

Improvement of energy recovery from the digestion of waste activated sludge (WAS) through intermediate treatments: The effect of the hydraulic retention time (HRT) of the

*Original*

Improvement of energy recovery from the digestion of waste activated sludge (WAS) through intermediate treatments: The effect of the hydraulic retention time (HRT) of the first-stage digestion / Ruffino, B., Cerutti, A., Campo, G., Scibilia, G., Lorenzi, E., Zanetti, M.. - In: APPLIED ENERGY. - ISSN 0306-2619. - STAMPA. - 240:(2019), pp. 191-204. [10.1016/j.apenergy.2019.02.061]

*Availability:*

This version is available at: 11583/2729654 since: 2019-03-28T17:35:16Z

*Publisher:*

Elsevier Ltd

*Published*

DOI:10.1016/j.apenergy.2019.02.061

*Terms of use:*

This article is made available under terms and conditions as specified in the corresponding bibliographic description in the repository

*Publisher copyright*

Elsevier preprint/submitted version

Preprint (submitted version) of an article published in APPLIED ENERGY © 2019,  
<http://doi.org/10.1016/j.apenergy.2019.02.061>

(Article begins on next page)

1  
2  
3  
4 **Improvement of the energy recovery from the digestion of waste activated sludge (WAS) through**  
5  
6 **intermediate treatments: the effect of the hydraulic retention time (HRT) of the first-stage**  
7  
8 **digestion**  
9

10  
11  
12  
13  
14 Barbara Ruffino<sup>1</sup>§ (\*), Alberto Cerutti<sup>1</sup>§, Giuseppe Campo<sup>1</sup>, Gerardo Scibilia<sup>2</sup>, Eugenio Lorenzi<sup>2</sup>,  
15  
16 Mariachiara Zanetti<sup>1</sup>  
17  
18

19  
20  
21 <sup>1</sup>Department of Environment, Land and Infrastructure Engineering, Politecnico di Torino, corso Duca  
22  
23 degli Abruzzi, 24 – 10129 Torino, Italy  
24  
25

26 <sup>2</sup>Research Center, Società Metropolitana Acque Torino S.p.A., Viale Maestri del Lavoro, 4 – 10127  
27  
28 Torino, Italy  
29  
30

31  
32  
33 (\*) Corresponding author

34  
35 Barbara RUFFINO

36  
37 DIATI, Department of Environment, Land and Infrastructure Engineering

38  
39 Politecnico di Torino

40  
41 Corso Duca degli Abruzzi, 24

42  
43 10129 Torino, ITALY

44  
45 Ph. +39.011.0907663

46  
47 Fax +39.011.0907699

48  
49 e-mail: [barbara.ruffino@polito.it](mailto:barbara.ruffino@polito.it)  
50

51  
52 § Authors contributed equally to the study  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 **Abstract**  
5

6 Reduced biodegradability of waste activated sludge (WAS) limits the production of methane and the  
7 consequent energy recovery in an anaerobic digestion (AD) process. Pre-treatments are a solution to  
8 increase the biodegradability of bacteria cell biomass, but a large part of poorly degradable organic  
9 matter is left after digestion. The utilization of intermediate hydrolysis treatments (IHTs) may help in  
10 converting even the most recalcitrant parts of organic matter in methane.  
11  
12  
13  
14  
15  
16  
17

18 This study employed a three-phase experimentation to assess the effect of the hydraulic retention time  
19 (HRT) of the digestion first stage, on the overall performance of a two-stage digestion process, with an  
20 in-between treatment, carried out on WAS. The three phases of the experimentation included a first-  
21 stage digestion (with HRTs = 5, 10 and 15 days), performed in a semi-continuous 10L-reactor,  
22 followed by a thermal (90°C) or a hybrid (thermal 90°C + chemical, 4% NaOH) IHT, completed by a  
23 second-stage digestion carried out in a batch mode. Both the digestion processes were performed in  
24 mesophilic conditions (38°C).  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34

35 The obtained results revealed that, in the presence of a thermal IHT and by fixing the duration of the  
36 second stage to 20 days, the overall specific methane production (SMP) tended to a constant value, in  
37 the order of 0.205 Nm<sup>3</sup>/kg VS added, irrespective of the duration of the first stage. Conversely, when a  
38 hybrid treatment was applied, the difference between a short (5 days) and a medium (10-15 days)  
39 duration of the digestion first stage became evident, with SMPs in the order of 0.247 and 0.230 Nm<sup>3</sup>/kg  
40 VS added, respectively.  
41  
42  
43  
44  
45  
46  
47  
48  
49

50 Energy and economic sustainability of the application of IHTs at a full scale plant required an adequate  
51 thickening of sludge / digestate matrices and an efficient heat exchange between donor (sludge after  
52 treatment) and acceptor (cold sludge before digestion) agents. It was demonstrated that for separated or  
53 joined digestion processes of primary sludge (7.0% TS) and treated digestates, with heat recovery and  
54 different combinations of the duration of the first and second stage of AD, TS contents in the order of  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

4%, 6% and 8-9% were required to make the thermal balance neutral for thermal exchanges efficiencies of 100%, 70% and 50%, respectively.

Keywords: intermediate hydrolysis treatments; two-stage digestion; energy balance; economic sustainability; digestate

## 1. Introduction

Anaerobic digestion (AD) is employed worldwide as the oldest and most important process for stabilization of both primary sludge (PS) and waste activated sludge (WAS) generated in wastewater treatment plants (WWTPs). However, it is well known that the aim of the AD process carried out in the sludge line of a WWTP is not only reducing the volume of sludge and making it stable, but primarily producing an amount of biogas sufficient to self-sustain the AD process and possibly supplying the WWTP and the close users with energy in the form of hot water and electricity. In this context, the reduced biodegradability of WAS represents a drawback because it limits the methane production and volatile solid (VS) reduction. In fact, most of the biodegradable material of WAS is either enclosed in the microbial cell wall or enmeshed in an extracellular polymeric matrix (Grübel and Suschka, 2015).

Various pre-treatment processes (e.g. thermal, mechanical, chemical and biological) have been developed with the aim of increasing the biodegradability of bacteria cell biomass and reducing volume and retention time of digesters. Extensive reviews of the effectiveness of each of the afore-mentioned types of process in terms of resulting properties of treated sludge, digester performance, energy balances, environmental benefits and current state of real-world application are reported among others in Carrère et al., 2010; Cano et al., 2015; Zhen et al., 2017; Waclawek et al., 2019. Pre-treatments, as their name suggests, are performed prior to the process of AD. In this way, the different fractions of organic matter, that is both complex compounds and readily bio-degradable substances (Jimenez et al., 2015), undergo the pre-treatment (Wang and Li, 2016). Therefore, the material that is already available to be converted into methane enters the pre-treatment despite it does not need to and does not benefit from such a treatment (Svensson et al., 2018). After pre-treatment, the treated substrate enters the anaerobic digesters where degradable organic matter is converted into methane, still leaving a large fraction of poorly degradable organic matter, that comes out from the digester and does not fully harness its energy potential (Aragón-Briceño et al., 2017). Some authors hypothesized that, in a

1  
2  
3  
4 digestion process involving WAS, this poorly degradable fraction was associated with the extracellular  
5  
6 polymeric substance (EPS), that is released from WAS during digestion because of microbial  
7  
8 metabolism, self-protective reaction and cell lysis (Shana et al., 2013; Williams et al., 2015).  
9  
10 Consequently, the use of sludge pre-treatment technologies may only help to hydrolyze the EPS in the  
11  
12 sludge feedstock but it is not of any benefit for the fraction produced during an AD process.  
13

14  
15  
16 A possible solution to deal with the EPS generated during sludge digestion and other poorly degradable  
17  
18 organic matter fractions is to make use of treatments that follow AD (named post-treatments, inter-  
19  
20 stage treatments, intermediate hydrolysis treatments, IHTs). Such treatments only concentrate on the  
21  
22 slowly degradable parts of the sludge, in contrast to pre-treatment methods (Shana et al., 2015). In this  
23  
24 new configuration, the AD process converts the readily biodegradable organic matter into methane, in  
25  
26 the first step, and the residual digestate containing the recalcitrant or more complex material undergoes  
27  
28 the IHT in a second phase. The “inter-treated” digestate enters the anaerobic digester either by  
29  
30 recirculating a certain fraction or, as a whole, entering a second-stage AD process. This will reduce the  
31  
32 residual organic matter and increase the energy benefit of the treatment (Ortega-Martinez et al., 2016).  
33  
34  
35

36  
37  
38 Despite the possible advantages, IHTs of WAS have until now received little attention in comparison to  
39  
40 pre-treatments. Moreover, due to the nature of the organic matter that remains after an AD process,  
41  
42 most part of the recent applications of IHTs have used high energy-demanding processes like  
43  
44 hydrothermal treatments (HTTs). The examples reported in the follow point out that some authors have  
45  
46 focused only on post-treatments, searching for the best operating conditions to extract the highest  
47  
48 amount of biogas from the second stage of digestion. Conversely, a limited number of authors have  
49  
50 compared the performances of HTTs used as a pre- or post-treatment.  
51  
52

53  
54  
55 With reference to the first group of studies, Aragón-Briceño and coauthors (2016) evaluated the effect  
56  
57 of the temperature parameter (160, 220 and 250 °C) in HTTs carried out on digestate samples collected  
58  
59 from Leeds (UK) WWTP. The performance of the IHTs was studied with respect to product yields  
60  
61  
62

1  
2  
3  
4 (that is the production of hydrochar), biomethane potential and solubilisation of organic carbon. They  
5  
6 obtained the highest hydrochar yields at 220°C. The solubilisation of carbon increased from 4.62% in  
7  
8 the raw feedstock to 31.68%, 32.56% and 30.48% after HTTs at 160, 220 and 250 °C, respectively.  
9  
10 Finally, the HTT enhanced the potential methane production up to 58% for both, the whole fraction  
11  
12 (hydrochar + processed water) and processed waters. Bjerg-Nielsen and coauthors (2018) used an  
13  
14 anaerobically digested WAS (HRT = 15 days, T = 55°C) from Aalborg Vest WWTP (Denmark), with  
15  
16 and without wheat straw as a co-substrate, to study the effect of time (30 and 60 min) and temperature  
17  
18 (120–190 °C) on the performances of thermal IHTs. They observed that, compared to non-treatment,  
19  
20 IHTs increased the secondary step methane yield from 52 to 222 L CH<sub>4</sub>/kg VS and from 147 to 224 L  
21  
22 CH<sub>4</sub>/kg VS for pre-digested WAS and pre-co-digested WAS respectively, at an optimum of 170 °C and  
23  
24 30 min.  
25  
26  
27  
28  
29

30  
31 Conversely, Svensson and coauthors (2018) compared the performances of pre- and post-treatments  
32  
33 that were carried out in a non-homogeneous way: low temperature (70°C) and HTT (134-175°C),  
34  
35 respectively. Specifically, they used a Cambi steam explosion unit to treat digestate samples from  
36  
37 source separated food waste (SSFW) digested in thermophilic conditions and from a mixture of sewage  
38  
39 sludge treated in mesophilic conditions. The thermal hydrolysis was found to be more efficient, on  
40  
41 WAS digestate than on SSFW digestate, in relation to methane yields and dewatering capacity. For  
42  
43 temperatures above 165°C, the authors observed increases in volumetric methane yield and COD  
44  
45 reduction of 7% and 70% respectively, compared to a conventional pre-treatment carried out at 70°C.  
46  
47  
48  
49

50  
51 Ortega-Martinez and coauthors (2016) demonstrated that intermediate HTTs could improve the  
52  
53 methane production of a mixed sludge by 45% and 20% compared to a conventional AD process and  
54  
55 pre-treatments followed by AD, respectively. They carried out pre- and inter-treatments in a laboratory-  
56  
57 scale thermal steam explosion system at temperatures from 110 to 200 °C and for contact times of 10,  
58  
59 30 and 50 min. Recently, Yuan and coauthors (2019) compared the efficiency of HTTs (130-210 °C) as  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 pre- and post-treatment on PS and anaerobically digested sludge. They demonstrated that, although  
5  
6 post-treatments achieved higher total methane yield and solids reduction, the increment of methane  
7  
8 yield was found to be similar through both strategies (pre- and post-treatments) compared to their  
9  
10 control counterparts (no HTT). The decomposition of insoluble organic carbon was also similar via  
11  
12 both strategies.  
13  
14

15  
16 Finally, Rennuit and coauthors (2018) tested a thermophilic (55°C) aerobic digestion (TAD) as a pre-  
17  
18 and inter-stage treatment of sludge and digestates samples coming from the Ejby Mølle WWTP  
19  
20 (Denmark, treating capacity 385,000 population equivalent, p.e.). Digestates were originated from a  
21  
22 mixture of primary (60%), dewatered secondary sludge (40%) and highly degradable organic waste  
23  
24 (depending on availability). The authors demonstrated that TAD as a pre-treatment decreased the  
25  
26 methane yield (up to -70%), due to oxidation losses, whereas, when used as an inter-stage treatment, it  
27  
28 slightly improved the overall methane yield (+2.6%) and the total COD removal (+5%) compared to  
29  
30 control.  
31  
32  
33  
34

35  
36 From the review of the existing studies, it can be seen that most part of the processes employed for  
37  
38 IHTs are high energy-demanding processes, like HTT Cambi, carried out at 140-180 °C. Moreover, all  
39  
40 the mentioned experiences used digested substrates from full scale plants (with a fixed HRT) or from  
41  
42 batch tests, in the case a comparison between pre- and post- treatments was carried out. In the specific  
43  
44 context of the application of low-temperature IHTs (that is treatments that employ only a limited  
45  
46 amount of energy compared to HTTs), the HRT of the AD process that generates the intermediate  
47  
48 digestate is a key issue. In fact, the HRT of the digestion first stage affects the amount of biogas  
49  
50 produced in the first stage, the state/characteristics of the organic matter after the partial digestion and  
51  
52 the volume of substrate to be fed to the second stage. All these elements concur in the biogas  
53  
54 production and consequent energy recovery from the whole (two-stage) digestion process.  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 In a previous study (Campo et al., 2018), the effect of low-temperature thermo-alkali IHTs on two  
5  
6 digestate samples with different HRTs (7 and 15 days) was compared. However, the two substrates  
7  
8 used for the study had different origin, the Castiglione Torinese full scale WWTP (Italy, 2,000,000  
9  
10 p.e.) and a pilot scale (300-L) digester fed with the WAS collected from the same WWTP, respectively.  
11  
12

13  
14 In the afore-mentioned study, the performances of the IHTs were compared with those of a process that  
15  
16 included a first stage of pre-treatment (same operating conditions as the IHT) and a subsequent stage of  
17  
18 batch digestion, both carried out on a WAS sample.  
19  
20

21 The aim of this new study was analyzing the effect of the duration of the first stage of an AD process  
22  
23 on the overall energy recovery, firstly in the form of biogas and consequently in the form of electric  
24  
25 energy and heat, from a system made of a two-stage AD with an in-between digestate treatment. To  
26  
27 achieve the prefixed aim, controlled-HRT (5, 10 and 15 days) digestate samples were produced in a  
28  
29 semi-continuous pilot scale apparatus (10 L), subsequently, they were treated (90°C, 4% NaOH, 90  
30  
31 minutes) and, finally, the residual methane production was quantified in a batch AD apparatus. To the  
32  
33 best of our knowledge, this is the first study that makes use of a first-stage controlled-HRT digestate to  
34  
35 evaluate the energy and economic sustainability of an AD process made of two stages. The results  
36  
37 obtained in these tests will subsequently be used to evaluate the future performances of the sludge line  
38  
39 of Castiglione Torinese plant, the largest Italian WWTP, once intermediate treatments are  
40  
41 implemented. The increase in the energy production, in the form of electricity that can be sold to  
42  
43 external users, compared to the present situation, of which the WWTP can benefit from the  
44  
45 implementation of the novel IHTs is the key point that will drive the WWTP managers in the decision  
46  
47 to go for the renovation of the sludge line.  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

## 2. Materials and Methods

### 2.1 Reactors set up and operation

The scheme of the system used for the study is shown in Figure 1. The study was articulated in three phases, each of which was repeated three times, one for each of the tested HRTs (5-10-15 days). The first phase was a semi-continuous digestion process, followed by an in-between treatment phase of the digestate and completed by a second-stage digestion process carried out in a batch mode. The experimental system consisted of a first semi-continuous anaerobic digester with a useful volume of 10 L, 500-ml jars for IHTs and a series of six 6 L batch anaerobic digesters (4.5 L working volume) for the second-stage AD process.

