= 0.9394 at 25 °C and 0.9305 at 55 °C, higher than those obtained with WAS (R^2=0.8474 at 25 and 55° C, Table S6). For CAS, the model contained all the parameters, whereas, for WAS, the model excluded SI and the quadratic effect of INOC. For all the models, all the parameters were statistically significant ($\alpha < 0.05$).

The biogas productions of HC at 25 and 55 °C with CAS (Fig.~~2~~3: C2 and D2) provided comparable results: a quadratic effect was evident for SI and INOC, proving their synergic effect in increasing the biogas production. The best conditions identified by the grid search algorithm were at high levels

of INOC, equal to 8 d for HC at 25 °C (predicted biogas production = 700.62 NL/kg $_{vs}$) and 9.5 d for HC at 55 °C (predicted biogas production = 798.07 NL/kg $_{vs}$), at SI = 3:1.

For HC at 25 and 55 °C with WAS (Fig.3 C1 and D1), the maximum biogas production occurred at high INOC values (from 5 to 10 d) and medium-high value of SI (from 2:1 to 3:1), due to its evident quadratic effect. The best conditions identified by the grid search algorithm (Table 2), for HC, at 25 and 55 °C, with WAS were: INOC = 10 d at SI = 2.74 (predicted biogas production = 670.44 NL/kg vs. for HC at 25°C and 705.36 NL/kg vs for HC at 55°C).

The AD of HC-OFMSW at 55 °C achieved higher biogas and CH_4 productions (Figure 2a: C, D) than HC at 25 °C, since during the collapse phase realised by HC, the highest temperature promoted the formation of more reactive microenvironments which boosted the diffuse turbulence, the phase changes, and the heat exchanges, occurring from macro to micro scales (Calcio et al., 2018).

3.2.4 Ultrasound pre-treatment

For US pre-treatment, the biogas and CH_4 productions and VS removal had similar trends for AD performed with WAS and CAS inocula, but the AD performed with CAS inoculum reached higher performances (Figure 2a: E; Figure 2b: E).

The model for biogas with CAS reached higher R^2 value (R^2= 0.9286, Table S7) than WAS (R^2=0.8756), with a model containing INOC, the quadratic effect of SI and the interaction between INOC and SI for both origins ($\alpha < 0.05$). By increasing INOC, the biogas production increased independently of SI (Figure 3: E1, E2), whereas SI showed a significant quadratic effect at high INOC values. These results proved the suitability of the incubated inoculum to treat higher OFMSW amounts and to promote its degradation (Zhang et al., 2019a).

The optimal conditions identified by the grid search algorithm corresponded to INOC =10 d: at SI = 2.74 for WAS (predicted biogas production = 683.95 NL/kg vs) and SI= 2.34 for CAS (predicted biogas production = 703.63 NL/kg vs) (Table ~~9~~2). These results agreed with (Rasapoor et al., 2016) where US was carried out on 6 % TS OFMSW for 30 min at 200 kHz.

406 Among the tested pre-treatments, US reached the lowest biogas and methane productions and VS
407 removals because probably the lipids accumulation negatively influenced AD since their degradation
408 was relatively slow and led to an accumulation of hydrophobic lipids which were adsorbed on the
409 microorganisms' surface (Hendriks and Zeeman, 2009) with the effect of limiting the mass transfer
410 process between microbial cells and dissolved organic matter (Scherzinger and Kaltschmitt, 2021).
411 The decrease of biodegradability after pre-treatment occurred for two main effects: formation of
412 refractory/toxic compounds (Carrère et al., 2009) and removal of organic material (Hendriks and
413 Zeeman, 2009). The US pre-treatment of the lignocellulosic fraction of OFMSW can release
414 hydroxymethylfurfural (HMF), furfural, and soluble phenols (Hendriks and Zeeman, 2009), or
415 produce melanoidins by uncompleted Maillard reactions (Pilli et al., 2011), which inhibit the AD.

416 **3.2.5 Comparison of the physical pre-treatments and no pre-treated OFMSW**

417 The performances of AD on physical pre-treated OFMSW were higher than those on untreated
418 OFMSW tested in (Demichelis et al., 2022) in 0.5 L bioreactors and re-tested in the present
419 manuscript in 2 L bioreactors to evaluate the scale effect (Figure S1).
420 The biogas production of mechanically pre-treated OFMSW was higher than untreated OFMSW, in
421 the ranges 3.0 and 7.3 % in agreement with (Coarita et al., 2020), 4.6- and 9.8 % for mild-thermal
422 pre-treatments in agreement with (Chen et al., 2020), 7.8 and 11.8 % with HC according to (Saxena
423 et al., 2019), 2.5- and 4.7 % with US as proved by (Rasapoor et al., 2016), since pre-treatments
424 increase the exposure of the biodegradable matter to microorganisms and vary the composition of
425 hardly degradable matter (Zhen et al., 2017).

