

Ecohydrology of street trees: design and irrigation requirements for sustainable water use

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Review

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3 **Ecohydrology of street trees: design and irrigation**
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6 **requirements for sustainable water use**
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28 (Dated: December 5, 2012)

Abstract

While the beneficial effects of urban vegetation have long been recognized, growing conditions in urban environments, especially for street trees, are typically harsh and limited by low water availability. Supplemental irrigation may be used to preserve aesthetic quality and ability to provide ecosystem services of urban vegetation but requires careful management of available economic and water resources to reduce urban water footprint. To this purpose, decision-makers need quantitative tools, requiring few, physically-based parameters and accounting for the uncertainties and future scenarios of the hydroclimatic forcing. Focusing on in-row and isolated trees, a minimalist description of street tree water balance is proposed here, including rainfed and irrigated conditions, and explicitly accounting for tree water requirements, growing conditions (in terms of soil properties and extension of bare soil, permeable and impervious pavements surrounding the tree), and rainfall unpredictability. The proposed model allows the quantification of tree cooling capacity, water stress occurrence, and irrigation requirements, as a function of soil, plant, and climate characteristics, thus providing indications regarding the tree ability to provide ecosystem services and management costs. In particular, an analysis of different planting designs suggests that a balanced design consisting in bare soil and permeable pavement with size equal to the lateral canopy extension is optimal for water conservation, tree cooling capacity, and health. The proposed model provides useful indications towards the definition of site-specific guidelines for species selection and planting design, for sustainable urban vegetation.

KEY WORDS: urban vegetation; street trees; soil moisture; stochastic rainfall; plant water stress; irrigation

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29 I. INTRODUCTION

30 While at the beginning of the 20th century only 14% of the world population lived in
31 urban settings, this percentage is now 50% (Konijnendijk, 2000), and it is predicted that over
32 5 billion people will reside in metropolitan areas by the year 2030 (United Nations, 2005;
33 Young, 2010). Increasing urban population causes the conversion of large parts of natural
34 landscape to urban environments, with significant repercussions on local climate, regional
35 hydrological cycle, as well as habitat and biodiversity presence (Kalnay and Cai, 2003;
36 Rees and Wackernagel, 2008). Within the urban environment, vegetation plays important
37 social, cultural, economic, and environmental roles, ranging from positive effects on human
38 health and improved social dynamics, increased housing prices and business district activity
39 (Jorgensen and Gobster, 2010; Kuo and Sullivan, 2001; Maas *et al.*, 2006; Payton *et al.*,
40 2008; Wolf, 2005), to beneficial environmental impacts such as reduced runoff, improved soil
41 drainage, soil erosion control, watershed protection, and provision of wildlife habitats and
42 ecological corridors (Fernandez-Juricic, 2000; Xiao and McPherson, 2002). Moreover, when
43 managed properly, urban vegetation provides local ecosystem services such as urban heat
44 island mitigation, cooling and reduction of energy demand in adjacent buildings (Imhoff
45 *et al.*, 2010; Shashua-Bar *et al.*, 2009), and alleviation of air pollution and dust (Beckett
46 *et al.*, 1998; McPherson *et al.*, 2011; Nowak *et al.*, 2006).

47 Despite the local variations in composition, pattern, and spatial extent of the urban
48 landscape (Quattrochi and Ridd, 1998; Thorsson *et al.*, 2011), urban vegetation is gener-
49 ally subject to biophysical and ecological conditions that are radically different from the
50 surrounding rural and natural environments, in particular regarding soil features and lo-
51 cal climate (Coder, 1996; Dwyer *et al.*, 1992, 2002; Gill *et al.*, 2007; Home *et al.*, 2010;
52 Konijnendijk, 2000; Lohr *et al.*, 2004; Swanwick *et al.*, 2003). Growth conditions are even
53 more severe for isolated trees located in parking lots and in-row along streets, and soils are
54 often characterized by high compaction levels and surface crusting, limiting water infiltra-
55 tion, drainage capacity, and oxygenation (Craul, 1999; Pauleit, 2003). Contamination by
56 anthropogenic materials (e.g., calcium from nearby construction weathering and de-icing
57 compounds) may further negatively impact soil quality (Pauleit, 2003). Finally, urban vege-
58 tation is subjected to the effects of dust and pollution (Takagi and Gyokusen, 2004) and may
59 need to withstand stringent pruning requirements for aesthetic reasons, as well as vandalism

60 and root injuries due to nearby construction (Foster and Blaine, 1978; Hauer *et al.*, 1994).