#### 2.1.1 Phase 1 – Semi-continuous tests on untreated WAS samples

In the first phase, a semi-continuous AD process was run on WAS samples collected from Castiglione Torinese SMAT WWTP (see Section 2.2). The test was carried out in mesophilic conditions (38°C). Once a day, aliquots of digestate were extracted from the digester and, subsequently, samples of fresh WAS were used to feed the reactor, for a number of days equal to at least three times the fixed HRT (5-10-15 days). This procedure was followed in order to reach a steady-state condition into the digester and obtain a digestate product representative of the fixed HRT. The OLR for each of the three HRTs, 5, 10 and 15 days, was of 6, 3 and 2 g TS/l·d, respectively.

At the end of each run, the digester was emptied and the digestate was used for the subsequent operations of IHTs and the second phase of the AD process in the batch apparatus (see Sections 2.1.2 and 2.1.3). Inoculum and fresh WAS (see Section 2.2) were subsequently added to the digester to start a new digestion run with a different HRT.

The temperature in the 10 L semi-continuous reactor was kept at 38°C (mesophilic conditions) by a temperature controlled water jacket. The biogas produced in each of the three digestion runs (HRT = 5-

1  
2  
3  
4 10-15 days) was collected in a gas bag. The volume of the gas was quantified and characterized daily  
5  
6 as described in Ruffino et al. (2016). Briefly, the volumetric composition of the biogas in terms of CH<sub>4</sub>,  
7  
8 CO<sub>2</sub>, O<sub>2</sub> was obtained by flushing 500 mL of biogas through a biogas analyzer (Biogas Check,  
9  
10 Geotechnical Instruments Ltd). The residual volume of the biogas after characterization was measured  
11  
12 by displacing volumes of water with the residual gas and referring the obtained value to the normal  
13  
14 conditions (273.15 K and 101.325 kPa). TS and VS concentrations of both the fed sludge (WAS) and  
15  
16 digested product from the 10 L semi-continuous reactor were measured every working day in order to  
17  
18 determine the extent of TS and VS destruction.  
19  
20  
21  
22  
23  
24  
25

### 26 **2.1.2 Phase 2 – Intermediate hydrolysis treatments (IHTs)**

27  
28 After three HRTs, the AD process was stopped, the semi-continuous anaerobic digester was emptied  
29  
30 and the digestate was subjected to the second phase of the study. The controlled-HRT digestate was  
31  
32 divided into three aliquots of the same volume. One aliquot was subjected to a low-temperature thermal  
33  
34 treatment (90 min, 90°C), another to a hybrid (thermo-alkali) treatment (90 min, 90°C, NaOH 4% TS)  
35  
36 and the third was used as a reference (untreated) sample. The conditions for the treatments and the  
37  
38 protocol followed for the tests came from the results of a previous study that involved samples of  
39  
40 digestate generated from the same WAS (Campo et al., 2018). For each of the tested conditions, three  
41  
42 replicates were carried out.  
43  
44  
45  
46  
47

48 The efficacy of the IHTs (low-temperature thermal and thermo-alkali) was preliminarily assessed by  
49  
50 using the disintegration rate (DR) parameter, as in Equation (1)  
51  
52  
53  
54

$$55 \quad DR = \frac{sCOD_t - sCOD_0}{tCOD_0 - sCOD_0} \quad (1)$$

56  
57  
58  
59  
60  
61 Where:

1  
2  
3  
4 sCOD<sub>1</sub> = soluble COD after the lysis process;  
5

6 sCOD<sub>0</sub> = soluble COD of the untreated sample;  
7

8  
9 tCOD<sub>0</sub> = total COD.  
10

11 The DR parameter has been largely employed in the literature to evaluate the efficacy of lysis processes  
12 (Dohányos et al. 1997). It relates the soluble COD released by the lysis treatment to the COD of the  
13  
14 particulate fraction (tCOD–sCOD), that is, the fraction that can be potentially hydrolyzed during the  
15  
16 treatment.  
17  
18  
19  
20  
21  
22

### 23 **2.1.3 Phase 3 – Batch tests on the digestate**

24  
25 After the IHT, each aliquot of treated digestate was divided into two sub-aliquots and subjected to the  
26  
27 third phase of the study, that is a second-stage AD in a batch mode. The AD process was carried out in  
28  
29 the 6 L batch reactors already used in previous experimentations (see Campo et al., 2017). Treated and  
30  
31 untreated digestate was mixed with the inoculum described in Section 2.2., according to a ratio in the  
32  
33 order of 1:1 on a VS basis (specifically, digestate : inoculum ratio = 0.99; 1.11 and 1.09 for HRT of 5;  
34  
35 10 and 15, respectively). No buffer agents and no nutrients were added to the digesters.  
36  
37  
38  
39

40 The biogas produced from each of the batch digester was collected in a gas bag and quantified and  
41  
42 characterized as described in Section 2.1.1. Blank tests were carried out in batch reactors with a  
43  
44 working volume of 2 liters, in order to evaluate the amount of biogas generated by the inoculum in  
45  
46 each experimental run (with 5-, 10- or 15-d digestate). In blank tests the volume of inoculum added to  
47  
48 the reactors was the same of that used for the tests with digestates.  
49  
50  
51  
52

## 53 **2.2 Substrate and inoculum**

54  
55 The sludge used in this study was collected from the secondary settlers of the SMAT WWTP located in  
56  
57 Castiglione Torinese (20 km from Turin, NW Italy). Details of the WWTP were provided in a previous  
58  
59  
60  
61

1  
2  
3  
4 paper (Ruffino et al., 2014). The WWTP has a standard configuration: preliminary treatments (grating  
5 and sand/oil removal), primary settling, pre-denitrification, biological oxidation with a SRT in the order  
6 of 25 days, secondary settling and final filtration on a gravel and anthracite bed.  
7  
8

9  
10 The WWTP has a treatment capacity of approximately 2,000,000 p.e. and generates 4320 kg TS/h of  
11 sludge, 64% of this amount is made of primary sludge. The mass flow rate of WAS is in the order of  
12 1555 kg TS/h, with an average TS content of 0.8%. Pre-thickeners located in the sludge line of the  
13 WWTP increase the solid content (TS) of the WAS from 0.8% to approximately 2-3%.  
14  
15

16  
17 Ten liters of WAS were collected weekly, for the whole duration of the study, and were used to feed  
18 the semi-continuous 10 L digester. The testing conducted in this study spanned approximately nine  
19 months (270 days) of the WWTP operation and hence it was anticipated that there were seasonal  
20 variations in the raw sludge properties. As shown in Figure 2, the ratio between volatile and total solids  
21 (% VS/TS), from the beginning to the end of the study, was appreciably variable. VS/TS ratio was in  
22 the order of 70-71% during the first digestion run (HRT = 5 days), it decreased to 66% during the  
23 second digestion run (HRT = 15 days) and, finally, it rose to 68% during the third digestion run (HRT  
24 = 10 days).  
25  
26

27  
28 Inocula employed for the semi-continuous and batch tests were obtained from a digester of the WWTP  
29 that was fed with sole primary sludge. The average TS content of the inoculum aliquots used for the  
30 experimental runs was in the order of 2%, the VS/TS ratio was of 58%. The inoculum aliquots were  
31 considered ready for the experimentation when their daily biogas production was of less than 1% of the  
32 overall production recorded during the whole period of permanence into the digesters used for the  
33 experimentation (VDI 4630, 2006).  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57

### 58 **2.3 Analytical methods**

59  
60  
61  
62  
63  
64  
65

All the analytical parameters monitored in the tests (TS, VS, pH, electric conductivity (EC), soluble COD (sCOD) were determined according to Standard Methods (APHA, AWWA, WEF, 2012). Soluble COD is the fraction of COD separated after a centrifugation at 4000 rpm for 15 min and a subsequent filtration on a 0.45 mm nylon membrane, as recommended by Roeleveld and van Loosdrecht (2002).

The FOS/TAC parameter was obtained by a potentiometric titration, according to the Nordmann method (Nordmann, 1977), by using a SI Analytics automatic titrator. FOS/TAC parameter is the ratio between volatile organic acids and alkaline buffer capacity; FOS/TAC is an easy-to-do and reliable measure of the risk of acidification of a biogas plant.

### 3. Results and Discussion

#### 3.1 Phase 1 – Semi-continuous tests on untreated WAS samples

Figure 3 presents the evolution of the specific methane production (SMP) from the semi-continuous first-stage digester in the three digestion runs (HRTs = 5, 10 and 15 days). It could be observed that in all cases a steady production of methane was obtained after three HRTs. Values of SMP and VS reduction from the three digestion runs, and pH and FOS/TAC values of the digestate samples with HRT of 5, 10 and 15 days are reported in Table 1.

Table 1. Values of SMP and VS reduction from the three digestion runs, and pH and FOS/TAC values of the digestate samples

HRT (d)	SMP Nm <sup>3</sup> /kg VS added	VS reduction (%)	pH	FOS/TAC
5	0.069±0.001	13	7.29±0.06	0.12±0.02
10	0.095±0.001	22	7.35±0.13	0.11±0.01
15	0.109±0.001	26	7.35±0.09	0.11±0.01

1  
2  
3  
4 The standard deviation of the specific production of methane from the average value was calculated  
5  
6 over the period in which a steady production was observed (see the circles in Figure 3). From these  
7  
8 results it can be concluded that the HRT of the first stage of the AD process played a fundamental role  
9  
10 in the methane production, especially for a low biodegradable substrate such a sludge from biological  
11  
12 origin.  
13  
14

15  
16 Both pH and FOS/TAC values indicated a good stability of the digestion process, even for short HRTs  
17  
18 and without the addition of buffering agents.  
19  
20

21 Sludge samples collected from the same WWTP and digested in a 300 L semi-continuous apparatus,  
22  
23 working with a HRT of 20 days, generated a SMP in the order of  $0.120 \text{ Nm}^3/\text{kg VS added}$  (Campo et  
24  
25 al., 2018). During the test carried out at a pilot scale (300 L) the average VS reduction was in the order  
26  
27 of 19%. In a similar experience carried out in a semi-continuous reactor ( $\text{OLR} = 2.1 \text{ gVS/l}\cdot\text{d}$ ,  $\text{HRT} = 15$   
28  
29 days), Wei and coauthors (2018) observed a SMP of  $0.132 \pm 0.005 \text{ m}^3/\text{kg VS added}$  and a VS reduction  
30  
31 of  $29.2 \pm 0.9\%$  for an untreated WAS. Leite and coauthors (2017) measured a SMP of  $0.067 \text{ Nm}^3/\text{kg}$   
32  
33 VS added in a semi-continuous mesophilic reactor ( $\text{HRT} = 7 \text{ days}$ ,  $\text{ORL} = 1.9 \text{ kg VS/m}^3\cdot\text{d}$ ) fed with a  
34  
35 WAS with a VS/TS ratio in the order of 70%. In that experience  $\text{NaHCO}_3$  was added to the fed sludge  
36  
37 in order to prevent bacteria inhibition or digester souring in the subsequent phase of thermophilic  
38  
39 digestion. In a test where sonication was used to pre-treat WAS before a semi-continuous AD process,  
40  
41 Córdova Lizama and coauthors (2018) obtained daily biogas productions for the raw WAS of  $0.040 \text{ L/d}$   
42  
43 and  $0.033 \text{ L/d}$  for OLRs of 1 and  $3 \text{ kg VS/m}^3\cdot\text{d}$  respectively, that correspond to approximately  $0.050 -$   
44  
45  $0.014 \text{ m}^3/\text{kg VS fed}$  in a  $0.8 \text{ L}$  working volume reactor. The production recorded for a HRT of 10 days  
46  
47 ( $0.037 \text{ L/d}$  or  $0.023 \text{ m}^3/\text{kg VS fed}$ ) was approximately five times lower than that of this study.  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

Table 2. Characteristics of the digestate samples used for the IHTs and the second-stage digestion process and results of the intermediate treatments

HRT (d)	TS (%)	VS/TS (%)	sCOD (mg/l)	thermal IHT DR (%)	hybrid IHT DR (%)
5	3.37±0.20	66.74±0.36	1005±9	21.5±1.1	32.1±1.3
10	2.65±0.07	63.72±0.12	453±6	14.5±0.8	29.2±1.5
15	2.65±0.05	61.29±0.46	823±7	13.3±0.6	19.4±1.0

### 3.2 Phase 2 – Intermediate hydrolysis treatments (IHTs)

The amount of TS, VS, sCOD in the three samples of digestate used for IHTs and for the second-stage digestion process are shown in Table 2. The effect of thermal and hybrid treatments on the samples of digestate collected from the 10-L semi-continuous digester was preliminarily evaluated by using the DR parameter. DR quantifies the sCOD released after the treatment with respect to the maximum amount of potentially hydrolysable COD (that is the particulate COD). Table 2 resumes the effect of IHTs in terms of DR values. From these results, it could be seen that, on average, the only thermal treatment was less efficient in COD solubilizing than the hybrid treatment. The efficiency of both thermal and hybrid treatments decreased with increasing HRTs, consequently, the degree of solubilization seemed to be dependent on the feedstock (Svensson et al., 2018). IHTs had probably a weaker effect on more stable structures, like those that are formed during an AD process (Rennuit et al., 2018). During an IHT, heat and/or alkali changed the floc structure of the digestate, caused the rupture of cell walls and/or extracellular polymeric substances (EPS) with the release of intracellular and/or extracellular dissolved organic compounds to the aqueous phase. These organic compounds are carbohydrates, proteins, lipids and VFAs, that have been proposed to be responsible for the significant increases in sCOD after treatment (Bjerg-Nielsen et al., 2018).