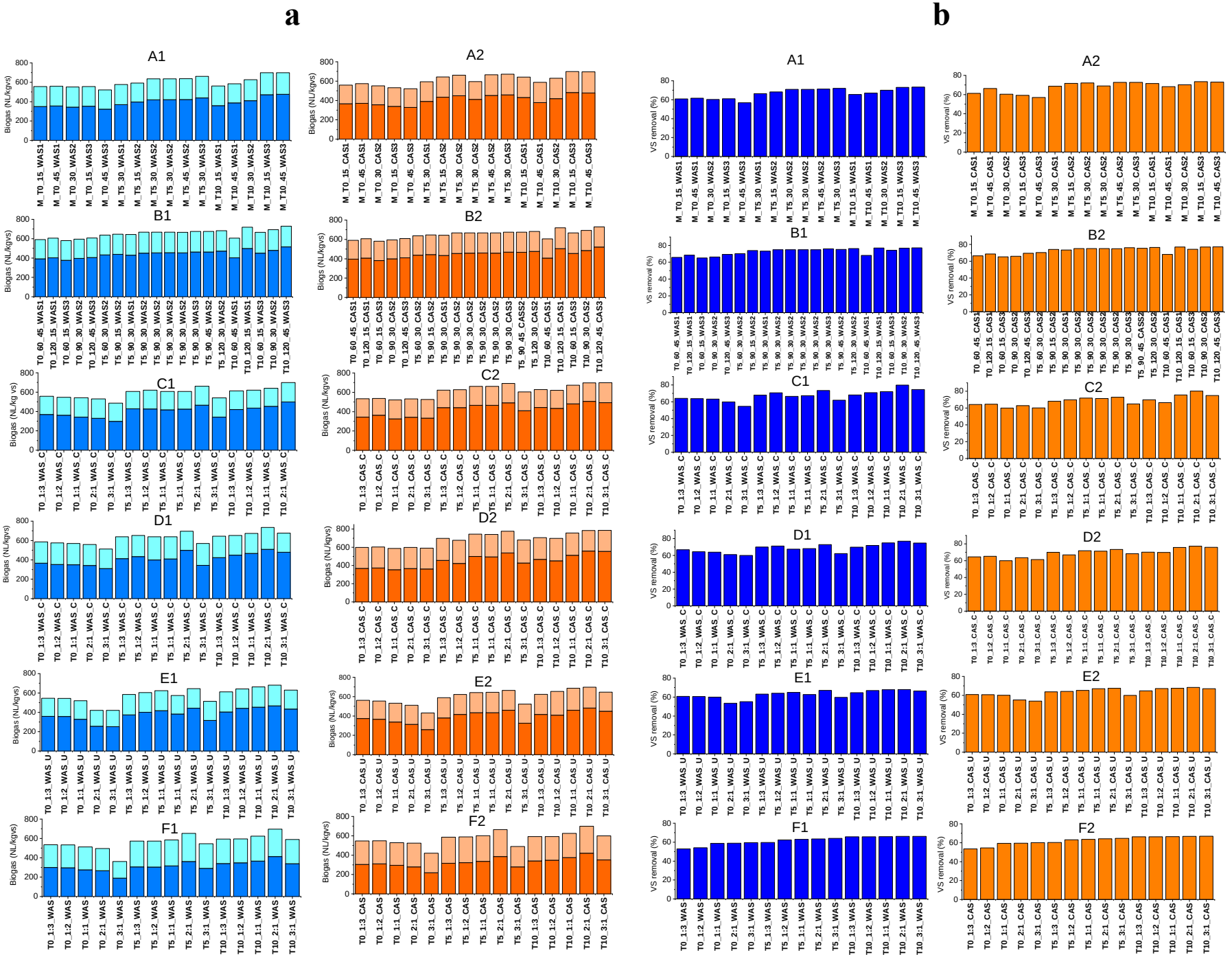
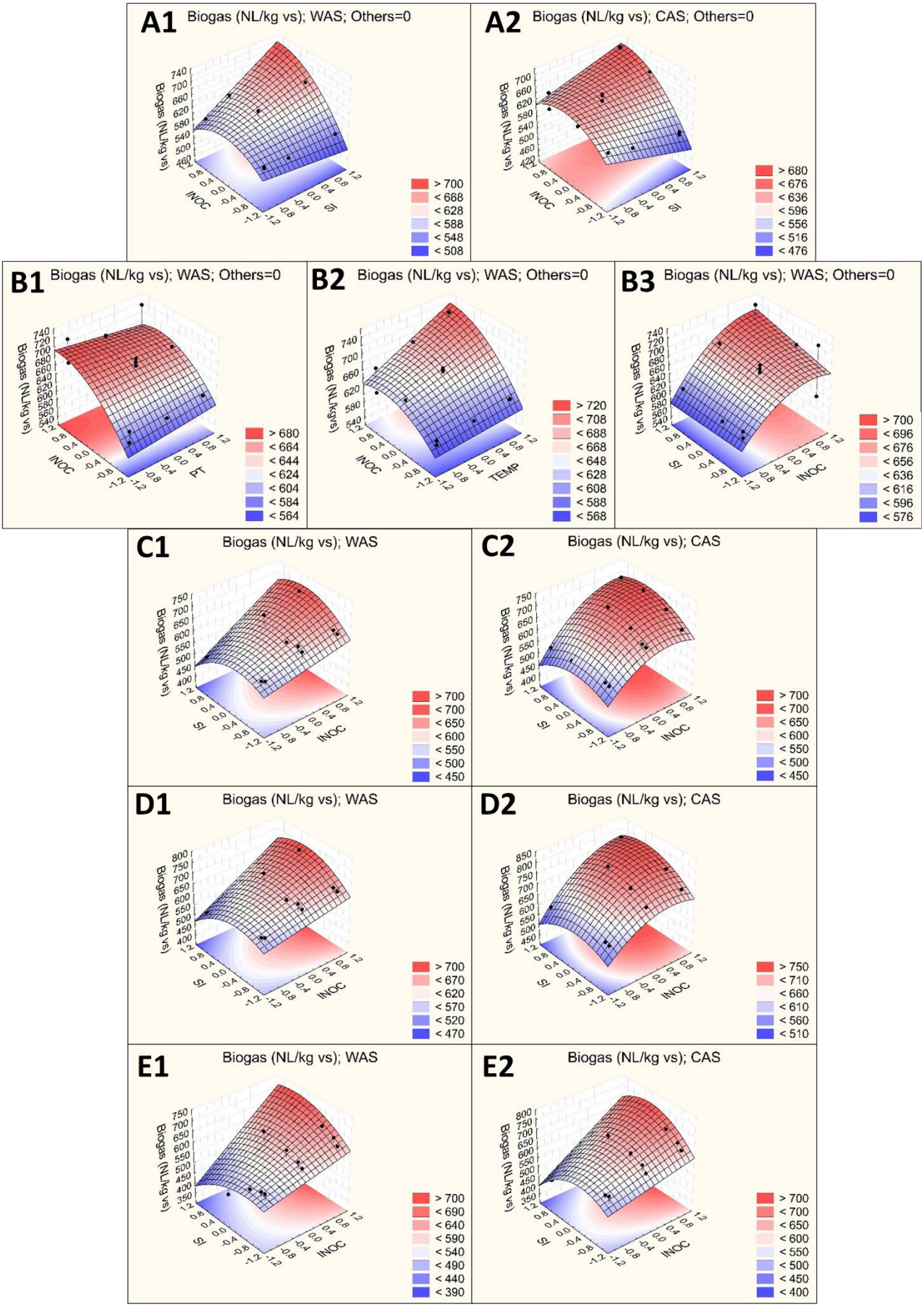

a
b
A1
A2
B1
B2
C1
C2
D1
D2
E1
E2
F1
F2
Biogas (NL/kgvs)
VS removal (%)

Figure 2a: Biogas production of AD on pre-treated and not pre-treated OFMSW: mechanical (A), thermal (B), hydrodynamic cavitation (HC) at 25°C (C), (HC) at 55 °C (D), ultrasound (E), no pre-treatment (F). On the left, AD configurations carried out with inoculum WAS are reported (dark blue is methane, light blue is carbon dioxide), while those carried out with the CAS origin are reported on the right (dark orange is methane, light orange is carbon dioxide).

Figure 2b: Volatile solids removal of AD performed on pre-treated and not pre-treated OFMSW: mechanical (A), thermal (B), HC at 25 °C (C), HC at 55 °C (D), ultrasound (E), no pre-treatment (F). On the left, AD configurations carried out with inoculum WAS are reported (blue), while those carried out with the CAS origin are reported on the right (orange).