61 Maintaining a viable urban vegetation requires significant resources (economic resources
62 for purchasing, planting, and maintenance of plants; supply of fertilizers and water for
63 irrigation). Thus decision makers are faced by the complex problem of evaluating the trade-
64 offs between the benefits of urban vegetation and the related costs, towards sustainable urban
65 tree design and management strategies (Clark *et al.*, 1997; Dwyer *et al.*, 2003; Ferrini and
66 Fini, 2011). In particular, species selection and planting design are key steps to facilitate
67 subsequent management and to enhance tree life span. Historically, species selection has
68 been mainly driven by aesthetic criteria (e.g., tree architectural features), often resulting in
69 the choice of non-native species, likely ill-adapted to the local climatic conditions (Balling
70 *et al.*, 2008) and potentially invasive (Niinemets and Penuelas, 2008). As a result, tree life
71 span in urban areas tends to be significantly reduced with respect to nearby rural areas
72 (Berrang *et al.*, 1985; Foster and Blaine, 1978; Nowak *et al.*, 1990). A more sustainable
73 species selection needs to represent a compromise between aesthetic appeal and functional
74 aspects and tolerance to the harsh conditions typical of urban sites (Pauleit, 2003; Richards,
75 1983; Sæbø *et al.*, 2003).

76 Among the limitations imposed by the urban environment, water deficit is generally
77 recognized as the principal limiting factor controlling the growth of urban trees (Clark and
78 Kjelgren, 1990; Cregg, 1995), particularly when combined with high air temperature and
79 low air humidity and insufficient nutrient availability (Flückiger and Braun, 1999). The
80 combination of poor soil infiltration, scarcity of the permeable surfaces (often concentrated
81 in the immediate vicinities of the tree trunk; Fig. 1a), and enhanced atmospheric water
82 demand results in frequent and intense episodes of plant water stress, which are not limited to
83 arid and semi-arid climates (Whitlow *et al.*, 1992). Plant water stress may negatively impact
84 vegetation growth and aesthetic quality, but also limit the beneficial cooling associated to
85 plant transpiration because of extended stomatal closure (Bowler *et al.*, 2010; Chen *et al.*,
86 2011; Jenerette *et al.*, 2011; Kjelgren and Clark, 1993; Porporato *et al.*, 2001). Furthermore,
87 water shortage interferes with plant defense mechanisms, increasing the predisposition to
88 parasite and pathogenic fungi attacks and tree mortality in general (Cregg and Dix, 2001;
89 Flückiger and Braun, 1999), with catastrophic losses in case of low species diversity, that
90 is typical of some but not all urban environments (Raupp *et al.*, 2006; Sjöman *et al.*, 2012;
91 Walker *et al.*, 2009). Hence, under specific climatic conditions, tree water requirements

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3 92 and planting design, supplemental irrigation may be a necessity to sustain transpiration and
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5 93 hence the beneficial effects of urban vegetation. Currently, water needs for public and private
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7 94 landscape represent 40-70% of total municipal requirements (Hilaire *et al.*, 2008). Such high
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9 95 water requirements are partly explained by past inadequate species selection and by poor
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11 96 planting design, but also by water applications often exceeding plant demands (Balling
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13 97 *et al.*, 2008; Salvador *et al.*, 2011). In light of recently reported water shortages (Jenerette
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15 98 and Larsen, 2006), enhanced governmental restrictions on agricultural and municipal water
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17 99 use (Brennan *et al.*, 2007; MacDonald *et al.*, 2010), and the projected climate change and
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19 100 increase in urban population, quantitative tools are needed to address the 'urban water
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21 101 challenge' (Pataki *et al.*, 2011a). Specifically, decision-makers increasingly require tools for
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23 102 optimal species selection and planting design (Sæbø *et al.*, 2003; Sjöman and Nielsen, 2010),
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25 103 to effectively manage available resources and limit city water footprints, particularly in
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27 104 semi-arid regions.

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29 105 The specificities of the urban environment make it difficult to exploit existing ecohydro-
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31 106 logical knowledge relative to natural and managed rural ecosystems. Furthermore, while
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33 107 some data have been published on irrigation requirements of container-grown ornamen-
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35 108 tal plants under nursery conditions (Drunasky and Struve, 2005; Hagishima *et al.*, 2007),
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37 109 data relative to water requirements of mature urban trees are still scarce (Pataki *et al.*,
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39 110 2011b; Roberts and Schnipke, 1994). More importantly, as typical of other ecohydrological
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41 111 problems, the question of sustainable water management is complicated by the inherent
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43 112 intermittency and unpredictability of rainfall occurrence (Rodriguez-Iturbe and Porporato,
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45 113 2004) and by the projected shifts in rainfall patterns in the next decades, which render the
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47 114 available historical climatological data insufficient for an effective long-term planning of wa-
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49 115 ter use. The few existing models describing soil water availability to urban trees are based
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51 116 on yearly-averaged rainfall input (e.g., Lindsey and Bassuk (1991)) or driven by relatively
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53 117 short meteorological observations (e.g., DeGaetano and Hudson (2000)), thus poorly char-
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55 118 acterizing extreme events, such as long dry spells. In what follows, focusing on the case of
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57 119 street trees, we propose an alternative approach explicitly including rainfall unpredictability
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59 120 by means of a probabilistic description of rainfall occurrence, thus avoiding computationally
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121 heavy simulations that needs to be forced by multi-decadal rainfall time series to include
122 extreme events. The proposed approach can be also applied for climate change scenario
123 analyses, for which only qualitative indications on projected changes are available at best.