DRs found in this work were similar to the values that were reported for sewage sludge (Ruffino et al., 2016) and for digestates in previous studies (Campo et al., 2018; Ortega Martinez et al., 2016). Ortega

1  
2  
3  
4 Martinez and coauthors (2016) reported solubilization degrees in the order of 25-40% for digestate  
5  
6 samples treated in a laboratory-scale thermal steam explosion system. However, a limited number of  
7  
8 information on the variation of the solubilization degree of digestate sludge subjected to IHTs is  
9  
10 available in the published literature.  
11  
12  
13  
14  
15

### 16 **3.3 Phase 3 – Batch tests on the digestate**

17  
18 In the third phase of the experimentation, the residual methane production of the digestates generated in  
19  
20 Phase 1, subsequently subjected to thermal or hybrid IHTs, was determined in a batch apparatus.  
21  
22 According to VDI 4630 rule (2006), the batch AD test was considered concluded when the daily  
23  
24 volume of the produced biogas was 1% or less of the volume produced during the whole test. All the  
25  
26 batch tests had a duration of approximately 25 days.  
27  
28  
29  
30

31 Figure 4 compares the overall SMP produced by the three systems (first stage HRT = 5, 10, 15 days).  
32  
33 Each bar combines the SMP obtained in the first-stage semi-continuous digestion (red bar), the SMP  
34  
35 obtained in the second-stage batch digestion from the untreated digestate and the SMP increments  
36  
37 (incr) due to the only thermal or hybrid IHT carried out on the digestate samples. It is necessary to  
38  
39 point out that the SMP obtained from the second-stage digestion was referred to the overall production  
40  
41 (25 days) of the batch tests, until exhaustion of the methane generation of the substrate (VDI 4630 rule,  
42  
43 2006).  
44  
45  
46  
47

48 It can be seen that the SMP from the second-stage batch digestion of untreated 5-, 10- and 15-day  
49  
50 digestates was of  $0.096 \pm 0.003$ ,  $0.072 \pm 0.002$  and  $0.053 \pm 0.003$  Nm<sup>3</sup> CH<sub>4</sub>/kg VS added, respectively. Not  
51  
52 surprisingly, digestates with a lower HRT in the first stage had a methane production higher than that  
53  
54 of the digestates that underwent longer digestion periods. In fact, as said in Section 3.1, the first-stage  
55  
56 digestion process consumed only 13%, 22% and 26% of the VS preliminarily available in the sludge  
57  
58 samples subjected to 5-, 10- and 15-day HRT digestion, respectively. The SMP values generated from  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 digestates in batch AD tests, after thermal IHTs, were of  $0.135\pm 0.006$ ,  $0.112\pm 0.004$  and  $0.100\pm 0.004$   
5  
6  $\text{Nm}^3 \text{CH}_4/\text{kg VS}$  added. Consequently, it could be stated that the only thermal treatment determined an  
7  
8 increase in the second-stage SMP of 41%, 56% and 89%, with respect to the untreated digestate, for 5-,  
9  
10 10- and 15-day digestates, respectively. Recently, Bjerg-Nielsen and coauthors (2018) observed that  
11  
12 thermal IHTs carried out at  $120^\circ\text{C}$  for 60 min on a predigested sludge (HRT = 15 days, thermophilic  
13  
14 conditions) determined a triplication in the SMP, that passed from  $0.052$  to  $0.147 \text{ m}^3/\text{kg VS}$  added.  
15  
16

17  
18 Conversely, the SMP values that resulted from the digestates treated with the hybrid IHT were of  
19  
20  $0.178\pm 0.008$ ,  $0.136\pm 0.004$  and  $0.120\pm 0.007 \text{ Nm}^3 \text{ CH}_4/\text{kg VS}$  added. The hybrid treatment determined  
21  
22 an increase in the SMP of 85%, 89% and 126%, in comparison with the untreated digestate, for 5-, 10-  
23  
24 and 15-day digestates, respectively. Hybrid IHTs proved to be more efficient in releasing  
25  
26 biodegradable substances and subsequent methane generation in the second-stage AD process than the  
27  
28 only thermal treatment. These observations were in agreement with the results obtained in a previous  
29  
30 study (Campo et al., 2018) carried out on WAS and digestates samples collected from the same WWTP  
31  
32 and tested in AD processes after low-temperature thermal and hybrid pre- and intermediate treatments.  
33  
34 Compared with IHTs carried out in other studies, the increase found in the present study was high.  
35  
36 Takashima and Tanaka (2014) observed increases in methane production in the order of 14-21% after  
37  
38 acidic thermal post-treatments carried out for 1 h at  $170^\circ\text{C}$  and pH 5-6. Rennuit and coauthors (2018)  
39  
40 in a recent study found very small increases in methane production (in the order of 2-3%) after aerobic  
41  
42 biological inter-stage treatments.  
43  
44  
45  
46  
47  
48

49  
50 The comparison among the three systems (first-stage HRT = 5, 10, 15 days) showed that the production  
51  
52 developed in the second-stage AD, irrespective of the presence of an IHT, tended to bring the overall  
53  
54 production (first plus second stage) to an approximately constant value. In fact, the sum of the  
55  
56 production of the two stages, without an IHT, was of  $0.165$ ,  $0.167$  and  $0.162 \text{ Nm}^3 \text{ CH}_4/\text{kg VS}$  added for  
57  
58 the systems characterized by a first-stage HRT equal to 5, 10 and 15 days, respectively. In the case of a  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 thermal IHT, the overall SMP values were of 0.204, 0.207 and 0.209 Nm<sup>3</sup> CH<sub>4</sub>/kg VS added for the  
5  
6 systems characterized by a first-stage HRT equal to 5, 10 and 15 days, respectively. Finally, where a  
7  
8 hybrid IHT was applied to the digestate samples, the overall SMP values were of 0.247, 0.231 and  
9  
10 0.229 Nm<sup>3</sup> CH<sub>4</sub>/kg VS added, respectively. Only in this last case the difference between a short and a  
11  
12 medium duration of the first-stage AD on the overall SMP became evident. As it can be seen in Figure  
13  
14 5, the combination of the two digestion processes with an in-between hybrid treatment, determined an  
15  
16 overall VS removal of 45%, 44% and 44% for the systems with a first-stage HRT equal to 5, 10 and 15  
17  
18 days, respectively. Digestates extracted and characterized after the second-stage digestion process had  
19  
20 an average VS/TS ratio in the order of 55% (data not shown). The systems that included the thermal  
21  
22 treatments experienced an overall VS removal of 37%, 40% and 41% with a final digestate VS/TS ratio  
23  
24 of 58%, 56% and 54% for first-stage HRTs equal to 5, 10 and 15 days, respectively (data not shown).

25  
26 These results demonstrated that a shorter first phase of the AD process (i.e. in the order of five days)  
27  
28 was more profitable in terms of the overall SMP than a longer one. This result became more evident if  
29  
30 the overall SMP was obtained by summing the SMP generated in the first-stage process (with a  
31  
32 duration of 5, 10 or 15 days), with the SMP obtained after having stopped the second-stage digestion  
33  
34 after a number of days equal to 15, 10 and 5, respectively, so that the overall duration of the two-stage  
35  
36 AD process was of 20 days. In this way, it could be seen (Figure 6) that, for the untreated digestate  
37  
38 samples, the useful SMPs were of 0.091, 0.059 and 0.028 Nm<sup>3</sup> CH<sub>4</sub>/kg VS added, only 95%, 82% and  
39  
40 53% of the values obtained at the end of the tests. That determined overall SMPs, after a 20-day, two-  
41  
42 stage digestion, of 0.160, 0.154 and 0.137 Nm<sup>3</sup> CH<sub>4</sub>/kg VS added for the systems with a first-stage  
43  
44 HRT of 5, 10 and 15 days, respectively.

45  
46 It can be pointed out that the above-reported overall SMPs for the untreated WAS samples were  
47  
48 obtained by summing the SMP generated in a semi-continuous (SC) test and the SMP generated in the  
49  
50 subsequent batch (B) test, with different durations, that is (5-d SC + 15-d B), (10-d SC + 10-d B) and  
51  
52

1  
2  
3  
4 (15-d SC + 5-d B). Consequently, values of SMP that increased from 0.137 to 0.160 Nm<sup>3</sup> CH<sub>4</sub>/kg VS  
5  
6 added, by combining shorter first-stage HRTs and longer durations of the batch test, were obtained for  
7  
8 the same substrate (untreated WAS). The observed trend was in good agreement with the well-known  
9  
10 result that batch tests produce from 15 to 30% more biogas than semi-continuous tests with the same  
11  
12 duration in terms of HRT (Ruffino et al., 2015b; Zhang et al., 2013; Zupančič and Jemec, 2010).  
13  
14

15  
16 For the same digestion conditions considered before, that is (5-d SC + 15-d B), (10-d SC + 10-d B) and  
17  
18 (15-d SC + 5-d B), with an in-between thermal treatment, the overall SMPs, after a 20-day, two-stage  
19  
20 digestion, were of 0.204, 0.194 and 0.196 Nm<sup>3</sup> CH<sub>4</sub>/kg VS added, respectively. Finally, the overall  
21  
22 SMPs, after a 20-day, two-stage digestion, with an in-between hybrid treatment, were of 0.244, 0.211  
23  
24 and 0.213 Nm<sup>3</sup> CH<sub>4</sub>/kg VS added.  
25  
26

27  
28 Tests carried out on a raw WAS sample, collected from the same WWTP, returned a SMP of 0.274  
29  
30 Nm<sup>3</sup> CH<sub>4</sub>/kg VS added when treated with a hybrid (90°C, NaOH 4% TS content, 90 minutes, same  
31  
32 conditions of this work) pre-treatment and digested in the same batch apparatus used in this work  
33  
34 (Campo et al., 2018). In the same study, a sample of digestate that came from a 7-day digestion process  
35  
36 in the WWTP digesters and was subsequently subjected to a hybrid IHT (90°C, NaOH 4% TS content,  
37  
38 90 minutes), produced an overall methane amount of 0.317 Nm<sup>3</sup>/kg VS added. The result obtained in  
39  
40 the present study for the most promising system (0.244 Nm<sup>3</sup> CH<sub>4</sub>/kg VS added, first-stage HRT = 5  
41  
42 days) was well below the values obtained in the previous experience, that is -11% if compared with the  
43  
44 pre-treatment and -23% if compared with the IHT. This demonstrates that batch tests were useful to  
45  
46 make a comparison between systems subjected to different treatment conditions but they were not  
47  
48 adequate to return a methane production value that can be representative of the productions of a plant  
49  
50 working in (semi-) continuous conditions.  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

### 3.4 Future application of IHTs to the sludge line of a full scale WWTP

#### 3.4.1 Technical assessments

In this section a technical analysis based on heat balances was carried out to evaluate the effects of the introduction of IHTs in the sludge line of Castiglione Torinese SMAT WWTP. The three scenarios schematized in Figure 7 were considered for the analysis. In Scenario 1 and Scenario 2 primary and secondary sludge are digested in separated lines. More in detail, in Scenario 1 the heat from the IHT of secondary sludge is recovered for heating primary sludge before digestion. Conversely, in Scenario 2, no heat recovery is applied. Finally, Scenario 3 considers that the primary sludge and the (X-day HRT) digestate from the secondary sludge are mixed after the IHT, and the mixture of the two substrates is subsequently digested in a reactor with HRT equal to 20 days. Both first- and second- stage digesters, for all the considered Scenarios, work in mesophilic conditions (38°C).

First of all, a “zero” Scenario, that refers to the current operating conditions of the sludge line in the WWTP, was presented and its performances were evaluated. Currently, the six 12,000 m<sup>3</sup> anaerobic digesters of Castiglione Torinese SMAT WWTP treat the sludge that come from both primary and secondary settlers. The ratio between primary and secondary sludge, generated in the water line and subsequently digested, is 65:35 on a TS basis (see Section 2.2).

At present, three of the five working digesters (one is periodically in maintenance) are employed for primary sludge and two for secondary sludge. As demonstrated in a previous study (Ruffino et al., 2014), the present modality of management of the sludge line, that includes a preliminary thickening with gravity devices able to thicken sludge to a final TS content of approximately 3.0% in both primary and secondary sludge, and a subsequent AD with no pre- or intermediate treatments, does not guarantee the complete thermal self-sustainment of the digesters, especially during the cold season.

Figure 8 shows the TS content that should be reached in both primary and secondary sludge, through an enhanced thickening process, in order to obtain the complete thermal self-sustainment of the digesters.

The key data used for this assessment were:

- volumes and characteristics of primary and secondary sludge currently produced and treated in the WWTP, that are listed in Table 3;
- SMP of primary and secondary sludge fixed to 0.280 and 0.090 Nm<sup>3</sup>/kg VS, respectively. These data were obtained from measurements of biogas and methane production, on the WWTP digesters fed with primary and secondary sludge, that have been carried out daily for approximately three years;
- the electrical and thermal efficiency of the combined heat and power (CHP) unit used in the WWTP that was of 41.9% and 42.4%, respectively;
- HRT of one-stage digesters fixed to 20 days;
- three values of heat exchange efficiency (50-70-100%) referred to the transfer of the heat generated in the CHP unit to the digesters.

Table 3. Volumes and characteristics of primary and secondary sludge currently produced and treated in the WWTP

	Primary sludge	Secondary sludge
Volumetric flow rate (m <sup>3</sup> /h)	39.5	51.8
Mass flow rate (kg TS/h)	2765	1554
VS/TS ratio	0.72	0.69
Sludge average temperature (°C)	15	15

As shown in Figure 8, the analysis revealed that TS contents of 4.5%, 6.5% and 9.1%, in both primary and secondary sludge, were necessary to make the thermal balance of the sludge line neutral in the case of heat exchange efficiencies of 100%, 70% and 50%, respectively.