437


A1
Biogas (NL/kg vs); WAS; Others=0
Biogas (NL/kg vs)
INOC
SI
> 700
< 668
< 628
< 588
< 548
< 508
A2
Biogas (NL/kg vs); CAS; Others=0
Biogas (NL/kg vs)
INOC
SI
> 680
< 676
< 636
< 596
< 556
< 516
< 476
B1
Biogas (NL/kg vs); WAS; Others=0
Biogas (NL/kg vs)
INOC
PT
> 680
< 664
< 644
< 624
< 604
< 584
< 564
B2
Biogas (NL/kg vs); WAS; Others=0
Biogas (NL/kg vs)
INOC
TEMP
> 720
< 708
< 688
< 668
< 648
< 628
< 608
< 588
< 568
B3
Biogas (NL/kg vs); WAS; Others=0
Biogas (NL/kg vs)
SI
INOC
> 700
< 696
< 676
< 656
< 636
< 616
< 596
< 576
C1
Biogas (NL/kg vs); WAS
Biogas (NL/kg vs)
SI
INOC
> 700
< 700
< 650
< 600
< 550
< 500
< 450
C2
Biogas (NL/kg vs); CAS
Biogas (NL/kg vs)
SI
INOC
> 700
< 700
< 650
< 600
< 550
< 500
< 450
D1
Biogas (NL/kg vs); WAS
Biogas (NL/kg vs)
SI
INOC
> 700
< 670
< 620
< 570
< 520
< 470
D2
Biogas (NL/kg vs); CAS
Biogas (NL/kg vs)
SI
INOC
> 750
< 710
< 660
< 610
< 560
< 510
E1
Biogas (NL/kg vs); WAS
Biogas (NL/kg vs)
SI
INOC
> 700
< 690
< 640
< 590
< 540
< 490
< 440
< 390
E2
Biogas (NL/kg vs); CAS
Biogas (NL/kg vs)
SI
INOC
> 700
< 700
< 650
< 600
< 550
< 500
< 450
< 400

438

439 Figure 3: Response surfaces of biogas production for the pre-treatments: mechanical (A1, WAS; A2,
440 CAS), thermal (B1, B2, B3, WAS). HC at 25°C (C1, WAS; C2, CAS), HC at 55°C (D1, WAS; D2,
441 CAS) and ultrasound (E1 for WAS; E2 for CAS). For thermal pre-treatment only the response
442 surfaces for WAS are depicted since the models for WAS and CAS are almost identical.

443 **3.3 Kinetic study**

444 The kinetic study proved the importance of the incubation time and of the origin of the inoculum
445 (Figure S1).

446 AD performed with INOC equal to 5 and 10 d promoted hydrolysis, acidogenesis, acetogenesis and
447 methanogenesis without inhibition, achieving the faster k_d and volumetric biogas rates.

448 The origin of the inoculum influenced the k_d, and CAS inoculum exhibited a faster degradation rate
449 than WAS, in agreement with (Kumar Biswal et al., 2020): because the proper C:N ratio of CAS
450 supported a correct development of the AD process (Calcio et al., 2018), and the incubation of the
451 inoculum provided acclimatised micro-organisms able to biodegrade the OFMSW (Zhang et al.,
452 2019a).

453 In all the tested pre-treatments, the values of k_d and biogas volumetric rate were linearly correlated.

454 In AD of mechanically pre-treated OFMSW, the highest volumetric biogas rate and k_d were achieved,
455 for the two inocula, by increasing INOC (5 and 10 d) and SI, without a significant effect of the
456 extension of the PT (k_d =0.33 and 0.50 1/d), as proved by (Gagić et al., 2018). The study of (Motte et
457 al., 2015) stated that the fine milling of organic waste may simultaneously increase the AD kinetic
458 and failure (Victorin et al., 2020), whereas in the present study these risks were limited by the
459 presence of the incubated inoculum.

460 For AD of thermally pre-treated OFMSW, the highest volumetric biogas rate and k_d were achieved
461 by AD performed with INOC =10 d, at SI = 2:1 and pre-treatment at 120 °C for 45 min: 3.31 with
462 WAS and 3.33 NL/kgvs·d with CAS (L. Zhang et al., 2019), and k_d equal to 0.51 and 0.53 1/d with
463 WAS and CAS respectively (Zhang et al., 2019b). These results proved that kinetic values increased
464 by increasing INOC (from 0 to 10 d) and pre-treatment temperature (Li et al., 2017), because the
465 incubation promoted the formation of acclimatised micro-organisms, while the temperature boosted
466 the solubilisation of the OFMSW improving its bio-degradation.

467 AD performed on HC-OFMSW at 25 and 55 °C with the two inocula, showed the same kinetic
468 configuration trends: the volumetric biogas rate and k_d increased by increasing the INOC (from 0 to

10 d and the HC temperature, due to the simultaneously effect of the inoculum incubation and the higher capacity of extraction of bioactive compounds characteristic of HC performed at higher temperature (Calcio et al., 2018).

AD of US pre-treated OFMSW reached the highest kd and volumetric biogas rates with the highest INOC = 10 d, notwithstanding the value of SI, for the two inocula.

For AD performed with incubated inocula, the k_d and volumetric biogas rates increased by increasing the SI ratio, because the inoculum with acclimatised micro-organisms could be employed ~~in~~ with a lower amount than a non-incubated inoculum (Zhang et al., 2019a).

The k_d of the four types of pre-treatments varied between 0.1 and 0.58 1/d , ranging from the lowest to the highest specific biogas production, according to the optimal range of 0.134-0.56 1/d stated by (Li et al., 2018). The k_d of carbohydrates ranged from 0.5 to 2.0 1/d, proteins varied between 0.25-0.8 1/d and lipids ranged between 0.1and 0.7 1/d , (Victorin et al., 2020). In the present study, mechanical, thermal and US pre-treatments could promote the release of carbohydrate compounds, whereas HC pre-treatment, a mix of carbohydrates and lipids.

The inocula incubation provided the optimal consortium of anaerobic microbes able to prevent inhibition, due to the higher SI ratio (from 3:1 to 2:1) (Browne and Murphy, 2013), whereas the non-incubated inocula negatively affected the lag phase (λ) and the maximum specific biogas production (Dasgupta and Chandel, 2019).

Among the tested AD configurations, AD of HC-OFMSW reached the highest k_d, since HC is a promising strategy to overcome the non-degradability of the recalcitrant components in AD (Naran et al., 2016) (Saxena et al., 2019).