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4 124 Our minimalist model, based on the probabilistic description of soil moisture and irrigation
5 125 (Porporato *et al.*, 2004; Rodriguez-Iturbe and Porporato, 2004; Vico and Porporato, 2010,
6 126 2011a), provides a quantitative tool to assess plant water status, effective cooling capacity,
7 127 and irrigation requirements, as a function of species selection and tree size (in terms of plant
8 128 water requirements), planting design (in terms of extension of permeable and impervious
9 129 surfaces around the tree trunk), rainfall patterns, and implemented irrigation strategy. This
10 130 model provides quantitative indications in support of strategic decision making for adequate
11 131 species selection, planting design, and management practices to maintain urban vegetation
12 132 ecosystem services while limiting water requirements, under current and future precipitation
13 133 patterns.

134 II. MODEL DESCRIPTION

135 A. Planting geometry of isolated and in-row street trees

136 Within the variety of growing conditions of urban vegetation, we focus here on isolated
137 or in-row trees growing in parking lots or along streets. In general, around these trees, it
138 is possible to distinguish (up to) three areas with different permeability properties (Fig.
139 1a), which in turn impact the soil water balance of the tree rooting volume: i) an area of
140 bare soil, A_B , often located immediately around the tree or shrub trunk; the infiltration
141 capacity of the bare soil is determined by soil permeability, $\eta_B \leq 1$ (depending on soil
142 properties, such as crusting, hydrophobicity, level of soil compaction, and mulching), and by
143 soil saturation; ii) a partially permeable area, A_P , which allows the infiltration of a fraction
144 $\eta_P < 1$ of the incoming rainfall; this area may be covered by tree grates or permeable
145 pavement (e.g., interlocking concrete permeable pavement); and iii) an area of impervious
146 pavement, A_I , which completely prevents water infiltration in the soil beneath, but that may
147 generate a runoff towards the more permeable areas if adequately sloped and designed (i.e.,
148 in absence of curbs preventing water flow); the fraction of rainfall on the impervious surface
149 that may potentially infiltrate in the more permeable areas is defined by the coefficient
150 $\eta_I < 1$. Pervious concrete (similar to standard concrete but lacking the fine aggregates)
151 is here assimilated to impervious pavements, on the basis of recent experimental results
152 suggesting insignificant differences between pervious and impervious paving (Morgenroth

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3 and Buchan, 2009; Viswanathan *et al.*, 2011). Fig. 1a shows a few examples of planting
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5 design: all the three above mentioned regions are apparent in III. In other locations, the bare
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7 soil may be reduced to a minimum (I), or the permeable pavement area may be altogether
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9 absent (VI), or a curb may prevent the free flowing of water from the impervious pavement
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11 to the permeable areas (IV). The extreme case of absence of permeable and impervious
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13 pavements, e.g., a tree located in a wide lawn (V) is equivalent to the case of an isolated
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15 tree in a natural environment. The lateral extensions of both canopy and root zone constitute
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17 further geometric constraints. To account for them, we define A_R as the area over which
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19 the root system extends horizontally and A_C the projected area of the canopy. The latter
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21 is relevant for the tree water balance because the vegetation canopy, in particular when leaf
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23 area index is high, may partially intercept rainfall, thus reducing the amount of water that
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25 can potentially infiltrate in the permeable areas or create a beneficial runoff from the nearby
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27 impervious surfaces. It is also useful to define the total area that contributes to the soil
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29 water balance pertaining a single tree, i.e., $A_T = A_B + A_P + A_I$.

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31 As apparent in Fig. 1a), the specific geometry of the areas surrounding the tree is highly
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33 variable, in compliance with aesthetic and practical reasons. Spacing between adjacent trees
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35 often represents a compromise among providing adequate soil volumes and water availability
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37 (DeGaetano and Hudson, 2000), achieving the required ecosystem services (aesthetic quality,
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39 air cooling, and pollution reduction), and preserving the ability to exploit the area under-
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41 neath for foot or vehicular traffic (e.g., McPherson (2001)). In the following quantitative
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43 analyses, we will focus on the case of circular symmetry, which works best for isolated trees.
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45 The radii r_i fully define the areas affecting the tree soil water balance, $A_i = \pi r_i^2$, where
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47 the subscript i may refer to bare soil ($i = B$), permeable ($i = P$) or impervious ($i = I$)
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49 pavements, canopy extension ($i = C$) or rooting zone ($i = R$). The geometry of the problem
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51 for this specific case is represented in Fig. 1b. The obtained results can be easily extended
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53 to other geometries, such as the squares employed in several locations (Fig. 1a).