1  
2  
3  
4 The thermal and economic assessment of the three Scenarios that consider the introduction of IHTs was  
5  
6 carried out by considering that the primary sludge is digested with a HRT of 20 days as it is, after being  
7  
8 thickened to 7% TS. This value came from the results of the analysis carried out on the “zero” Scenario  
9  
10 that revealed that the self-sustainment of the sludge line, under a mid-thermal efficiency condition  
11  
12 (70%), could be obtained provided that both primary and secondary sludge were thickened to a TS  
13  
14 content in the order of 7%.  
15  
16

17  
18 This assessment considers that the secondary sludge is partially digested for X days (with X = 5, 10 and  
19  
20 15 days, the three HRTs considered in this study), subsequently the first-stage digestate undergoes a  
21  
22 IHT and, finally, it is digested in a second-stage process with a (20-X)-day HRT. For the assessment,  
23  
24 the TS content of secondary sludge was made to change from 3% to 11% and the efficiency of the  
25  
26 thermal exchange was fixed to 50%, 70% and 100%, considering that the heat produced from the  
27  
28 biogas had to be used for the IHTs and the self-sustainment of the digestion processes in mesophilic  
29  
30 conditions. Furthermore, it could be demonstrated that, in the present conditions, the heat lost in the  
31  
32 sludge/digestate treatment, that is the heat necessary to compensate the exchange with the external  
33  
34 environment and to maintain the requested temperature (90°C) for all the duration of the treatment (90  
35  
36 minutes), was only a small fraction (less than 1%) of the heat required for the process (Ruffino et al.,  
37  
38 2014).  
39  
40  
41  
42  
43  
44

45 Figure 9, 10 and 11 and Table S1, S2 and S3 refer to Scenario 1, 2 and 3, respectively. Letters A, B and  
46  
47 C in Figures 9, 10 and 11 refer to HRTs of the first-stage digestion of 5, 10 and 15 days, respectively.  
48  
49

50 Figure 9 highlights two main outcomes of the analysis. On the one hand, it could be observed that the  
51  
52 efficiency with which the heat produced from the CHP unit was transferred to the processes that need  
53  
54 heating (IHTs and digestion processes) heavily affected the performances of the whole system. On the  
55  
56 other hand, irrespective of the thermal exchange efficiency, the solution that considered a 5-day long  
57  
58 first-stage digestion, followed by an IHT and completed by a final 15-day digestion, for secondary  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 sludge, proved to be the most efficient. More in detail, in the presence of a thermal exchange efficiency  
5  
6 of 100%, a TS content of secondary sludge from 4.1 to 4.3%, depending on the combination of the  
7  
8 duration of the first and second stage of AD, was sufficient to make the thermal balance neutral. If the  
9  
10 thermal exchange efficiency decreased to 70% or 50%, the TS content had to be increased to  
11  
12 approximately 6% or 8%, respectively. It has to be taken into account that, in the case of IHTs  
13  
14 application, only the second-stage digestion process could benefit of the improved flowing capacity of  
15  
16 the sludge samples that originated from the application of hybrid treatments. In a previous study, it was  
17  
18 demonstrated that thermal treatments, carried out at 90°C for at least 90 minutes, made the flowing  
19  
20 capacity of a 9% TS sludge equal to the one of a sludge with half TS content (Ruffino et al., 2015a). It  
21  
22 is worth mentioning that a Cambi process can easily work with sludge with a TS content in the order of  
23  
24 16-17% (Svensson et al., 2018).  
25  
26  
27  
28  
29

30  
31 Not surprisingly, the results of Scenario 2 (shown in Figure 10), compared to those of Scenario 1,  
32  
33 highlighted that when the heat recovered from the IHT was not used to heat the primary sludge before  
34  
35 digestion, higher TS contents were required to make the thermal balance neutral. Theoretically, in the  
36  
37 case of a thermal exchange efficiency of 50%, the TS content of secondary sludge rose to values in the  
38  
39 order of 25%.  
40  
41  
42

43 Results of Scenario 3, shown in Figure 11, were quite similar to those of Scenario 1 in terms of TS  
44  
45 content of secondary sludge necessary to make the thermal balance of the overall system neutral. The  
46  
47 required TS content of secondary sludge was of approximately 4.2, 6.1 and 8.7% for thermal exchange  
48  
49 efficiencies of 100%, 70% and 50%, respectively. Even in this case, due to the weight of the thermal  
50  
51 exchange efficiency on the overall process, the distribution of the length of the first and second stage of  
52  
53 the AD process was only a minor issue in defining the TS content of secondary sludge.  
54  
55 Notwithstanding the similarities of Scenario 1 and Scenario 3 in the TS content of secondary sludge,  
56  
57 Scenario 3 had to be preferred to Scenario 1 because of the best conditions that could be found in the  
58  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 second-stage digester. In fact, it is necessary to mention the effect of hybrid processes on ammonia  
5  
6 release (Kuglarz et al., 2013; Ruffino et al., 2016) and final pH of the treated substrates. The mixing  
7  
8 with primary sludge helps in controlling ammonia concentration, adjusts pH to neutral values and  
9  
10 dilutes the load of sodium ions introduced with the NaOH used for the treatment (Pinto et al., 2016;  
11  
12 Sarwar et al., 2018).  
13  
14  
15  
16  
17

### 18 **3.4.2 Economic balance**

19 The economic assessment carried out in this section of the study was aimed to:  
20  
21

- 22 • identify the cost of the initial investment that can be borne by the WWTP and recovered after a  
23  
24 fixed time because of the gain that would derive from the application of the novel IHTs under  
25  
26 Scenarios 1 or 3;  
27  
28  
29  
30
- 31 • compare the revenues that the WWTP presently obtains from the production of electricity with  
32  
33 the gain that would derive from the implementation of one of the new Scenarios (1 or 3).  
34  
35

36 The Net Present Value (NPV) method was employed for the economic analysis. The NPV is the  
37  
38 summation of the present value of a series of present and future cash flows as shown in Equation 2.  
39  
40 Because NPV accounts for the time value of money, NPV provides a method for evaluating and  
41  
42 comparing products with cash flows spread over many years.  
43  
44

$$45 \quad NPV(i, N) = \sum_{t=0}^N \frac{R_t}{(1+i)^t} \quad (2)$$

46  
47  
48  
49  
50

51 Where

52 t - time of the cash flow;

53 i – discount rate, i.e. the return that could be earned per unit of time on an investment with similar risk;

54 R<sub>t</sub> – net cash flow (i.e cash inflow minus cash outflow), at time t  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 In this study the unknown cost of the initial investment ( $I_0$ ) was evaluated by using Equation 3 and  
5  
6 assuming a pay-back time of three or five years  
7

$$NPV(i, N) = -I_0 + \sum_{t=0}^N \frac{R_t}{(1+i)^t} \quad (3)$$

8  
9  
10  
11 Where the time of cash flow and the discount rate were fixed to 20 years and 5%, respectively.  
12

13  
14 For the estimation of cash inflows and outflows, the four items listed in the follow were considered:  
15

- 16  
17 1. the cost of NaOH used for the application of the novel IHTs, fixed equal to 0.450 €/kg (Solvay  
18  
19 2018);  
20
- 21  
22 2. the avoided cost of the methane, used as an auxiliary fuel, of which the WWTP must presently  
23  
24 be supplied to keep the energy balance of the sludge line neutral, especially in the worst  
25  
26 weather conditions – i.e. winter time (Ruffino et al., 2014);  
27
- 28  
29 3. the cost of energy to thicken the sludge in a dynamic centrifuge and, finally,  
30
- 31  
32 4. the bonus granted for the production of green energy.  
33

34  
35  
36 With reference to item n. 3, it has to be considered that TS contents in the order of 6% or more, like  
37  
38 those required to make the energy balance neutral under Scenario 1 or 3, can be obtained only by using  
39  
40 specific dynamic thickeners. The average volumetric flow rate of sludge generated in the WWTP  
41  
42 (approximately 540 m<sup>3</sup>/h), with their average TS content (from 0.8 to 2.5% TS, depending on the  
43  
44 sludge type, primary or secondary), requires a rotary screw thickener with a specific power of 35  
45  
46 W/m<sup>3</sup>. This piece of equipment consumes on average 0.03 kWh/m<sup>3</sup> to thicken sludge to TS values up to  
47  
48 7%. It is worthwhile to note that the hourly unit cost for the energy required for the dynamic thickener  
49  
50 is of approximately one order of magnitude less than the hourly unit cost required for the NaOH used in  
51  
52 IHTs (2.35 €/h vs. 28.0 €/h).  
53  
54  
55

56  
57  
58 With reference to item n. 4, it must be mentioned that in Italy the electric energy produced from  
59  
60 renewable sources benefits from incentives. According to DM 6/07/2012, biogas from AD of sewage  
61  
62

1  
2  
3  
4 sludge that fuels endothermic engines is included in renewable sources. From 2016, the bonuses  
5  
6 granted for the production of green energy are calculated by using GRIN application (GSE, 2018).

7  
8  
9 According to this method, the incentive rate (IR) is calculated by the formula

10  
11 
$$IR = k \cdot (180 - Re) \cdot 0.78$$

12  
13  
14 Where

- 15  
16  
17
  - 180 is the reference value of a green certificate (equal to 180 €/MWh);
  - Re is equal to the sale price of electricity defined by the Authority annually on the basis of the  
18  
19 economic conditions recorded on the market in the previous year. For year 2016 Re was of  
20  
21 42.38 €/MWh;
  - k is a constant, the value of which depends on the type of used renewable source; for the case  
22  
23 considered in this study k was equal to 0.8.

24  
25  
26  
27  
28  
29  
30

31 The WWTP considered in this study will be having access to incentives until May 2023. Furthermore,  
32  
33 for the economic analysis here reported, the value of Re was kept constant and equal to the value fixed  
34  
35 (42.38 €/MWh) for 2016 (i.e. the first year in which GRIN application started to be used).

36  
37  
38 From the outcomes of the economic analysis, as can be seen in Figure 12, it resulted that the WWTP  
39  
40 could bear investments in the order of 4.7 M€ and 7.2 M€, that could be recovered after three or five  
41  
42 years respectively, to enhance the performances of the sludge line. The change of the curves' slope  
43  
44 recorded after 2023 depended on the expiration of incentives.

45  
46  
47  
48 In the current operating condition, the AD process carried out on the sludge from the WWTP generated  
49  
50 2730 kWe. Future potential increases in the amount of produced energy, that could derive from the  
51  
52 application of IHTs under Scenario 1 or Scenario 3, and in the consequent revenues from the produced  
53  
54 electricity are detailed in Table 4.  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

Table 4. Increases in the amount of produced energy and in the revenues from the produced electricity deriving from the application of IHTs under Scenario 1 or Scenario 3

Scenario	Energy increase (%)	Revenue increase (%)	AD stage I duration (d)	AD stage II duration (d)
1	25.3	34.3	5	15
1	19.9	28.2	10	10
1	20.2	28.6	15	15
3	25.8	34.8	5	20
3	23.1	31.9	10	20
3	22.8	31.5	15	20

It can be seen from Table 4 that, under both Scenarios, the introduction of IHTs determined an average energy increase in the order of 20%; the consequent increase in the revenues from the production of electricity was of 30% or more.

### Conclusions

This study wanted to analyze the effect of the duration of the first-stage AD on the overall performance of a system made of two stages of digestions with an in-between digestate treatment. The three digestates with controlled HRTs (5 – 10 and 15 days) were obtained by using a 10L semi-continuous mesophilic reactor.

The obtained results demonstrated that:

- A two-stage digestion, with a low-temperature (90°C) thermal IHT, produced a SMP in the order of 0.205 Nm<sup>3</sup>/kg VS added, irrespective of the duration of the first stage, when the duration of the second-stage digestion was fixed to 20 days.
- In the presence of an IHT, the difference between a short (5 days) and a medium (10-15 days) duration of the first-stage digestion became evident, with SMPs in the order of 0.247 and 0.230

1  
2  
3  
4 Nm<sup>3</sup>/kg VS added, respectively. Even in this case the duration of the second-stage digestion  
5  
6 was fixed to 20 days.  
7

- 8  
9 • When the duration of the overall process was fixed to 20 days (i.e. not all the methane  
10 generated in the batch tests was used to calculate the overall production), the effect of the first-  
11 stage HRT was amplified, with a SMP reduction, from short to medium first-stage durations, in  
12 the order of 15%.  
13  
14 • Thermal sustainability of the application of IHTs at a full scale plant required an adequate  
15 thickening of sludge and an efficient heat exchange between donor (sludge after treatment) and  
16 acceptor (cold sludge before digestion) agents. In the case of separated digestion of raw  
17 primary, thickened to 7%, and treated digestate from secondary sludge (Scenario 1), a TS  
18 content of secondary sludge in the range 4.1-4.3%, depending on the combination of the  
19 duration of the first and second stage of AD, was required to make the thermal balance neutral,  
20 with a thermal exchange efficiency equal to 100%. If the thermal exchange efficiency decreased  
21 to 70% or 50%, the TS content had to be increased to 6% or 8%, respectively.  
22  
23 • When primary sludge (7%) was mixed with the treated digestate from secondary sludge to  
24 undergo a final 20-day digestion (Scenario 3), TS contents of secondary sludge similar to those  
25 of Scenario 1 (that is of 4.2, 6.1 and 8.7%, for thermal exchange efficiencies of 100%, 70% and  
26 50%, respectively) were required.  
27  
28 • Irrespective of the considered Scenario (1 or 3), the introduction of IHTs in the WWTP would  
29 determine an average energy increase in the order of 20% and a consequent increase in the  
30 revenues from the production of electricity of 30% or more.  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4  
5  
6  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

## Acknowledgements

This research was funded by SMAT, Societa' Metropolitana Acque Torino.

The authors wish to thank Rocco Eufemia and Luca Polimeno for the support in the experimental activity.

Three anonymous reviewers are gratefully acknowledged for comments and suggestions useful to improve the quality of the manuscript.