3.4 Energy sustainable index

3.4.1 Mechanical pre-treatment

The ESI major than 1 for AD performed on mechanically pre-treated OFMSW was reached by the same ~~four~~ configurations with WAS and CAS, (Figure 4A1-A2, and Figure S2 and Table S4) but different models were obtained for AD performed with CAS and WAS ~~-~~(Table S4). The model for WAS reached a lower R^2 (equal to 0.8963) and contained only INOC, while the model for CAS contained INOC, PT and two interactions (SI * PT and SI * INOC), proving the relevant interaction between the pre-treatment phase and the AD (R^2 = 0.9369).

In the case of WAS, no surface responses are provided since the model was simple and the best conditions were obtained with high values of INOC (10 d) notwithstanding the values applied for PT and SI; these two factors can be therefore kept at the most convenient value equal to PT =15 min at SI= 2:1.

In the case of the CAS origin, looking at the ~~For the~~ surface responses , ~~for CAS~~ (Figure ~~3~~4: A1, A2), considering INOC*SI, the ESI increased by increasing INOC at high and low SI, since INOC (Table S4) played the most significant role (Zhang et al., 2019a).

The best configuration stated by the grid search algorithm (Table 2) was, for the two inoculum origins, PT ~~=~~15 min, SI =2:1and INOC =10 d, with a calculated response of 1.17 and 1.01 for CAS and WAS (close to the experimental ones, 1.14 and 1.04).

3.4.2 Thermal pre-treatment

The ESI of AD on thermally pre-treated OFMSW reached the same trends with WAS and CAS (Figure S2, 4B).

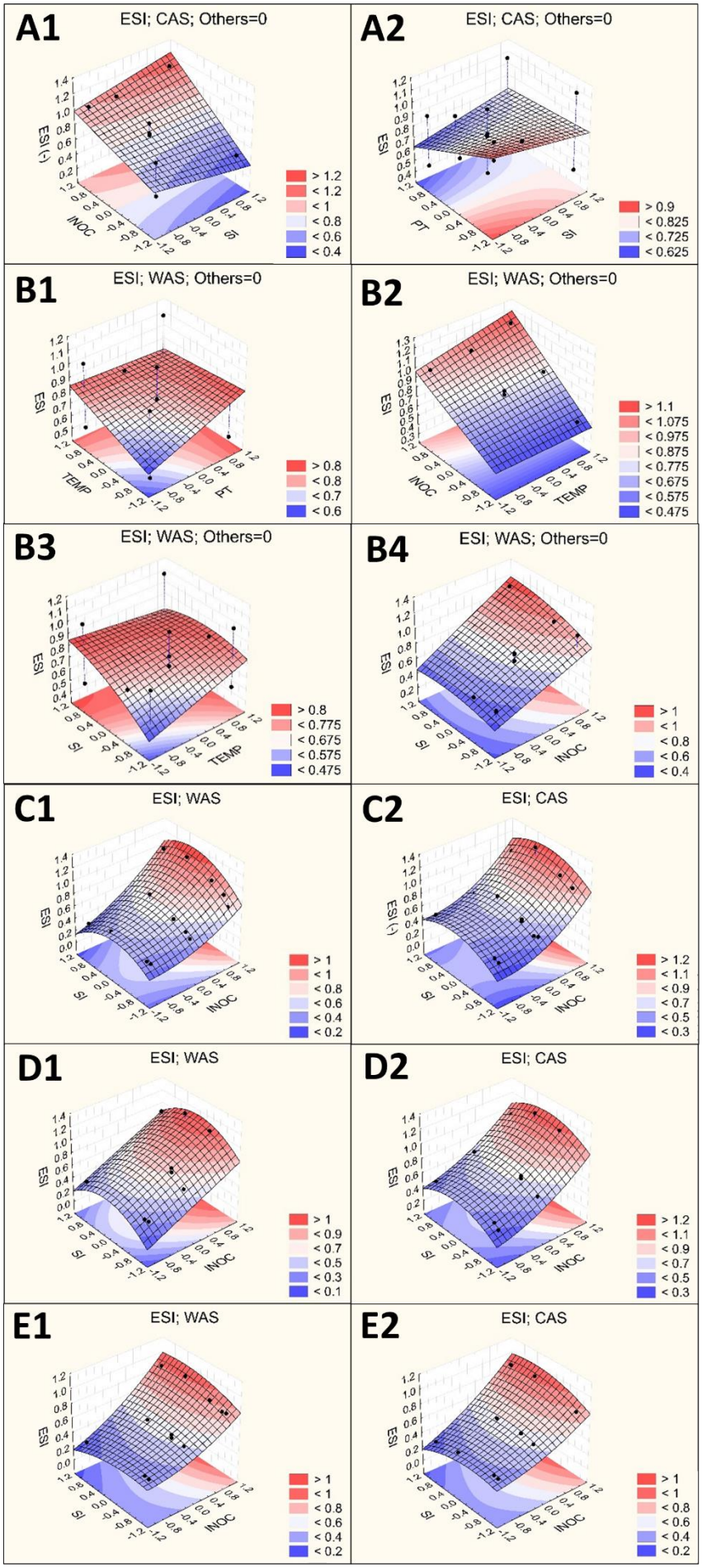
For thermal pre-treatment, the models with CAS and WAS reached high R^2 values (R^2= 0.9882 for WAS and R^2 = 0.9878 for CAS, Table ~~6~~ S5), and contained the same parameters ($\alpha < 0.05$). The two models were similar (Table S5), therefore, the response surfaces for the four significant interactions were reported just for WAS (Figure 4: B1-B4). Considering the interactions, the trend of TEMP*PT (Figure 4 B1) and TEMP*SI (Figure 4 B3) was similar one to each other and the same can be observed

517 for INOC*TEMP (Figure 4 B2) and INOC*SI (Figure 4 B4). Increasing TEMP, the ESI increased at
518 low PT values, according to (Rittmann et al., 2018), while an increase of PT increased ESI at low
519 TEMP values, without producing relevant effects at high TEMP values, in agreement with (Menardo
520 et al., 2015).
521 This result proved that the pre-treatment temperature was more effective than time since the
522 temperature boosted the solubilisation of the OFMSW reducing the ammonia concentration as the
523 result of caramelization or Maillard reactions occurring at temperature above 90 °C, preserving the
524 AD process.
525 The best configurations identified by the grid search algorithm for thermal pre-treatment (Table 2)
526 were AD with INOC=10 d at SI = 2:1, TEMP = 120°C and PT = 45 min for CAS (predicted ESI =
527 1.20), followed by AD with INOC = 10 d and SI = 1.74, TEMP =120 °C and PT = 45 min for WAS
528 (predicted ESI = 2.13).
529 **3.4.3 HC pre-treatment**
530 For HC at 25 °C, the models with the two inoculum origins reached high R^2 values (R^2= 0.9635 for
531 WAS and R^2 = 0.9390 for CAS, Table 2), whereas for HC at 55 °C, the model with CAS reached a
532 R^2 value higher than WAS (R^2= 0.9455 for CAS, R^2 = 0.8842 for WAS, Table 2).
533 The best conditions of ESI (Table 2, Figure 4: C1, C2, D1, D2) were: INOC = 10 d and SI values
534 ranging from 3:1 (WAS) to 2.87 (CAS) for HC at 25 °C and from 2.6 (WAS) and 2.74 (CAS) for HC
535 at 55 °C. The predicted ESI values ranged between 1.07 and 1.12. These results proved the importance
536 of temperature on the HC performance (Calcio et al., 2018).
537 The energy cost to carry out HC was covered by the biogas surplus produced by AD of HC-OFMSW,
538 assessing HC as an effective pre-treatment (Saxena et al., 2019). The HC had a positive ESI due to
539 lower plant and operating costs, higher process yields and energy savings due to shorter process times
540 (Calcio et al., 2018).
541 **3.4.4 Ultrasound pre-treatment**