54 55 56 57 58 59 60 179 **B. Soil moisture balance over tree rooting volume**

180 A previously proposed stochastic model of the soil water balance suitable for natural and
181 agricultural environments (see e.g., Laio *et al.* (2001); Rodriguez-Iturbe *et al.* (1999); Vico
182 and Porporato (2011a)) is here adapted to the case of isolated or in-row trees. The temporal

183 dynamics of the soil water content in the plant rooting zone can be effectively described by
 184 the following water balance

$$nA_R Z_R \frac{ds(t)}{dt} = R(t) + I(s(t)) - ET(s(t)) - LQ(s(t)). \quad (1)$$

185 The state variable $s(t)$ represents the relative soil moisture averaged over the soil volume
 186 $A_R Z_R$, where most of the plant roots are located and over which soil features are assumed
 187 uniform, with Z_R being the characteristic rooting depth, A_R the area over which the root
 188 system extends (see Fig. 1b), and n the soil porosity. The main input to the soil water
 189 balance, $R(t)$, is represented by rainfall, either directly falling over the permeable area or
 190 falling over nearby areas and being brought over by runoff. Water may also be supplied by
 191 irrigation applications, $I(s(t))$. The main losses occur through soil water evaporation and
 192 plant transpiration, $ET(s(t))$, runoff and deep percolation, $LQ(s(t))$. It is assumed that
 193 there is no interaction between the root volume and any existing water table. All the fluxes
 194 in Eq. (1) are interpreted at the daily time scale.

195 The actual volumetric input by rainfall to the rooting zone depends on the interaction
 196 among surface permeabilities (and runoff generating capacity), canopy, and root lateral
 197 extension. Regarding the impact of canopy, experimental evidence suggests that canopy
 198 interception may reduce both the frequency of effective (i.e., non-canopy intercepted) rainfall
 199 events and their effective depths (Daly *et al.*, 2008; Guswa, 2005). To quantify the volume of
 200 potentially infiltrating rainfall it is thus necessary to consider the geometry of the problem,
 201 by accounting for the fraction of area subjected to canopy interception effect, the extension
 202 and permeability of bare soil and permeable pavement, and the distance of the contributing
 203 permeable and impervious areas from the edge of the rooting zone, r_{nR} . Depending on root
 204 lateral extension, three cases need to be considered: i) the rooting zone extends under the
 205 entire permeable surface till the permeable/impervious surface interface (i.e., $r_R = r_B + r_P$),
 206 ii) the rooting zone does not extend under the entire bare soil and permeable area (i.e.,
 207 $r_R < r_B + r_P$), and iii) the rooting zone extends also under the impervious surface (i.e.,
 208 $r_R > r_B + r_P$ depicted in Fig. 1b). The occurrence of the latter case, i.e., the rooting zone
 209 extending also beyond the permeable/impervious surface interface, depends on soil type
 210 and compaction level, construction details, water availability and its location, and features
 211 of the nearby areas. The site-specificity of these features partly explains the contrasting
 212 conclusions from studies on root extension under impervious pavement, with stunted root

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213 growth in certain locations and relatively well developed, but concentrated, roots in others
214 (e.g., Reichwein (2003); Čermák *et al.* (2000)). Due to its complexity, the third case is
215 not considered here, i.e., it is assumed that roots do not generally extend under impervious
216 areas. For the purposes of defining the occurrence of water stress and irrigation requirements,
217 assuming that roots do not extend under the impervious pavement results in conservative
218 estimates of water stress frequency and severity, because the assumed smaller rooting zone
219 has lower buffering capacity against water dynamics and does not allow the exploitation of
220 other water stores that might be available with more extensive rooting systems.

221 In the first case (i.e., $r_R = r_B + r_P$), all the infiltrated water from the permeable area A_P
222 and a fraction η_I of surface runoff generated by the nearby impervious surface A_I contribute
223 to the rooting zone soil water content. For a generic rainfall event of depth $h(t)$, the water
224 volume contributing to the soil water balance (1) can be quantified as

$$R(t) = (A_T^{(\eta)} - A_T^{(\eta^k)}) h(t) + A_T^{(\eta^k)} h'(t) \quad (2)$$

225 where $A_T^{(\eta)} = \sum_{i=B,P,I} \eta_i A_i$ and $A_T^{(\eta^k)} = \sum_{i=B,P,I} \eta_i k_{i,C} A_i$; $h(t)$ is the rainfall event depth,
226 and $h'(t)$ is the effective rainfall depth below the canopy (i.e., after canopy interception;
227 see II.C below). The coefficients $k_{i,C}$ (with $i = B, P, I$) are the fractions of the bare soil,
228 permeable, and impervious areas respectively influenced by the presence of the canopy, while
229 η_i are the respective surface permeability, driving the fraction of rainfall volume infiltrating
230 in non-saturated soils. In this case, lateral water redistribution between the root volume and
231 the nearby soil is neglected, even though it may contribute to root volume water depletion,
232 unless artificial boundaries are present.