1  
2  
3  
4 References  
5

6 APHA, AWWA, WEF, 2012. Standard methods for the examination of water and wastewater, 22st  
7 ed., Washington: American Public Health Association, ISBN 978-087553-013-0.  
8

9 Aragón-Briceño C., Ross A.B., Camargo-Valero M.A., 2017. Evaluation and comparison of  
10 product yields and bio-methane potential in sewage digestate following hydrothermal treatment.  
11 *Appl. Energy*, 208, 1357–1369. DOI: 10.1016/j.apenergy.2017.09.019.  
12

13 Bjerg-Nielsen M., Ward A.J., Møller H.B., Ottosen L.D.M., 2018. Influence on anaerobic digestion  
14 by intermediate thermal hydrolysis of waste activated sludge and co-digested wheat straw. *Waste*  
15 *Manage.* 72, 186–192. DOI: 10.1016/j.wasman.2017.11.021.  
16  
17

18 Campo G., Cerutti A., Zanetti M.C., Scibilia G., Lorenzi E., Ruffino B., 2017. Pre- and  
19 intermediate hybrid treatments for the improvement of anaerobic digestion of sewage sludge:  
20 Preliminary results. *J. Environ. Eng.* 143(9), 04017052. DOI: 10.1061/(ASCE)EE.1943-  
21 7870.0001249.  
22

23 Campo G., Cerutti A., Zanetti M.C., Scibilia G., Lorenzi E., Ruffino B., 2018. Enhancement of  
24 waste activated sludge (WAS) anaerobic digestion by means of pre- and intermediate treatments.  
25 Technical and economic analysis at a full scale WWTP. *J. Environ. Manage.* 216, 372-382. DOI:  
26 10.1016/j.jenvman.2017.05.025.  
27

28 Cano R., Pérez-Elvira S.I., Fdz-Polanco F., 2015. Energy feasibility study of sludge pretreatments:  
29 a review. *Appl. Energy* 149, 176-185. DOI: 10.1016/j.apenergy.2015.03.132.  
30

31 Carrère H., Dumas C., Battimelli A., Batstone D.J., Delgenès J.P., Steyer J.P., Ferrer I., 2010.  
32 Pretreatment methods to improve sludge anaerobic degradability: a review. *J. Hazard. Mater.* 183  
33 (1-3), 1-15. DOI: 10.1016/j.jhazmat.2010.06.129.  
34

35 Córdova Lizama A., Carrera Figueiras C., Zepeda Pedreguera A., Ruiz Espinoza J.E., 2018. Effect  
36 of ultrasonic pretreatment on the semicontinuous anaerobic digestion of waste activated sludge with  
37 increasing loading rates. *Int. Biodeter. Biodegr.* 130, 32–39. DOI: 10.1016/j.ibiod.2018.03.013.  
38

39 Dohányos M., Záborská J., Jeníček P., 1997. Enhancement of sludge anaerobic digestion by use of  
40 a special thickening centrifuge. *Water Sci. Technol.* 36, 145-153. DOI: 10.1016/S0273-  
41 1223(97)00677-X  
42

43 Grübel K., Suschka J., 2015. Hybrid alkali-hydrodynamic disintegration of waste-activated sludge  
44 before two-stage anaerobic digestion process. *Environ. Sci. Pollut. Res.* 22, 7258–7270. DOI:  
45 0.1007/s11356-014-3705-y.  
46

47 GSE, 2018. <https://www.gse.it/servizi-per-te/fonti-rinnovabili/impianti-a-fonti-rinnovabili-grin> (in  
48 Italian). Last accessed 12/2018.  
49

50 Jimenez J., Aemig Q., Doussiet N., Steyer J.P., Houot S., Patureau D., 2015. A new organic matter  
51 fractionation methodology for organic wastes: bioaccessibility and complexity characterization for  
52 treatment optimization. *Bioresource Technol.* 194, 344–353. DOI: 10.1016/j.biortech.2015.07.037.  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

- 1  
2  
3  
4 Kuglarz M., Karakashev D., Angelidaki I., 2013. Microwave and thermal pretreatments as methods  
5 for increasing the biogas potential of secondary sludge from municipal wastewater treatment plants.  
6 *Bioresour. Technol.* 134, 290–297. DOI: 10.1016/j.biortech.2013.02.001.  
7  
8  
9 Leite W., Scandolaro Magnus B., Bittencourt Guimarães L., Gottardo M., Belli Filho P., 2017.  
10 Feasibility of thermophilic anaerobic processes for treating waste activated sludge under low HRT  
11 and intermittent mixing. *J. Environ. Manage.* 201, 335-344. DOI: 10.1016/j.jenvman.2017.06.069.  
12  
13 Nordmann W., 1977. Die Überwachung der Schlammfäulung. KA-Informationen für das  
14 Betriebspersonal, Beilage zur Korrespondenz Abwasser, 3/77.  
15  
16 Ortega-Martinez E., Sapkaite I., Fdz-Polanco F., Donoso-Bravo A., 2016. From pre-treatment  
17 toward inter-treatment. Getting some clues from sewage sludge biomethanation. *Bioresource*  
18 *Technol.* 212, 227–235. DOI: 10.1016/j.biortech.2016.04.049.  
19  
20 Pinto N., Carvalho A., Pacheco J., Duarte E., 2016. Study of different ratios of primary and waste  
21 activated sludges to enhance the methane yield. *Water Environ. J.* 30, 203–210. DOI:  
22 10.1111/wej.12188.  
23  
24 Rennuit C., Triolo J.M., Eriksen S., Jimenez J., Carrère H., Hafner S.D., 2018. Comparison of pre-  
25 and inter-stage aerobic treatment of wastewater sludge: Effects on biogas production and COD  
26 removal. *Bioresource Technol.* 247, 332–339. DOI: 10.1016/j.biortech.2017.08.128.  
27  
28 Roeleveld P.J., van Loosdrecht M.C., 2002. Experience with guidelines for wastewater  
29 characterisation in The Netherlands. *Water Sci. Technol. J. Int. Assoc. Water Pollut. Res.* 45 (6),  
30 77-87.  
31  
32 Ruffino, B., Campo, G., Zanetti, M.C., Genon, G., 2014. Improvement of activated sludge  
33 anaerobic digestion: thermal and economical perspectives. *WIT Trans. Ecol. Environ.* 190, 979–  
34 991. DOI: 10.2495/EQ140922.  
35  
36 Ruffino B., Campo G., Genon G., Lorenzi E., Novarino D., Scibilia G., Zanetti M.C., 2015a.  
37 Improvement of anaerobic digestion of sewage sludge in a wastewater treatment plant by means of  
38 mechanical and thermal pre-treatments: Performance, energy and economical assessment.  
39 *Bioresource Technol.* 175, 298-308. DOI: 10.1016/j.biortech.2014.10.071.  
40  
41 Ruffino B., Fiore S., Roati C., Campo G., Novarino D., Zanetti M.C., 2015b. Scale effect of  
42 anaerobic digestion tests in fed-batch and semi-continuous mode for the technical and economic  
43 feasibility of a full scale digester. *Bioresour Technol.* 182, 302-313. DOI:  
44 10.1016/j.biortech.2015.02.021.  
45  
46 Ruffino B., Campo G., Cerutti A., Zanetti M.C., Lorenzi E., Scibilia G., Genon G., 2016.  
47 Preliminary technical and economic analysis of alkali and low temperature thermo-alkali  
48 pretreatments for the anaerobic digestion of waste activated sludge. *Waste Biomass Valor.* 7, 667-  
49 675. DOI: 10.1007/s12649-016-9537-x.  
50  
51 Sarwar R., Elbeshbishy E., Parkera W.J., 2018. Codigestion of high pressure thermal hydrolysis-  
52 treated thickened waste activated sludge with primary sludge in two-stage anaerobic digestion.  
53 *Environ. Progr. Sustain.* 37(1), 425-433. DOI: 10.1002/ep.12700.  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

- 1  
2  
3  
4 Shana A., Ouki S., Asaadi M., Pearce P., Mancini G., 2013. The impact of intermediate thermal  
5 hydrolysis on the degradation kinetics of carbohydrates in sewage sludge. *Bioresource Technol.*  
6 137, 239–244. DOI: 10.1016/j.biortech.2013.03.121.  
7  
8 Shana A.D., Ouki S., Asaadi M., Pearce P., 2015. The impact of intermediate thermal hydrolysis  
9 process and conventional thermal hydrolysis process on biochemical composition during anaerobic  
10 digestion of sewage sludge. *Proceedings of 20th European Biosolids & Organic Resources*  
11 *Conference & Exhibition, Manchester, 9-11/11/2015*, pp. 1-15.  
12  
13 Solvay, 2018. <http://www.solvaychemicals.com/EN/Home.aspx>. Last accessed 12/2018.  
14  
15 Svensson K., Kjølraug O., Higgins M.J., Linjordet R., Horn S.J., 2018. Post-anaerobic digestion  
16 thermal hydrolysis of sewage sludge and food waste: Effect on methane yields, dewaterability and  
17 solids reduction. *Water Res.* 132, 158-166. DOI: 10.1016/j.watres.2018.01.008.  
18  
19 Takashima M., Tanaka Y., 2014. Acidic thermal post-treatment for enhancing anaerobic digestion  
20 of sewage sludge. *J. Environ. Chem. Eng.* 2 (2), 773-779. DOI: 10.1016/j.jece.2014.02.018.  
21  
22 VDI Standard, 2006. VDI 4630 Fermentation of Organic Materials. Characterization of the  
23 substrate, sampling, collection of material data, fermentation tests, p. 92. available on line at:  
24 <http://www.vdi.eu/guidelines>. Last accessed 12/2018.  
25  
26 Waclawek S., Grübel K., Silvestri D., Padil V.V.T., Waclawek M., Černík M., Varma R.S., 2019.  
27 Disintegration of Wastewater Activated Sludge (WAS) for Improved Biogas Production. *Energies*,  
28 12, 21. DOI: 10.3390/en1201002.  
29  
30 Wang J., Li Y., 2016. Synergistic pretreatment of waste activated sludge using CaO<sub>2</sub> in  
31 combination with microwave irradiation to enhance methane production during anaerobic  
32 digestion. *Appl. Energy*, 183, 1123–1132. DOI: 10.1016/j.apenergy.2016.09.042.  
33  
34 Wei W., Wang Q., Zhang L., Laloo A., Duan H., Batstone D.J., Yuan Z., 2018. Free nitrous acid  
35 pre-treatment of waste activated sludge enhances volatile solids destruction and improves sludge  
36 dewaterability in continuous anaerobic digestion. *Water Res.* 130, 13-19. DOI:  
37 10.1016/j.watres.2017.11.050.  
38  
39 Williams T.O., Burrowes P., Fries K., Newbery C., Whitlock D., 2015. Treatment of WAS with  
40 thermal hydrolysis and mesophilic anaerobic digestion. *Proceedings of 20th European Biosolids &*  
41 *Organic Resources Conference & Exhibition, Manchester, 9-11/11/2015*, pp. 1-8.  
42  
43 Yuan T., Cheng Y., Zhang Z., Lei Z., Shimizu K., 2019. Comparative study on hydrothermal  
44 treatment as pre- and post-treatment of anaerobic digestion of primary sludge: Focus on energy  
45 balance, resources transformation and sludge dewaterability. *Appl. Energy*, 239, 171-180. DOI:  
46 10.1016/j.apenergy.2019.01.206  
47  
48 Zhen G., Lu X., Kato H., Zhao Y., Li Y.Y., 2017. Overview of pretreatment strategies for  
49 enhancing sewage sludge disintegration and subsequent anaerobic digestion: Current advances,  
50 full-scale application and future perspectives. *Renew. Sust. Energ. Rev.* 69, 559-577. DOI:  
51 10.1016/j.rser.2016.11.187.  
52  
53 Zhang C., Su H., Tan T., 2013. Batch and semi-continuous anaerobic digestion of food waste in a  
54 dual solid–liquid system. *Bioresour. Technol.* 145, 10–16. DOI: 10.1016/j.biortech.2013.03.030.  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

1  
2  
3  
4 Zupančič G.D., Jemec A., 2010. Anaerobic digestion of tannery waste: semicontinuous and  
5 anaerobic sequencing batch reactor processes. *Bioresour. Technol.* 101, 26–33. DOI:  
6 10.1016/j.biortech.2009.07.028.  
7  
8  
9  
10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21  
22  
23  
24  
25  
26  
27  
28  
29  
30  
31  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47  
48  
49  
50  
51  
52  
53  
54  
55  
56  
57  
58  
59  
60  
61  
62  
63  
64  
65