All tested configurations for AD carried out on ultrasound pre-treated OFMSW, with CAS and WAS, were energetically unsustainable (Figure S2E).

The model with WAS and CAS contained all the parameters (R^2= 0.9832 for WAS, R^2=0.9775 for CAS, Table S7). The response surfaces (Figure 4: E1, E2) with WAS and CAS, proved that the increase of INOC increased ESI notwithstanding SI values, whereas ESI increased and reached almost a plateau by increasing SI at high INOC.

The best conditions identified by the grid search algorithm (Table 2) corresponded to INOC =10 d and SI =3:1 for both origins, but the calculated responses with these conditions reached ESI<1. The energetic unsustainability of AD performed on US pre-treated OFMSW was due to the high energy required to carry out the pre-treatments and lower CH_4 produced during AD (59.80 and 68.90 %v/v with WAS and 60.34 and 69.52 %v/v with CAS) compared to the other pre-treatments.


A1
ESI; CAS; Others=0
ESI (-)
INOC
SI
> 1.2
< 1.2
< 1
< 0.8
< 0.6
< 0.4
A2
ESI; CAS; Others=0
ESI
PT
SI
> 0.9
< 0.825
< 0.725
< 0.625
B1
ESI; WAS; Others=0
ESI
TEMP
PT
> 0.8
< 0.8
< 0.7
< 0.6
B2
ESI; WAS; Others=0
ESI
INOC
TEMP
> 1.1
< 1.075
< 0.975
< 0.875
< 0.775
< 0.675
< 0.575
< 0.475
B3
ESI; WAS; Others=0
ESI
SI
TEMP
> 0.8
< 0.775
< 0.675
< 0.575
< 0.475
B4
ESI; WAS; Others=0
ESI
SI
INOC
> 1
< 1
< 0.8
< 0.6
< 0.4
C1
ESI; WAS
ESI
SI
INOC
> 1
< 1
< 0.8
< 0.6
< 0.4
< 0.2
C2
ESI; CAS
ESI (-)
SI
INOC
> 1.2
< 1.1
< 0.9
< 0.7
< 0.5
< 0.3
D1
ESI; WAS
ESI
SI
INOC
> 1
< 0.9
< 0.7
< 0.5
< 0.3
< 0.1
D2
ESI; CAS
ESI
SI
INOC
> 1.2
< 1.1
< 0.9
< 0.7
< 0.5
< 0.3
E1
ESI; WAS
ESI
SI
INOC
> 1
< 1
< 0.8
< 0.6
< 0.4
< 0.2
E2
ESI; CAS
ESI
SI
INOC
> 1
< 1
< 0.8
< 0.6
< 0.4
< 0.2

554

Figure 4: Response surfaces of ESI for the four physical pre-treatments: mechanical (A1, A2), thermal (B1, B2, B3, B4), HC at 25°C (C1, C2), HC at 55°C (D1, D2) and ultrasound (E1, E2). For the mechanical and thermal pre-treatments only the response surfaces for WAS are reported since the models for WAS and CAS are almost identical.

Table 2: Response prediction of AD of pre-treated OFMSW (/) is reported when the parameter was not included in the optimization; - is reported when the parameter can be kept at any level).

			Best conditions ranged between -1 and +1					**Best conditions reported in the original measure units**			
			PT	**TEMP**	**INOC**	**SI**	**Y best**	**PT (min)**	**TEMP (°C)**	**INOC (d)**	**SI**
Mechanical	**Biogas**	WAS	-	/	1	1	703.00	-	/	10	2:1
		CAS	-	/	1	1	684.70	-	/	10	2:1
	ESI	WAS	-	/	1	-	1.01	-	/	10	-
		CAS	-1	/	1	1	1.17	15	/	10	2:1
Thermal	**Biogas**	WAS	-1	1	1	1	752.70	15	120	10	2:1
		CAS	-1	1	1	1	753.20	15	120	10	2:1
	ESI	WAS	1	1	1	0.65	2.13	45	120	10	1.74:1
		CAS	1	-1	1	1	1.20	45	60	10	2:1
HC Cavitation 25°C	**Biogas**	WAS	/	/	1	0.8	670.44	/	/	10	2.74:1
		CAS	/	/	0.8	1	700.62	/	/	8	3:1
	ESI	WAS	/	/	1	1	1.09	/	/	10	3:1
		CAS	/	/	1	0.9	1.12	/	/	10	2.87:1
HC Cavitation 55°C	**Biogas**	WAS	/	/	1	0.8	705.36	/	/	10	2.74:1
		CAS	/	/	0.9	1	798.07	/	/	9.5	3:1
	ESI	WAS	/	/	1	0.7	1.07	/	/	10	2.60:1
		CAS	/	/	1	0.8	1.12	/	/	10	2.74:1
Ultrasound	**Biogas**	WAS	/	/	1	0.8	683.95	/	/	10	2.74:1
		CAS	/	/	1	0.5	703.63	/	/	10	2.34:1
	ESI	WAS	/	/	1	1	0.90	/	/	10	3:1
		CAS	/	/	1	1	0.92	/	/	10	3:1

4.Conclusions

565 This study evaluated the AD of real OFMSW prior pre-treated with four types of pre-treatments:
566 mechanical, thermal, HC, and US, to optimise the whole process. The tested pre-treatments and AD
567 configurations were selected through DoE, considering the interaction between the pre-treatment and
568 the AD phases. The experiments of each DoE were evaluated by measuring the biogas production,
569 the VS removal, and the ESI. The results were used to build regression models correlating the
570 responses to the factors involved in the study, their interactions, and their quadratic effects.
571 The best configurations of pre-treatments and AD were the ones performed with thermal pre-
572 treatment at 120° C for 45 min and inoculum incubation of 10 d at SI equal to 2:1, due to the thermal
573 solubilisation effect, and HC at 55 °C and inoculum incubation of 10 d at SI equal to 3:1, for the
574 combined heat-bubbling effect, which enhanced the availability of the digestible fraction of OFMSW.
575 Pre-treatment time was significant only in the case of thermal pre-treatment and it showed a
576 significant interaction with the inoculum incubation time.
577 The AD of US-OFMSW achieved the lowest performances since inhibition occurred. In the future
578 the combined environmental and economic assessments of the four pre-treatments and AD will be
579 investigated.