233 In the second case, where the lateral extension of roots is less than the bare soil and
234 permeable pavement combined areas (i.e., $r_R < r_B + r_P$, as in the case depicted in Fig.
235 1b), the infiltration from the excess permeable surface and the runoff generated by the
236 surrounding impervious pavement does not directly contribute to soil water content of the
237 rooting zone, but rather it enhances water content outside the rooting volume. In absence
238 of artificial boundaries, soil water beyond the rooting volume may be laterally redistributed
239 according to existing soil water potential gradients. While a precise description of the soil
240 water lateral redistribution lies beyond the scope of the proposed model, it is simply assumed
241 here that only a fraction of the water volume infiltrating over a ring of width dr at a generic
242 distance r from the edge of the rooting zone will finally contribute to the tree available

243 water. The contributed fraction is assumed to decrease exponentially with the distance
 244 from the rooting zone edge. Thus, the contribution of the infinitesimal ring of permeable
 245 surface is $dR_{P,nR}(t) = 2\pi\eta_P h(t)r_{nR}e^{-r}dr$, where the rainfall depth $h(t)$ is substituted by $h'(t)$
 246 should the infinitesimal area $2\pi r_{nR}dr_{nR}$ be subject to canopy interception. With the further
 247 assumption to simplify the notation that the areas beyond the rooting zones are not subject
 248 to canopy interception (as discussed below, canopy seldom extends beyond the rooting zone),
 249 the total contribution of the permeable area beyond the rooting extension is $R_{P,nR}(t) =$
 250 $\int_{r_R}^{r_B+r_P} dR_{P,nR} = 2\pi\eta_P h(t) [1 + r_R - (1 + r_B + r_P)e^{-r_{nR}}]$, where $r_{nR} = r_B + r_P - r_R$ is the
 251 lateral extension of the area with permeable pavement beyond the rooting zone (Fig. 1b).
 252 Similarly, the water volume contributed by the nearby impervious surface to plant accessible
 253 soil moisture is here considered exponentially decreasing with increasing distance between
 254 the edge of the rooting area and the position of the permeable/impervious surface interface,
 255 i.e. $R_I(t) = e^{-r_{nR}}\eta_I h(t)A_I$ (where $k_{I,C}$ has been set to zero under the assumption that
 256 $r_C \leq r_R$; see below). Accordingly, the water volume contributed to the soil water balance
 257 for $r_R < r_B + r_P$ from a generic event of depth $h(t)$ is

$$R(t) = h'(t) \sum_{i=B,P} \eta_i k_{i,C} A_i + h(t) \left\{ 2\pi\eta_P [1 + r_R - (1 + r_B + r_P)e^{-r_{nR}}] + \eta_I A_I e^{-r_{nR}} \right\}. \quad (3)$$

258 Regarding the losses, the individual plant is responsible for a daily volumetric water
 259 uptake, which in general depends on species, amount of transpiring leaves (and hence tree
 260 size), plant activity (driven by temperature, solar radiation, and plant water status, in turn
 261 function of soil moisture), plant general conditions (e.g., impact of pollutants, diseases,
 262 and pest infestations), and atmospheric water demand (as defined by air temperature and
 263 humidity, and wind speed). As such, water uptake accounts for the specificities of the urban
 264 growing environment, including potentially higher temperatures and vapor pressure deficits
 265 (Kjelgren and Clark, 1993; Litvak *et al.*, 2012; McCarthy and Pataki, 2010; Wang *et al.*,
 266 2011). Losses through soil water evaporation are driven by soil moisture in the superficial
 267 layers of the bare soil area and, to a lesser extent, under the permeable pavement. Because of
 268 the geometry typical of street trees, often implying relatively large trees growing on a rather
 269 small areas of bare or mulched soil, soil water evaporation is generally much less relevant
 270 than plant transpiration. Thus, in what follows, we focus on losses by plant transpiration.
 271 While most of the results presented below are valid for a generic transpiration function

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272 $ET(s(t))$, for the quantitative results below a piecewise linear dependence of transpiration
273 water volume on soil moisture is assumed (Rodriguez-Iturbe *et al.*, 1999), i.e.,

$$\rho_w ET(s(t)) = \begin{cases} T_{max} \frac{s(t)}{s^*} & s(t) < s^* \\ T_{max} & s(t) \geq s^* \end{cases}, \quad (4)$$

274 where T_{max} represents the mass of transpired water per tree per day when the plant is under
275 well watered conditions (depending on species and amount of transpiring leaves), ρ_w is the
276 density of water, and s^* is the soil moisture level corresponding to incipient plant water
277 stress (i.e., below which plant transpiration is reduced because of stomatal closure). In
278 what follows, we will often refer to the volumetric water losses per unit rooting volume, i.e.,
279 $\rho(s(t)) = (nZ_R A_R)^{-1} ET(s(t))$.