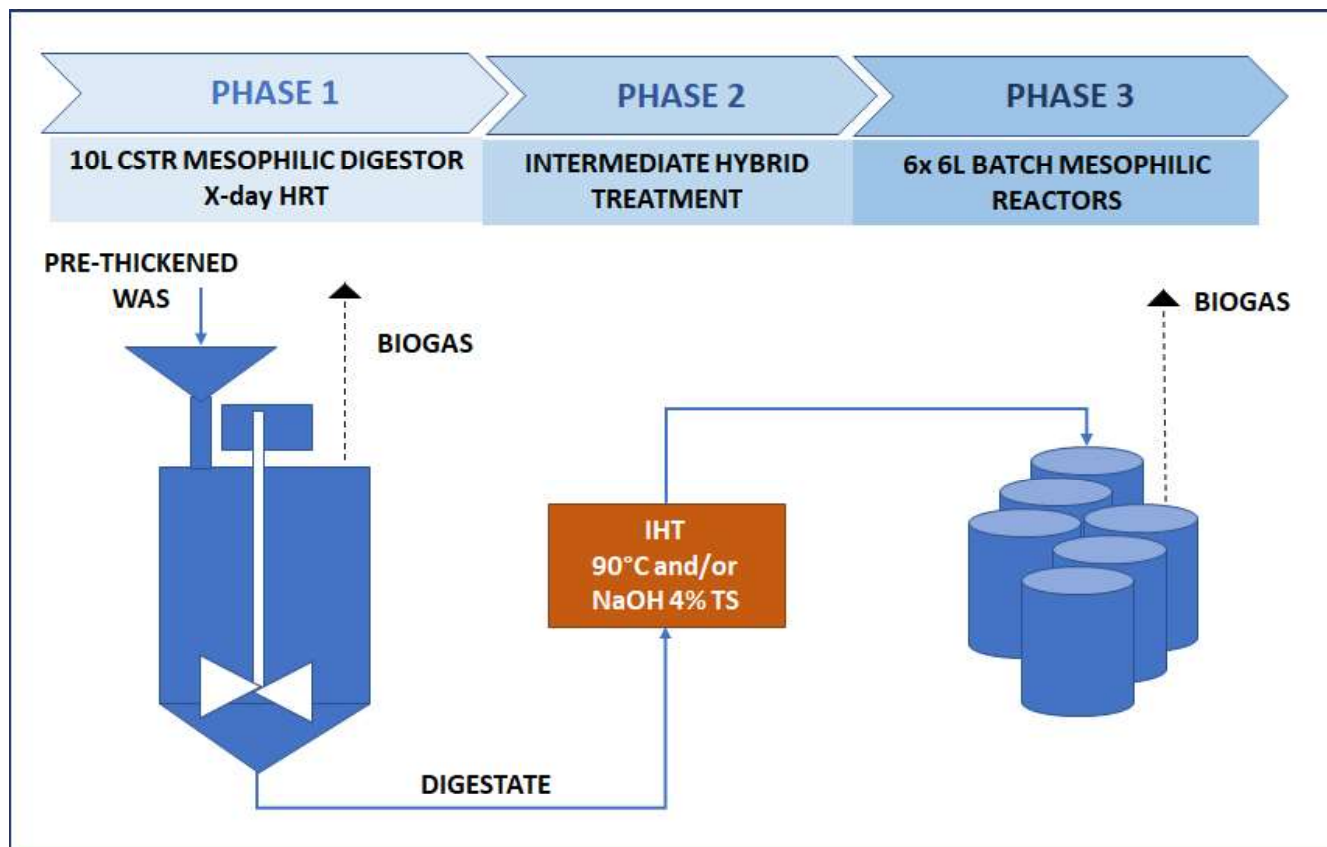


Figure 1. Scheme of the system used for the study

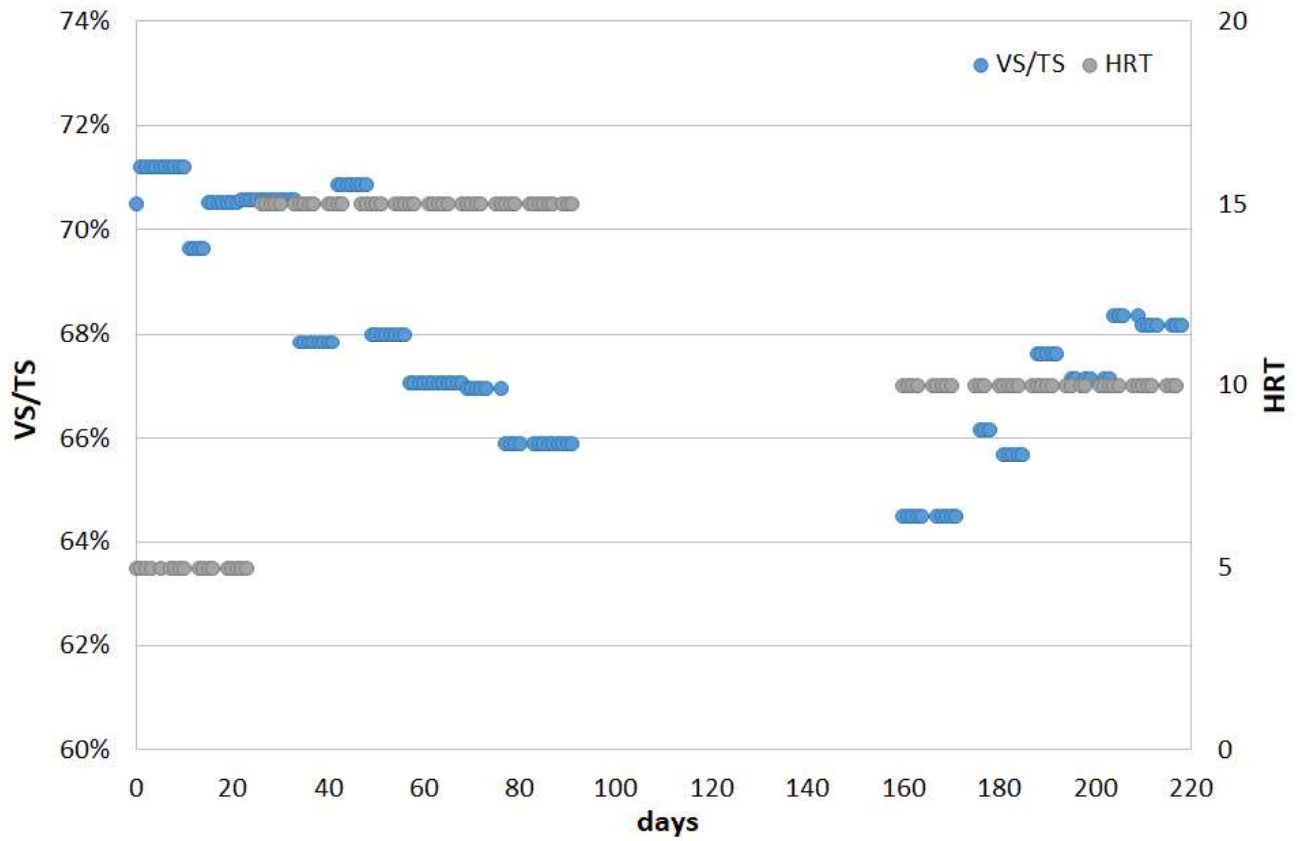


Figure 2. Ratio between volatile and total solids (% VS/TS) for the whole duration of the study

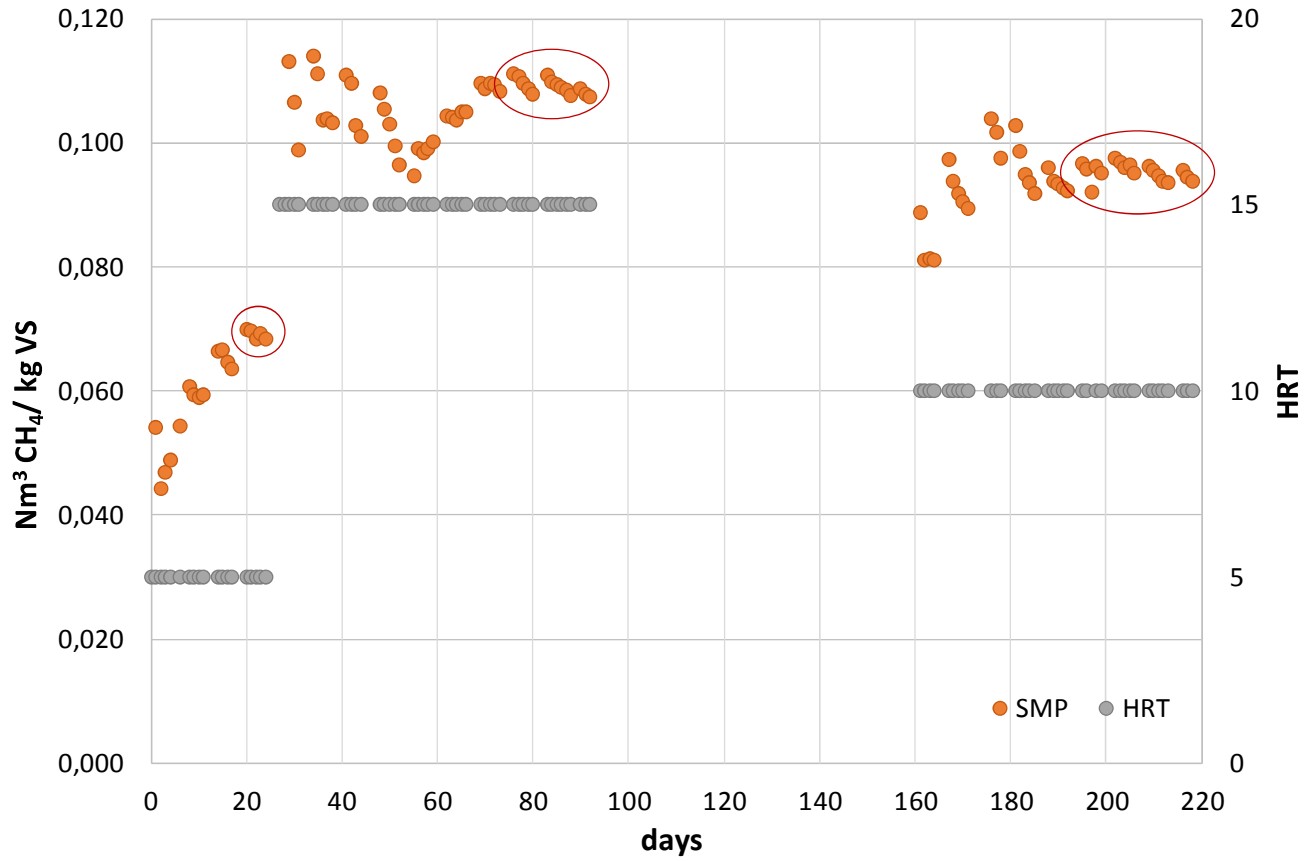


Figure 3. Evolution of the SMP from the semi-continuous first-stage digester in the three digestion runs (HRTs = 5, 10 and 15 days)

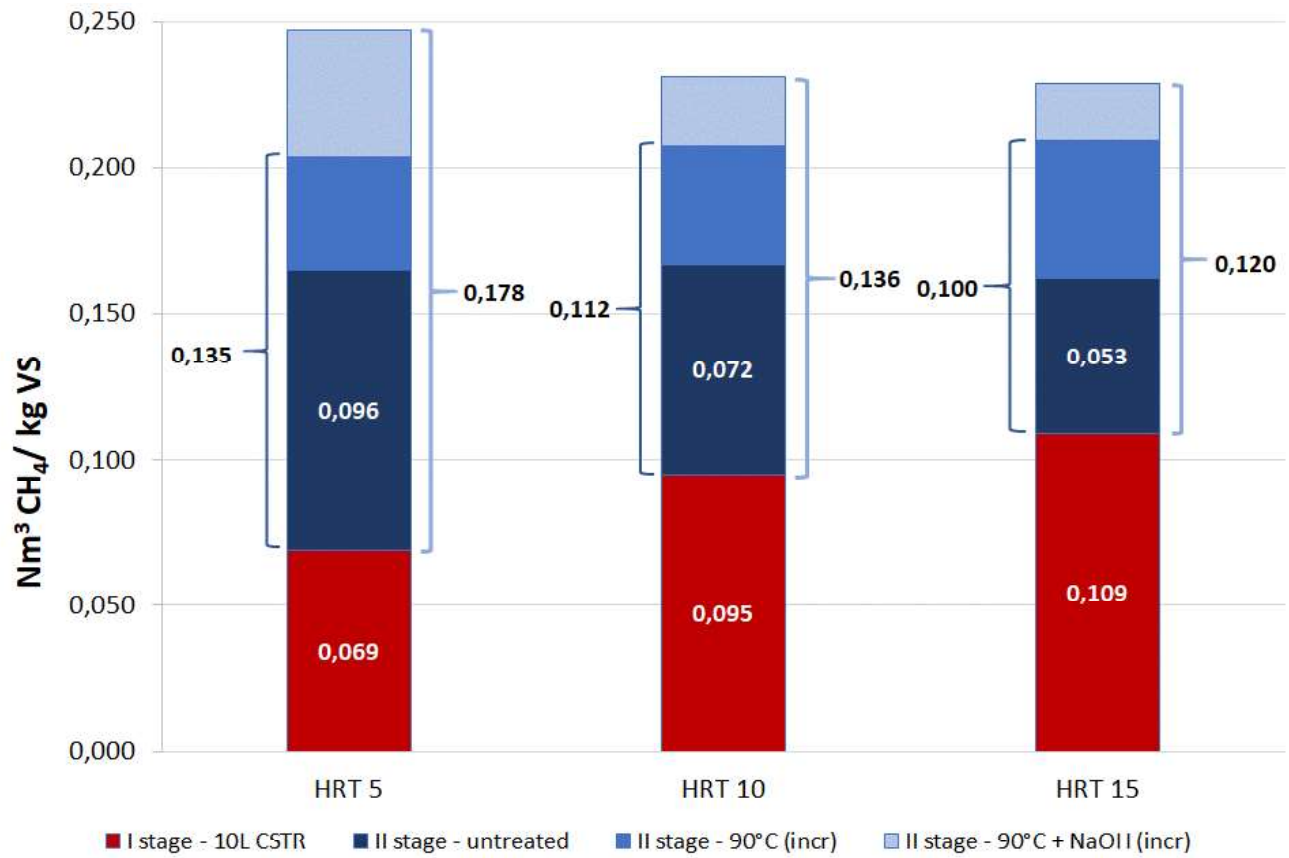


Figure 4. Overall SMP generated by the three systems (first stage HRT = 5, 10, 15 days)

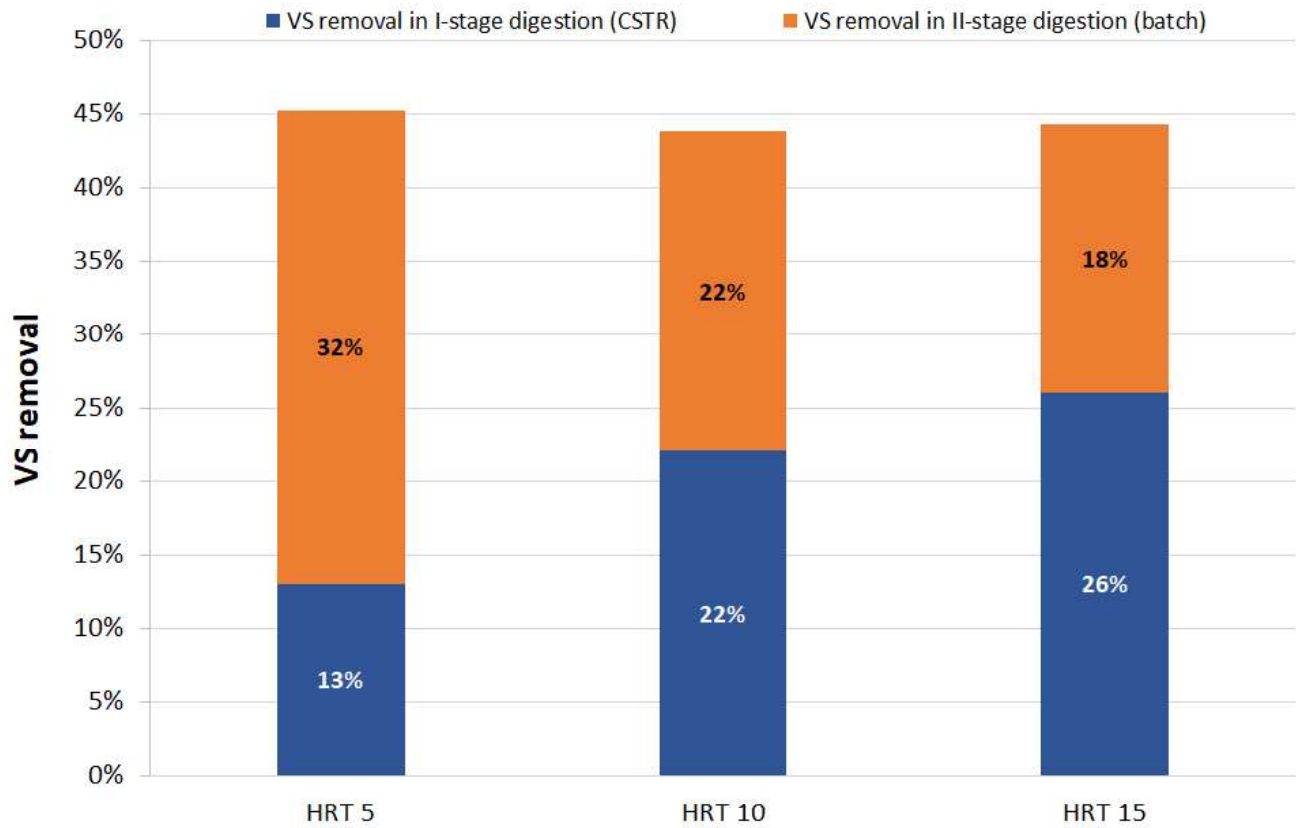


Figure 5. Overall VS removal obtained in the combination of the two digestion processes, with an in-between hybrid treatment, for the systems with a first-stage HRT equal to 5, 10 and 15 days