580 **Statements and Declarations**

581 **Funding**

582 The authors gratefully acknowledge financial support from Region Piemonte (Italy), POR FESR
583 2014/2020, Project BIOENPRO4TO.

584 **Competing interests**

585 The authors declare no competing interests.

586 **Author's contributions**

587 Authors 'contributions are detailed in the following. F. Demichelis carried out the conceptualization
588 and experimental tests and wrote part of the paper. F.A. Deorsola contributed to realise the
589 methodology and wrote and reviewed the paper, T. Tommasi supported the structure of the study. G.
590 Cravotto optimised the hydrodynamic-cavitation and ultrasound pre-treatment tests and contributed

to write part of the paper; G.Grillo performed hydrodynamic-cavitation and ultrasound pre-treatment tests; E. Robotti coordinated and realised the modelling study and contribute to write and review the manuscript. E. Marengo contributed to modelling analysis and D. Fino realised the conceptualization to supervision of the study.

Data availability

Data are complete reported in the manuscript and in the supplementary material.

Acknowledgements

The authors are grateful to Mr. Daniele Crudo (E-PIC srl - Mongrando (BI), Italy) for his valuable technical assistance.

References

Abraham, A., Mathew, A.K., Park, H., Choi, O., Sindhu, R., 2020. Pretreatment strategies for enhanced biogas production from lignocellulosic biomass. Bioresour. Technol. 301, 122725. https://doi.org/10.1016/j.biortech.2019.122725

Angelidaki, I., Alves, M., Bolzonella, D., Borzacconi, L., Campos, J.L., Guwy, A.J., Kalyuzhnyi, S., Jenicek, P., Lier, J.B. Van, 2009. Defining the biomethane potential (BMP) of solid organic wastes and energy crops : a proposed protocol for batch assays. Water Sci. Technol 927–934. https://doi.org/10.2166/wst.2009.040

Atelge, M. R., Atabani, A.E., Banu, J.R., Krisa, D., Kaya, M., Eskicioglu, C., Kumar, G., Lee, C., Yildiz, Y., Unalan, S., Mohanasundaram, R., Duman, F., 2020. A critical review of pretreatment technologies to enhance anaerobic digestion and energy recovery. Fuel 270, 117494. https://doi.org/10.1016/j.fuel.2020.117494

Bougrier, C., Carrère, H., Delgenès, J.P., 2005. Solubilisation of waste-activated sludge by ultrasonic treatment. Chem. Eng. J. 106, 163–169. https://doi.org/10.1016/j.cej.2004.11.013

Box, G., Hunter, S., H., 2005. Statistics for Experimenters., Second Edi. ed. Wiley., New Jersey, USA

Browne, J.D., Murphy, J.D., 2013. Assessment of the resource associated with biomethane from food waste. Appl. Energy 104, 170–177. https://doi.org/10.1016/j.apenergy.2012.11.017

Bruni, E., Jensen, A.P., Angelidaki, I., 2010. Steam treatment of digested biofibers for increasing biogas production. Bioresour. Technol. 101, 7668–7671. https://doi.org/10.1016/j.biortech.2010.04.064

Calcio, E., Tabasso, S., Grillo, G., Cravotto, G., Dreyer, T., Schories, G., Altenberg, S., Jashina, L., Telysheva, G., 2018. Comptes Rendus Chimie Wheat straw lignin extraction with bio-based solvents using enabling technologies. Comptes rendus - Chim. 21, 563–571. https://doi.org/10.1016/j.crci.2018.01.010

Carmona-Cabello, M., Garcia, I.L., Leiva-Candia, D., Dorado, M.P., 2018. Valorization of food

627 waste based on its composition through the concept of biorefinery. Curr. Opin. Green Sustain.
628 Chem. 14, 67–79. https://doi.org/10.1016/j.cogsc.2018.06.011
629 Carrère, H., Sialve, B., Bernet, N., 2009. Improving pig manure conversion into biogas by thermal
630 and thermo-chemical pretreatments. Bioresour. Technol. 100, 3690–3694.
631 https://doi.org/10.1016/j.biortech.2009.01.015
632 Cesaro, A., Velten, S., Belgiorno, V., Kuchta, K., 2014. Enhanced anaerobic digestion by ultrasonic
633 pretreatment of organic residues for energy production. J. Clean. Prod. 74, 119–124.
634 https://doi.org/10.1016/j.jclepro.2014.03.030
635 Cheah, Y.K., Vidal-Antich, C., Dosta, J., Mata-Álvarez, J., 2019. Volatile fatty acid production
636 from mesophilic acidogenic fermentation of organic fraction of municipal solid waste and food
637 waste under acidic and alkaline pH. Environ. Sci. Pollut. Res. 26, 35509–35522.
638 https://doi.org/10.1007/s11356-019-05394-6
639 Chen, H., Yi, H., Li, H., Guo, X., Xiao, B., 2020. Effects of thermal and thermal-alkaline
640 pretreatments on continuous anaerobic sludge digestion: Performance, energy balance and,
641 enhancement mechanism. Renew. Energy 147, 2409–2416.
642 https://doi.org/10.1016/j.renene.2019.10.051
643 Chen, Y., Cheng, J.J., Creamer, K.S., 2008. Inhibition of anaerobic digestion process : A review.
644 Bioresour. Technol. 99, 4044–4064. https://doi.org/10.1016/j.biortech.2007.01.057
645 Chen, Y., Ping, Q., Li, D., Dai, X., Li, Y., 2022. Comprehensive insights into the impact of
646 pretreatment on anaerobic digestion of waste active sludge from perspectives of organic matter
647 composition, thermodynamics, and multi-omics. Water Res. 226, 119240.
648 https://doi.org/10.1016/j.watres.2022.119240
649 Coarita Fernandez, H., Teixeira Franco, R., Bayard, R., Buffiere, P., 2020. Mechanical Pre-
650 treatments Evaluation of Cattle Manure Before Anaerobic Digestion. Waste Biomass
651 Valorization. https://doi.org/10.1007/s12649-020-01022-4
652 Dasgupta, A., Chandel, M.K., 2019. Enhancement of biogas production from organic fraction of

municipal solid waste using hydrothermal pretreatment. Bioresour. Technol. Reports 7, 100281. https://doi.org/10.1016/j.biteb.2019.100281