280 The other loss term included in the soil water balance (1), $LQ(s(t))$, combines losses
281 through surface runoff and deep percolation from the bottom of the rooting volume. For
282 simplicity, following Milly (2001) and Porporato *et al.* (2004), it is assumed that deep perco-
283 lation and runoff take place instantaneously (at the daily time scale) whenever soil moisture
284 reaches a threshold s_1 , typically slightly above soil field capacity. For soils within closed-
285 bottom containers or other confined spaces, the threshold s_1 may approach soil saturation,
286 to mimic the poor drainage typical of these growing conditions.

287 Finally, depending on tree water requirements, rainfall input, and landscaping strategy,
288 an irrigation system may be implemented. This additional water input is included in the
289 modeling scheme as detailed in Vico and Porporato (2010, 2011a). In particular, as opposed
290 to fixed-schedule water applications, we consider the case of demand-based irrigation, where
291 irrigation applications are triggered by soil moisture reaching a pre-set 'intervention point'
292 \tilde{s} (Vico and Porporato, 2011a). If a certain water stress is considered tolerable, the inter-
293 vention point can be set below the incipient stomatal closure s^* , thus performing a deficit
294 irrigation (English and Raja, 1996). Currently deficit irrigation of urban vegetation is often
295 applied as the result of municipal level water efficiency ordinances and limited technical or
296 economic resources, rather than in response to environmental concerns (Parés-Franzi *et al.*,
297 2006). Several deficit irrigation applications are under active consideration (Delcambre and
298 Rossignol, 1999; Shoostarian *et al.*, 2011; Suleiman *et al.*, 2011) and increasingly limited
299 water availability will likely force the adoption of new guidelines for species selection and
300 management, favoring species for which a limited water stress does not significantly impact

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4 301 the tree aesthetic quality. Furthermore, depending on the employed irrigation technique,
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6 302 we distinguish between i) a more traditional irrigation, in which each irrigation application
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8 303 will provide instantaneously (at the daily time scale) a given amount of water, that restores
9
10 304 soil moisture to a pre-set 'target' level, \hat{s} ; each irrigation application provides a volume
11
12 305 $nZ_R A_R(\tilde{s} - \hat{s})$; and ii) the more sophisticated micro-irrigation, which is idealized here as a
13
14 306 continuous supply of water that balances losses through evapotranspiration (i.e., providing
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16 307 a volume $ET(\tilde{s})$ per day), initiated when the soil moisture reaches the intervention point,
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18 308 thus maintaining soil moisture at the intervention point till the next (effective) rainfall event
19
20 309 (Vico and Porporato, 2010). In an urban setting, the first strategy may correspond to rather
21
22 310 labor-intensive activities, such as periodic water applications through plant water bags (Fig.
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24 311 1a, V) or direct manual watering with hoses or trucked water. Conversely, micro-irrigation
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26 312 requires the installation of a permanent irrigation system (e.g., sub-irrigation and drip ir-
27
28 313 rigation systems), allowing more frequent or even continuous water applications. As such,
29
30 314 micro-irrigation is currently limited to specific locations where economic resources and in-
31
32 315 frastructures are available, and there is the need to maintain certain vegetation, e.g., for
33
34 316 touristic reasons.

317 C. Inclusion of rainfall stochasticity

318
319 Rainfall unpredictability can be explicitly included in the above soil water balance by
320
321 idealizing rainfall occurrence as a series of instantaneous events occurring according to a
322
323 marked Poisson process, with average frequency λ . Rainfall event depths are assumed to be
324
325 exponentially distributed, with average depth α (Rodriguez-Iturbe *et al.*, 1999; Rodriguez-
326
327 Iturbe and Porporato, 2004). Within this framework, the effective rainfall depth under the
328
329 canopy can be well described as a censored Poisson process, occurring according to frequency
330
331 $\lambda' = \lambda e^{-\Delta/\alpha}$, where Δ is a vegetation-dependent depth-threshold (Rodriguez-Iturbe *et al.*,
332
333 1999). Rainfall events smaller than Δ are completely intercepted by the canopy. We further
334
335 assume that the mean effective rainfall depth is reduced to $\alpha' = \kappa_C \alpha$ (Daly *et al.*, 2008).
336
337 The presence of the canopy (and the existence of an interception threshold Δ) generates two,
338
339 partially dependent, Poisson processes: i) the uncensored rainfall process providing water to
340
341 areas unaffected by the canopy, which occurs with mean frequency λ , and ii) the censored
342
343 Poisson process driving precipitation under the canopy, which occurs with mean frequency

331 λ' . Considering in a rigorous way both processes would undermine the analytical tractability
 332 of the whole problem. As an approximation, effective rainfall is assumed to reach the ground
 333 with average frequency λ_{eff} , representing the area-weighted average of λ and λ' , i.e.,

$$\lambda_{\text{eff}} = \left(1 - \frac{A_T^{(k)}}{A_T}\right) \lambda + \frac{A_T^{(k)}}{A_T} \lambda' \quad (5)$$

334 where $A_T^{(k)} = \sum_{i=B,P,I} k_{i,C} A_i$. This approximation works particularly well when the
 335 vegetation-dependent threshold Δ is small with respect to the average event depth α (so
 336 that λ' does not significantly differ from λ), a relatively common case even in presence of
 337 large canopies (Daly *et al.*, 2008; Guswa, 2005).