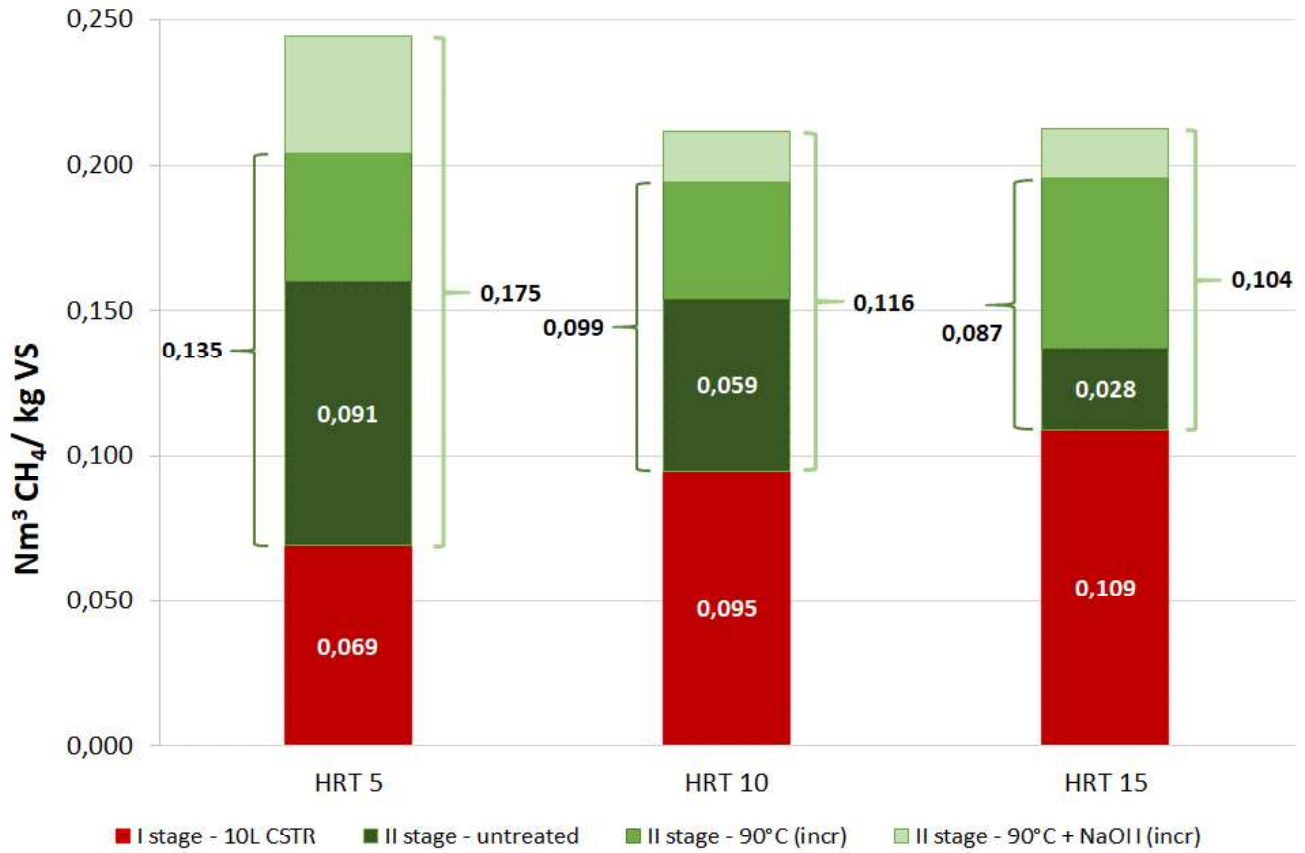


Figure 6. Overall SMP generated by the three systems (first stage HRT = 5, 10, 15 days) when the overall duration of the two-stage system was fixed to 20 days.

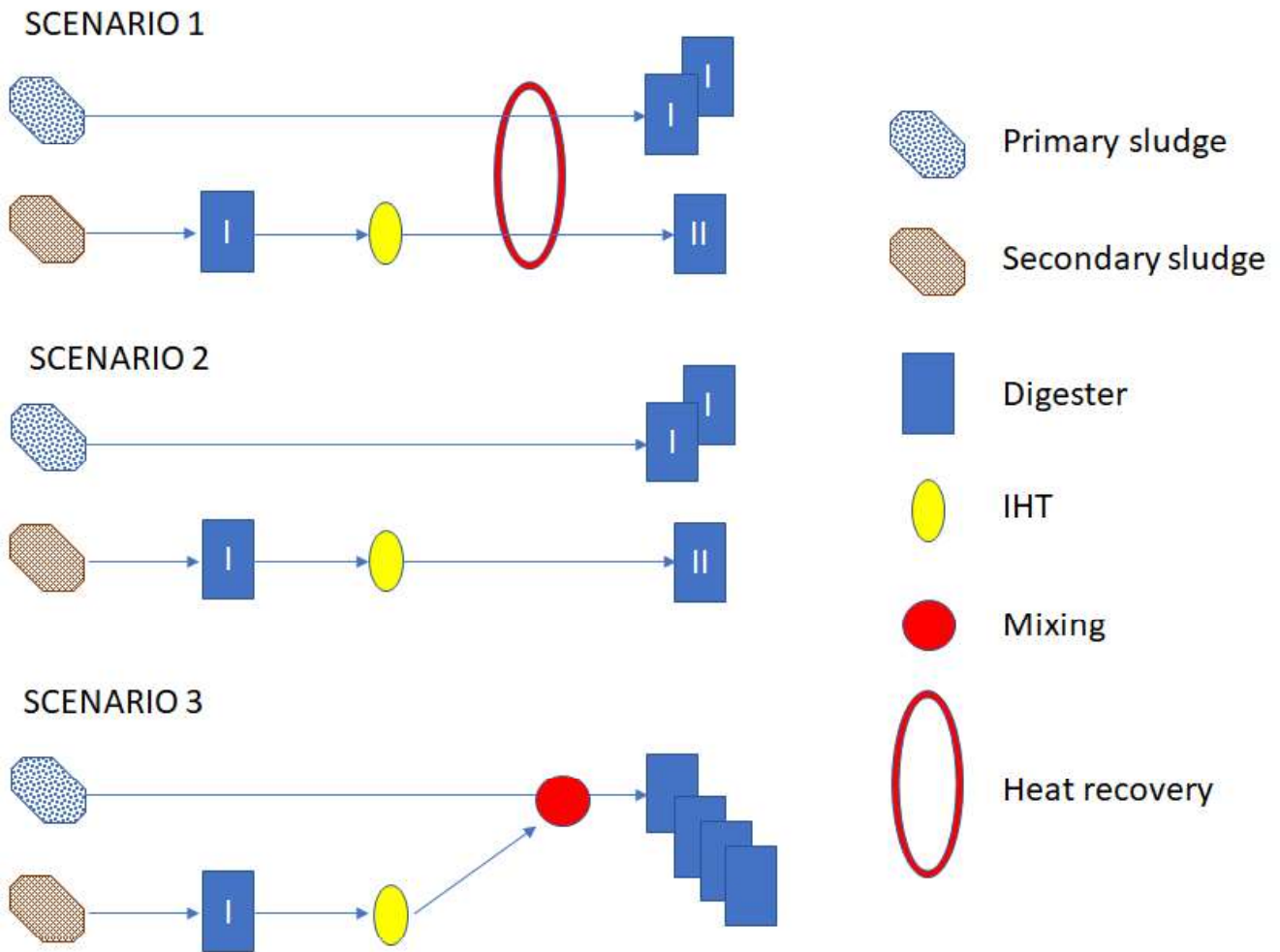


Figure 7. Scenarios considered for the technical and economic assessments at a full scale

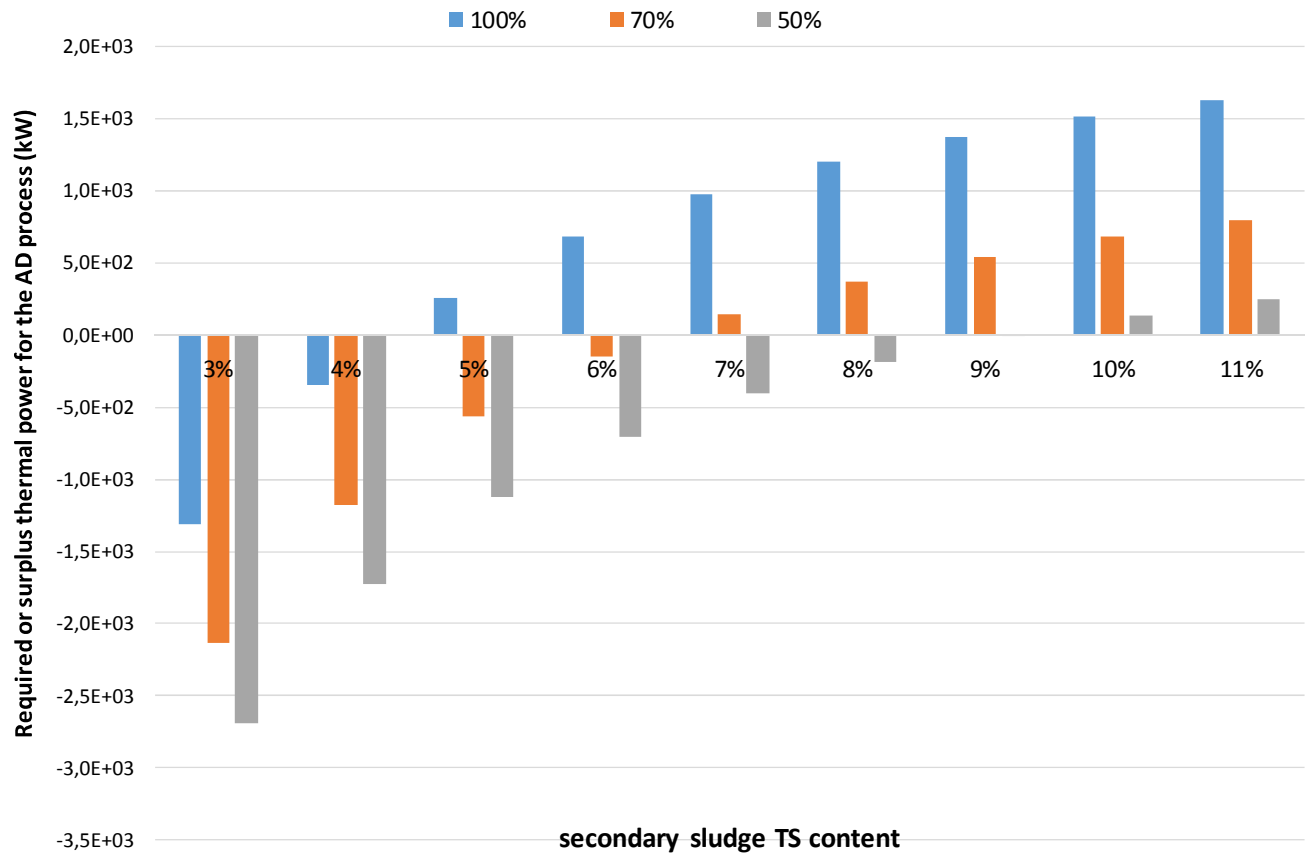


Figure 8. Required or surplus thermal power for the AD process for Scenario 0 that results from the technical assessment. Percentage values (100% - 70% - 50%) refer to the efficiency with which the heat produced from the CHP unit was transferred to the processes that need heating.

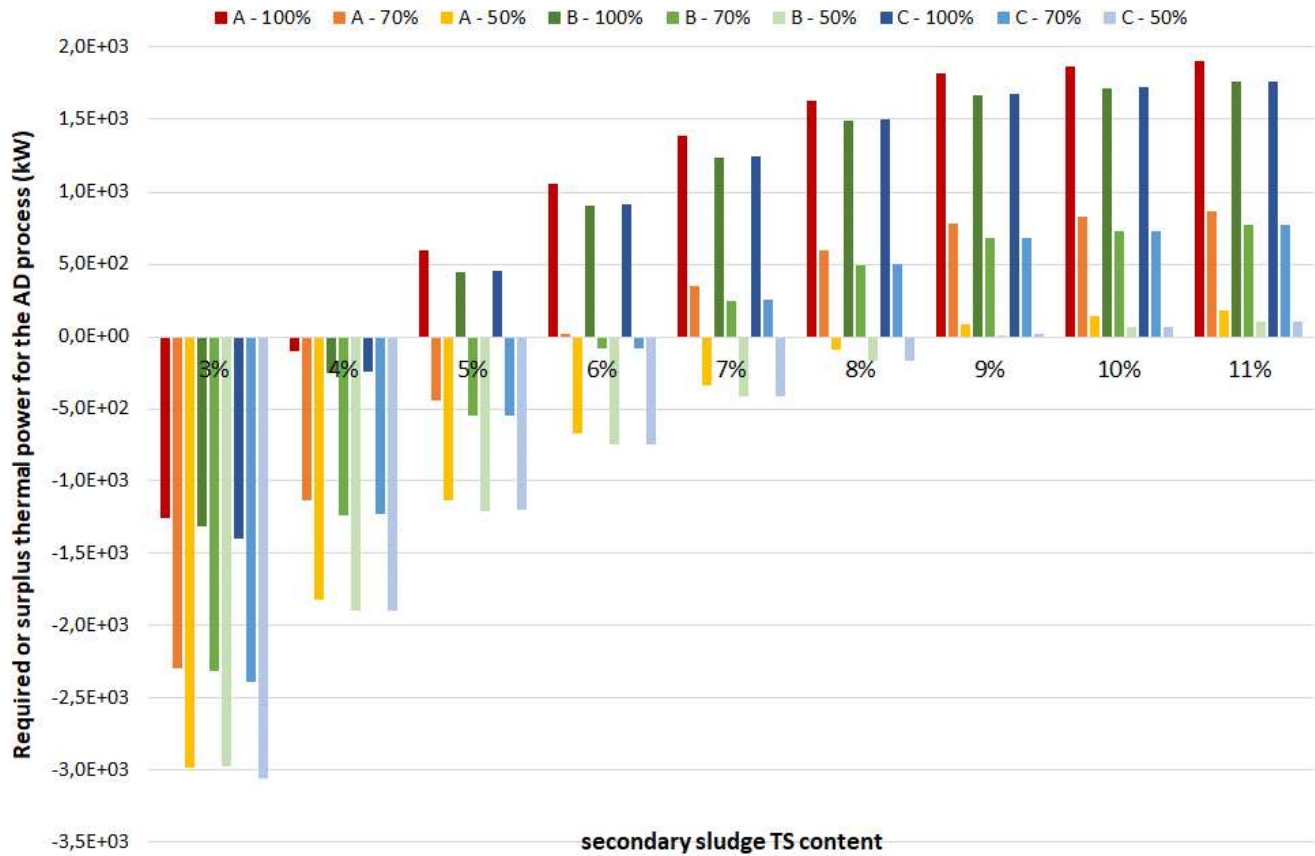


Figure 9. Required or surplus thermal power for the AD process for Scenario 1 that results from the technical assessment. Letters A, B and C refer to HRTs of the first-stage digestion of 5, 10 and 15 days, respectively. Percentage values (100% - 70% - 50%) refer to the efficiency with which the heat produced from the CHP unit was transferred to the processes that need heating (IHTs and digestion processes).