Demichelis, F., Fiore, S., Onofrio, M., 2018. Pre-treatments aimed at increasing the biodegradability of cosmetic industrial waste. Process Saf. Environ. Prot. 118. https://doi.org/10.1016/j.psep.2018.07.001

Demichelis, F., Tommasi, T., Deorsola, F.A., Marchisio, D., Fino, D., 2022. Effect of inoculum origin and substrate-inoculum ratio to enhance the anaerobic digestion of organic fraction municipal solid waste (OFMSW). J. Clean. Prod. 351, 131539. https://doi.org/10.1016/j.jclepro.2022.131539

Gagić, T., Perva-Uzunalić, A., Knez, Ž., Škerget, M., 2018. Hydrothermal Degradation of Cellulose at Temperature from 200 to 300 °c. Ind. Eng. Chem. Res. 57, 6576–6584. https://doi.org/10.1021/acs.iecr.8b00332

Gu, Y., Chen, X., Liu, Z., Zhou, X., Zhang, Y., 2020. Effect of inoculum sources on the anaerobic digestion of rice straw. Bioresour. Technol. 158, 149–155. https://doi.org/10.1016/j.biortech.2014.02.011

Hai, T., Dhahad, H.A., Kumar, P., Fahad, S., Ibrahim, A., Fahmi, A., Attia, E., Shamseldin, M.A., Najat, A., 2022. Innovative proposal of energy scheme based on biogas from digester for producing clean and sustainable electricity , cooling and heating : Proposal and multi-criteria optimization. Sustain. Energy Technol. Assessments 53, 102618. https://doi.org/10.1016/j.seta.2022.102618

Harun, N., Othman, N.A., Zaki, N.A., Mat Rasul, N.A., Samah, R.A., Hashim, H., 2019. Simulation of Anaerobic Digestion for Biogas Production from Food Waste Using SuperPro Designer. Mater. Today Proc. 19, 1315–1320. https://doi.org/10.1016/j.matpr.2019.11.143

Hendriks, A.T.W.M., Zeeman, G., 2009. Pretreatments to enhance the digestibility of lignocellulosic biomass. Bioresour. Technol. 100, 10–18. https://doi.org/10.1016/j.biortech.2008.05.027

679 Huang, Y., Ma, Y., Wan, J., Wang, Y., 2021. Modeling the Performance of Full-Scale Anaerobic
680 Biochemical System Treating Deinking Pulp Wastewater Based on Modified Anaerobic
681 Digestion Model No . 1 12, 1–12. https://doi.org/10.3389/fmicb.2021.755398
682 Jain, Siddharth, Jain, Shivani, Wolf, I.T., Lee, J., Tong, Y.W., 2015. A comprehensive review on
683 operating parameters and different pretreatment methodologies for anaerobic digestion of
684 municipal solid waste. Renew. Sustain. Energy Rev. 52, 142–154.
685 https://doi.org/10.1016/j.rser.2015.07.091
686 Kainthola, J., Kalamdhad, A.S., Goud, V. V., 2019. A review on enhanced biogas production from
687 anaerobic digestion of lignocellulosic biomass by different enhancement techniques. Process
688 Biochem. 84, 81–90. https://doi.org/10.1016/j.procbio.2019.05.023
689 Karthikeyan, O.P., Trably, E., Mehariya, S., Bernet, N., Wong, J.W.C., Carrere, H., 2018.
690 Pretreatment of food waste for methane and hydrogen recovery : A review. Bioresour.
691 Technol. 249, 1025–1039. https://doi.org/10.1016/j.biortech.2017.09.105
692 Kawai, M., Nagao, N., Tajima, N., Niwa, C., Matsuyama, T., Toda, T., 2014. The effect of the
693 labile organic fraction in food waste and the substrate / inoculum ratio on anaerobic digestion
694 for a reliable methane yield. Bioresour. Technol. 157, 174–180.
695 https://doi.org/10.1016/j.biortech.2014.01.018
696 Kianmehr, P., Parker, W., Seto, P., 2010. An evaluation of protocols for characterization of ozone
697 impacts on WAS properties and digestibility. Bioresour. Technol. 101, 8565–8572.
698 https://doi.org/10.1016/j.biortech.2010.06.061
699 Kovalovszki, A., Treu, L., Ellegaard, L., Luo, G., Angelidaki, I., 2020. Modeling temperature
700 response in bioenergy production : Novel solution to a common challenge of anaerobic
701 digestion. Appl. Energy 263, 114646. https://doi.org/10.1016/j.apenergy.2020.114646
702 Kumar Biswal, B., Huang, H., Dai, J., Chen, G.H., Wu, D., 2020. Impact of low-thermal
703 pretreatment on physicochemical properties of saline waste activated sludge, hydrolysis of
704 organics and methane yield in anaerobic digestion. Bioresour. Technol. 297.

https://doi.org/10.1016/j.biortech.2019.122423

Lauberte, L., Telysheva, G., Cravotto, G., Andersone, A., Janceva, S., Dizhbite, T., Arshanitsa, A., Jurkjane, V., Vevere, L., Grillo, G., Calcio, E., Tabasso, S., 2021. Lignin e Derived antioxidants as value-added products obtained under cavitation treatments of the wheat straw processing for sugar production. J. Clean. Prod. 303, 126369. https://doi.org/10.1016/j.jclepro.2021.126369

Li, L., He, Q., Zhao, X., Wu, D., Wang, X., Peng, X., 2018. Anaerobic digestion of food waste: Correlation of kinetic parameters with operational conditions and process performance. Biochem. Eng. J. 130, 1–9. https://doi.org/10.1016/j.bej.2017.11.003

Li, Y., Chen, Y., Wu, J., 2019. Enhancement of methane production in anaerobic digestion process: A review. Appl. Energy 240, 120–137. https://doi.org/10.1016/j.apenergy.2019.01.243

Li, Y., Jin, Y., Li, J., Li, H., Yu, Z., Nie, Y., 2017. Effects of thermal pretreatment on degradation kinetics of organics during kitchen waste anaerobic digestion. Energy 118, 377–386. https://doi.org/10.1016/j.energy.2016.12.041

Liu, Y., Fang, J., Tong, X., Huan, C., Ji, G., Zeng, Y., 2019. Change to biogas production in solid-state anaerobic digestion using rice straw as substrates at different temperatures. Bioresour. Technol. 293, 122066. https://doi.org/10.1016/j.biortech.2019.122066