338 The effective rainfall contributing to the soil water balance occurs according to a modified,
 339 censored Poisson process, with frequency λ_{eff} , and providing water volumes extracted by an
 340 exponential distribution with average volume α_V , obtained by setting $h = \alpha$ and $h' = \kappa_C \alpha$
 341 in Eqs. (2) and (3) for $r_R = r_B + r_P$ and $r_R < r_B + r_P$ respectively. The effective depth
 342 contributing to the soil water content in the rooting zone is given by

$$\alpha_{\text{eff}} = \alpha_V / A_R. \quad (6)$$

343 D. Soil moisture probability density function (pdf) and irrigation requirements

344 With the above simplifications and assuming stochastic steady state, it is possible to
 345 obtain analytically the soil moisture probability density function (pdf), $p(s)$, both in absence
 346 of irrigation and with a generic demand-based irrigation scheme, by exploiting the crossing
 347 properties of the soil moisture process. In fact, after the soil moisture process has reached
 348 the stochastic steady state (i.e., $\partial p(s)/\partial t = 0$), the frequency of upcrossing of a generic soil
 349 moisture threshold must equal the frequency of downcrossing of the same threshold. For a
 350 generic normalized loss function $\rho(s) = (nZ_R A_R)^{-1} ET(s)$ and including irrigation, the soil
 351 moisture pdf reads (Vico and Porporato, 2011a) is

$$p(s) = C \frac{e^{\int_{\tilde{s}}^s \left(\gamma - \frac{\lambda_{\text{eff}}}{\rho(u)}\right) du}}{\rho(s)} \left\{ 1 + \int_{\tilde{s}}^s [\gamma \theta(\hat{s} - u) - \delta(\hat{s} - u)] e^{\int_{\tilde{s}}^u \left(\gamma - \frac{\lambda_{\text{eff}}}{\rho(y)}\right) dy} du \right\}, \quad (7)$$

352 where $\gamma = nZ_R / \alpha_{\text{eff}}$, $\theta(\cdot)$ is the Heaviside function, and $\delta(\cdot)$ the Dirac delta function. The
 353 normalization constant C can be obtained by imposing $\int_{\tilde{s}}^{s_1} p(s) ds = 1$. For $\tilde{s} \rightarrow 0$ and $\hat{s} \rightarrow 0$,

the above pdf simplifies to the case of absence of irrigation. For $\hat{s} \rightarrow \tilde{s}$, the case of micro-irrigation is retrieved, even though a more straightforward derivation of the soil moisture pdf for micro-irrigation is also available (as detailed in Vico and Porporato (2010)). Eq. (7) can be easily particularized for the piecewise linear loss function in (4), which is used for the quantitative analyses below.

The crossing properties of the soil moisture process can also be exploited to obtain the average irrigation requirements in terms of irrigation frequency and required water volumes. Following Vico and Porporato (2010, 2011a), the average frequency of the irrigation treatment is the frequency of downcrossing of the threshold $\xi = \tilde{s}$, i.e., $\nu^\downarrow(\tilde{s}) = \rho(\tilde{s})p(\tilde{s})$, while the volume of irrigation water applied over a period of duration T_{seas} is given by the amount of water provided at each application times the number of applications over the period, i.e.,

$$V_t = nZ_R A_R(\hat{s} - \tilde{s})\nu^\downarrow(\tilde{s})T_{seas} = nZ_R A_R(\hat{s} - \tilde{s})\rho(\tilde{s})p(\tilde{s})T_{seas}. \quad (8)$$

E. Plant average transpiration and water stress

The above described stochastic framework allows also the quantification of plant average transpiration and the occurrence and severity of plant water stress, as a function of species, tree size, planting design, root zone features, and precipitation patterns, under unpredictable rainfall. Plant transpiration and water stress are key quantities to describe plant ability to provide ecosystem services: on the one hand, average transpiration over the season is a measure of the effective capacity of the tree to provide its potential cooling effect; on the other, water stress provides some indications regarding tree growth and aesthetic value as well as its health and susceptibility to pest attacks, even though the response to water stress is highly species-specific.