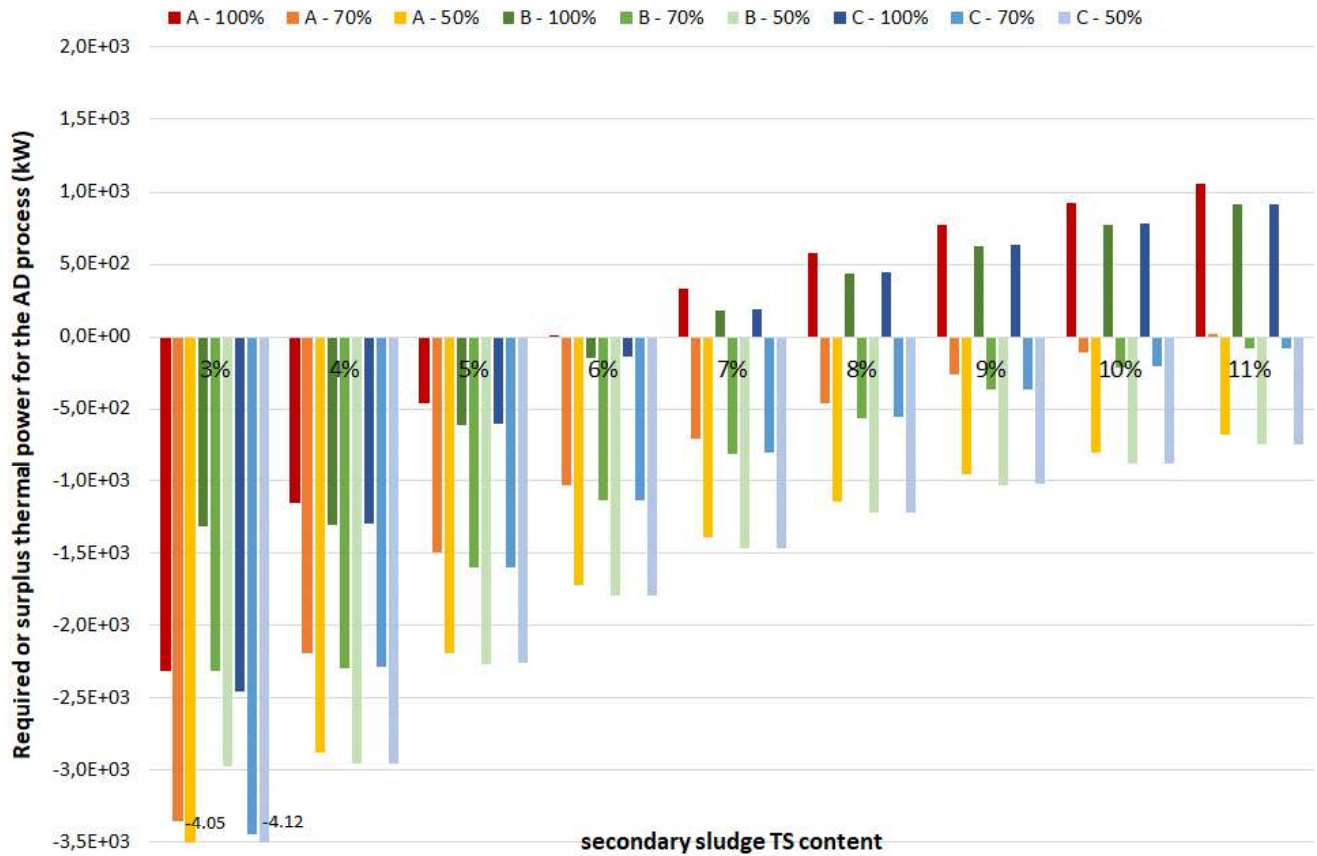


Figure 10. Required or surplus thermal power for the AD process for Scenario 2 that results from the technical assessment. Letters A, B and C refer to HRTs of the first-stage digestion of 5, 10 and 15 days, respectively. Percentage values (100% - 70% - 50%) refer to the efficiency with which the heat produced from the CHP unit was transferred to the processes that need heating (IHTs and digestion processes).

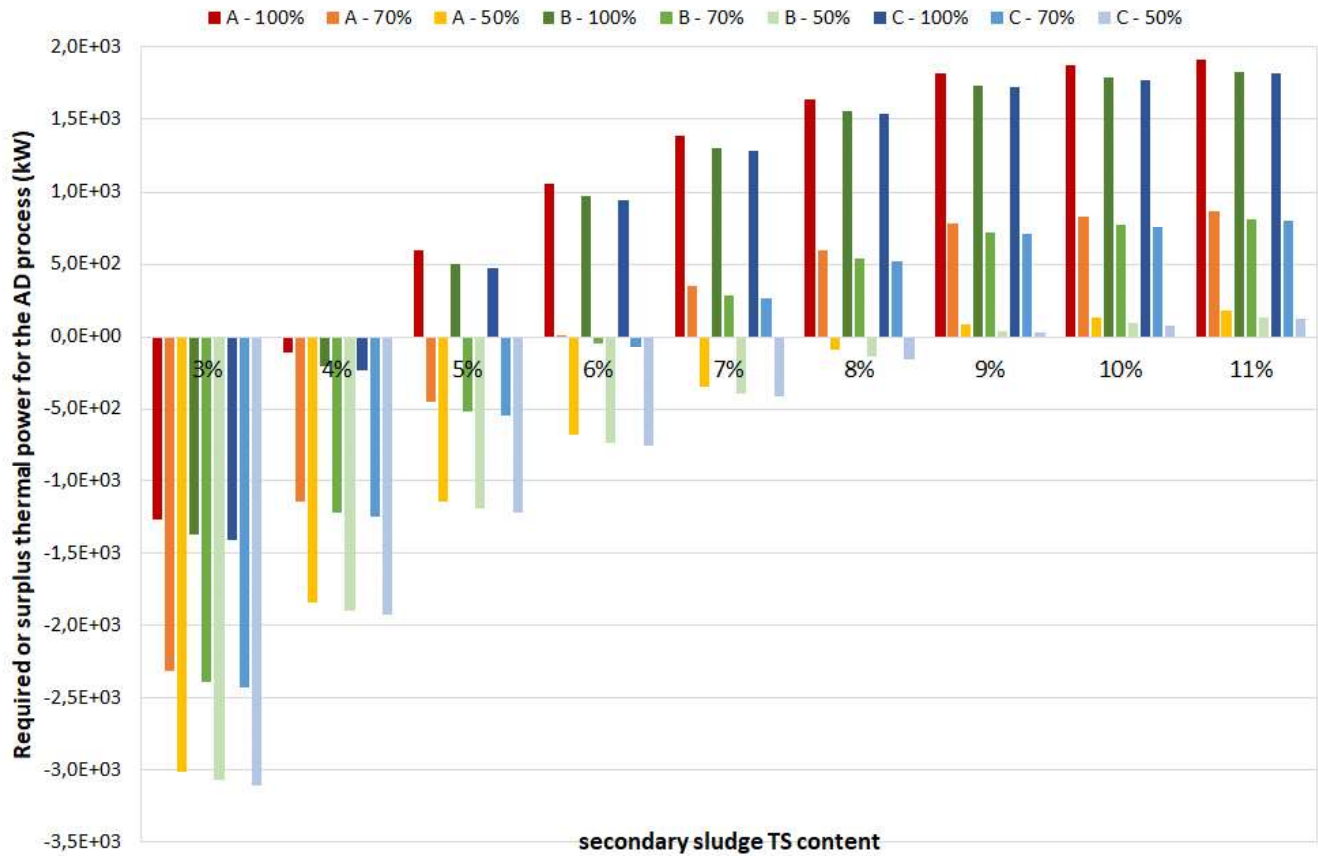


Figure 11. Required or surplus thermal power for the AD process for Scenario 3 that results from the technical assessment. Letters A, B and C refer to HRTs of the first-stage digestion of 5, 10 and 15 days, respectively. Percentage values (100% - 70% - 50%) refer to the efficiency with which the heat produced from the CHP unit was transferred to the processes that need heating (IHTs and digestion processes).

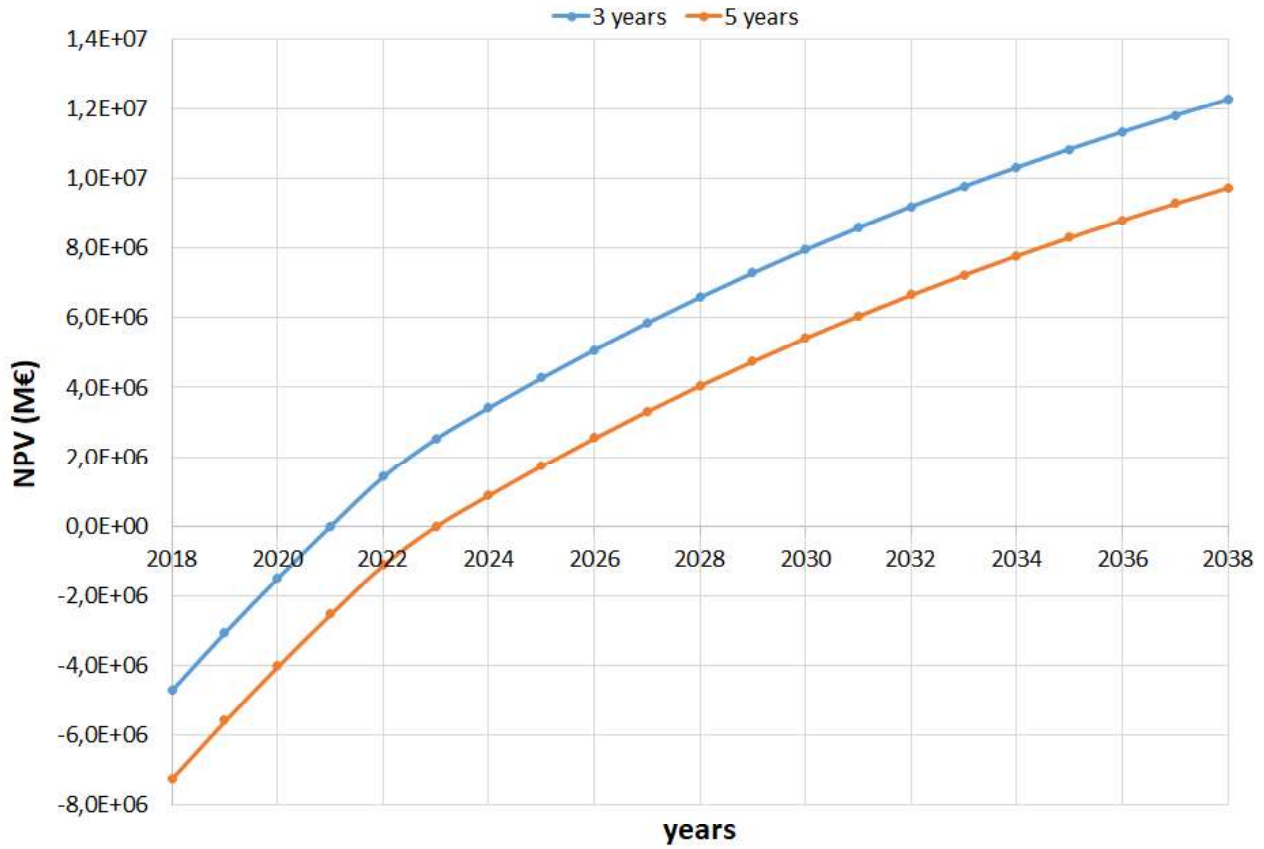


Figure 12. Net Present Value analysis for the assessment of the initial investment that the WWTP can bear, with pay-back times of three and five years, to renew the sludge line

Table S1. Main results of the technical assessment carried out for Scenario 1 (Primary sludge TS content 7%)

First-stage digestion N. of digesters	Second-stage digestion N. of digesters	Primary sludge digestion N. of digesters	Total number of digesters	Heat exchange efficiency	%TS secondary sludge (*)	Duration of the first and second stage digestion (days)
1	2	2	5	100%	4.12	5+15
1	1	2	4		4.31	10+10
2	1	2	5		4.30	15+5
1	1	2	4	70%	5.95	5+15
1	1	2	4		6.23	10+10
1	1	2	4		6.21	15+5
1	1	2	4	50%	8.46	5+15
1	1	2	4		8.86	10+10
1	1	2	4		8.84	15+5

(\*) that made the thermal balance neutral

Table S2. Main results of the technical assessment carried out for Scenario 2 (Primary sludge TS content 7%)

First-stage digestion N. of digesters	Second-stage digestion N. of digesters	Primary sludge digestion N. of digesters	Total number of digesters	Heat exchange efficiency	%TS secondary sludge (*)	Duration of the first and second stage digestion (days)
1	1	2	4	100%	6.00	5+15
1	1	2	4		6.41	10+10
1	1	2	4		6.38	15+5
1	1	2	4	70%	10.9	5+15
1	1	2	4		11.8	10+10
1	1	2	4		11.8	15+5
1	1	2	4	50%	23.7	5+15
1	1	2	4		27.1	10+10
1	1	2	4		26.9	15+5

(\*) that made the thermal balance neutral

Table S3. Main results of the technical assessment carried out for Scenario 3 (Primary sludge TS content 7%)

First-stage digestion N. of digesters	Second-stage digestion (**) N. of digesters	Primary sludge digestion N. of digesters	Total number of digesters	Heat exchange efficiency	%TS secondary sludge (*)	Duration of the first and second stage digestion (days)
1	4	-	5	100%	4.13	5+20
1	4	-	5		4.24	10+20
2	4	-	6	70%	4.28	15+20
1	4	-	5		5.96	5+20
1	4	-	5	50%	6.13	10+20
1	4	-	5		6.19	15+20
1	3	-	4	50%	8.47	5+20
1	3	-	4		8.72	10+20
1	3	-	4		8.80	15+20

(\*) that made the thermal balance neutral

(\*\*) digestion of the mixture made of primary sludge and treated digestate from secondary sludge