Mancini, E., Raggi, A., 2021. A review of circularity and sustainability in anaerobic digestion processes. J. Environ. Manage. 291, 112695. https://doi.org/10.1016/j.jenvman.2021.112695

Mehr, A.S., Gandiglio, M., Mosayebnezhad, M., Lanzini, A., Mahmoudi, S.M.S., Yari, M., Santarelli, M., 2017. Solar-assisted integrated biogas solid oxide fuel cell (SOFC) installation in wastewater treatment plant : Energy and economic analysis. Appl. Energy 191, 620–638. https://doi.org/10.1016/j.apenergy.2017.01.070

Menardo, S., Cacciatore, V., Balsari, P., 2015. Batch and continuous biogas production arising from feed varying in rice straw volumes following pre-treatment with extrusion. Bioresour. Technol. 180, 154–161. https://doi.org/10.1016/j.biortech.2014.12.104

731 Motte, J.C., Escudié, R., Hamelin, J., Steyer, J.P., Bernet, N., Delgenes, J.P., Dumas, C., 2015.
732 Substrate milling pretreatment as a key parameter for Solid-State Anaerobic Digestion
733 optimization. Bioresour. Technol. 173, 185–192.
734 https://doi.org/10.1016/j.biortech.2014.09.015
735 Naran, E., Toor, U.A., Kim, D.J., 2016. Effect of pretreatment and anaerobic co-digestion of food
736 waste and waste activated sludge on stabilization and methane production. Int. Biodeterior.
737 Biodegrad. 113, 17–21. https://doi.org/10.1016/j.ibiod.2016.04.011
738 Panigrahi, S., Sharma, H.B., Dubey, B.K., 2020. Anaerobic co-digestion of food waste with
739 pretreated yard waste: A comparative study of methane production, kinetic modeling and
740 energy balance. J. Clean. Prod. 243, 118480. https://doi.org/10.1016/j.jclepro.2019.118480
741 Pilli, S., Bhunia, P., Yan, S., LeBlanc, R.J., Tyagi, R.D., Surampalli, R.Y., 2011. Ultrasonic
742 pretreatment of sludge: A review. Ultrason. Sonochem. 18, 1–18.
743 https://doi.org/10.1016/j.ultsonch.2010.02.014
744 Pleissner, D., Peinemann, J.C., 2020. The Challenges of Using Organic Municipal Solid Waste as
745 Source of Secondary Raw Materials. Waste Biomass Valorization 11, 435–446.
746 https://doi.org/10.1007/s12649-018-0497-1
747 Rajput, A.A., Zeshan, Visvanathan, C., 2018. Effect of thermal pretreatment on chemical
748 composition, physical structure and biogas production kinetics of wheat straw. J. Environ.
749 Manage. 221, 45–52. https://doi.org/10.1016/j.jenvman.2018.05.011
750 Rasapoor, M., Ajabshirchi, Y., Adl, M., Abdi, R., Gharibi, A., 2016. The effect of ultrasonic
751 pretreatment on biogas generation yield from organic fraction of municipal solid waste under
752 medium solids concentration circumstance. Energy Convers. Manag. 119, 444–452.
753 https://doi.org/10.1016/j.enconman.2016.04.066
754 Rillo, E., Gandiglio, M., Lanzini, A., Bobba, S., Santarelli, M., Blengini, G., 2020. Life Cycle
755 Assessment (LCA) of biogas-fed Solid Oxide Fuel Cell (SOFC) plant. Energy 126, 585–
756 602. https://doi.org/10.1016/j.energy.2017.03.041

Rittmann, S.K.M.R., Seifert, A.H., Bernacchi, S., 2018. Kinetics, multivariate statistical modelling, and physiology of CO2-based biological methane production. Appl. Energy 216, 751–760. https://doi.org/10.1016/j.apenergy.2018.01.075

Saxena, S., Saharan, V.K., George, S., 2019. Modeling & simulation studies on batch anaerobic digestion of hydrodynamically cavitated tannery waste effluent for higher biogas yield. Ultrason. Sonochem. 58, 104692. https://doi.org/10.1016/j.ultsonch.2019.104692

Scherzinger, M., Kaltschmitt, M., 2021. Thermal pre-treatment options to enhance anaerobic digestibility – A review. Renew. Sustain. Energy Rev. 137, 110627. https://doi.org/10.1016/j.rser.2020.110627

Suksong, W., Mamimin, C., Prasertsan, P., Kongjan, P., 2019. Effect of inoculum types and microbial community on thermophilic and mesophilic solid-state anaerobic digestion of empty fruit bunches for biogas production. Ind. Crop. Prod. 133, 193–202. https://doi.org/10.1016/j.indcrop.2019.03.005

Victorin, M., Davidsson, Å., Wallberg, O., 2020. Characterization of Mechanically Pretreated Wheat Straw for Biogas Production. Bioenergy Res. 13, 833–844. https://doi.org/10.1007/s12155-020-10126-7

Wang, L., Long, F., Liao, W., Liu, H., 2020. Prediction of anaerobic digestion performance and identi fi cation of critical operational parameters using machine learning algorithms. Bioresour. Technol. 298, 122495. https://doi.org/10.1016/j.biortech.2019.122495

Zhang, J, Luo, W., Wang, Y., Li, G., Liu, Y., Gong, X., 2019a. Anaerobic cultivation of waste activated sludge to inoculate solid state anaerobic co-digestion of agricultural wastes : Effects of di ff erent cultivated periods. Bioresour. Technol. 294, 122078. https://doi.org/10.1016/j.biortech.2019.122078

Zhang, J, Mao, L., Nithya, K., Loh, K.C., Dai, Y., He, Y., Wah Tong, Y., 2019b. Optimizing mixing strategy to improve the performance of an anaerobic digestion waste-to-energy system for energy recovery from food waste. Appl. Energy 249, 28–36.

https://doi.org/10.1016/j.apenergy.2019.04.142

Zhang, L., Loh, K.C., Zhang, J., 2019. Enhanced biogas production from anaerobic digestion of solid organic wastes: Current status and prospects. Bioresour. Technol. Reports 5, 280–296. https://doi.org/10.1016/j.biteb.2018.07.005

Zhang, Y., Banks, C.J., 2013. Impact of different particle size distributions on anaerobic digestion of the organic fraction of municipal solid waste. Waste Manag. 33, 297–307. https://doi.org/10.1016/j.wasman.2012.09.024

Zhen, G., Lu, X., Kato, H., Zhao, Y., Li, Y.Y., 2017. Overview of pretreatment strategies for enhancing sewage sludge disintegration and subsequent anaerobic digestion: Current advances, full-scale application and future perspectives. Renew. Sustain. Energy Rev. 69, 559–577. https://doi.org/10.1016/j.rser.2016.11.187