Average mass daily transpiration over the season can be obtained from the soil moisture pdf (7) as

$$\langle T \rangle = \int_{\tilde{s}}^{s_1} \rho_w ET(s)p(s)ds, \quad (9)$$

where $\rho_w ET(s)$ is defined in (4). The ratio $\langle T \rangle / T_{max}$ quantifies how the specific growing conditions (climate, planting geometry, irrigation) reduce the ability of the street tree to

379 provide cooling, with reference to the maximum potential cooling effect (proportional to
380 T_{max}).

381 Regarding water stress, to account for frequency, duration, and intensity of plant water
382 stress within a single indicator, we employ the 'dynamic water stress' or mean dynamic
383 stress over the growing season $\bar{\theta}$ (Porporato *et al.* (2001)):

$$\bar{\theta} = \begin{cases} \left(\frac{\bar{\zeta}' \bar{T}^\downarrow(s^*)}{k T_{seas}} \right) (\nu^\downarrow(s^*) T_{seas})^{-\frac{1}{2}} & \text{if } \bar{\zeta}' \bar{T}^\downarrow(s^*) < k T_{seas} \\ 1 & \text{otherwise} \end{cases} \quad (10)$$

384 In the above definition, $\bar{\zeta}'$ is the average static water stress, $\bar{T}^\downarrow(s^*)$ is the average time spent
385 by the soil moisture process below the threshold s^* , $\nu^\downarrow(s^*) T_{seas}$ is the average number of
386 downcrossings of the threshold s^* over the period T_{seas} , and k is an index of plant resistance
387 to water stress. The interested reader is referred to Porporato *et al.* (2001) for a discussion
388 on the rationale behind these stress measures. The frequency of downcrossing of level s^* ,
389 $\nu^\downarrow(s^*)$, is linked to the soil moisture pdf as $\nu^\downarrow(s^*) = \rho(s^*) p(s^*)$, while the average time spent
390 by the process below the same threshold, $\bar{T}^\downarrow(s^*)$, can be obtained as $\bar{T}^\downarrow(s^*) = \nu^\downarrow(s^*)^{-1} P(s^*)$,
391 where $P(\cdot)$ is the cumulative density function. In turn, the average static stress, $\bar{\zeta}'$, is defined
392 as mean level of plant water stress, provided that the plant is under stress, i.e.,

$$\bar{\zeta}' = P(s^*)^{-1} \bar{\zeta} = P(s^*)^{-1} \int_0^1 \zeta p_Z(\zeta) d\zeta, \quad (11)$$

393 where ζ depends on soil moisture as $\zeta(t) = \max\{s^{*-q} (s^* - s(t))^q, 0\}$, and $p_Z(\zeta)$ is the
394 probability density function of the static stress, obtained from $p(s)$ through the derived
395 distribution technique (see Porporato *et al.* (2001) for details). The parameter q is a measure
396 of the nonlinearity of the effects of soil moisture on plant status, with higher q for plants
397 more sensitive to a small change in water availability. While in principle this definition of
398 water stress can be employed both in absence and in presence of irrigation, we limit the
399 analyses of tree water stress to the case of absence of irrigation. In fact, the choice of the
400 irrigation strategy should be based on considerations relative to acceptable plant water stress
401 levels, thus making the quantification of water stress in irrigated settings less relevant.

402 F. Model parameterization

403 1. Tree size and water requirements

404 To fully characterize street tree water balance, information on tree level transpiration
405 rates, canopy and root extensions are needed. Because little information is currently avail-
406 able for mature urban trees (Pataki *et al.*, 2011b), parameterization of the above model
407 may require resorting to additional assumptions or to the combination of different sources
408 of information, as discussed next.

409 Regarding tree water requirements, leaf level transpiration rates for specific species/location
410 combinations can be obtained e.g. by means of gas exchange measurements, to quantify
411 stomatal conductance. Upscaling such leaf area transpiration rate to the canopy level re-
412 quires knowledge of the total tree transpiring leaf area (a function of tree size). For most
413 of currently available data on transpiration rates in urban settings, the only information
414 available on tree dimension is trunk diameter at breast height (DBH), with canopy exten-
415 sions being reported only in some cases. To circumvent such lack of information, existing
416 allometric relationship may be used to estimate canopy height and radius from DBH (see,
417 e.g., McHale *et al.* (2009)); alternatively, realistic assumptions are to be made on canopy
418 radius of the species under scrutiny. A selection of existing data on transpiration rate and
419 canopy extension relative to the most common species in North American cities, growing in
420 parks or along streets, is reported in Table I.

421 Regarding root dimensions, to our knowledge no dataset on plant transpiration includes
422 information about root extension, depth, and role played by the specific geometry of the
423 planting site. Thus the choice of related model parameters needs to rely on other indirect
424 information and assumptions. In absence of external constraints, root and canopy radial
425 extensions tend to be similar (Craul, 1985; Schenk and Jackson, 2002). Furthermore, as
426 discussed in II.B, roots extending under impervious surfaces tend to contribute little to tree
427 available soil water. Hence, analyses are limited to the case of roots not extending beyond the
428 permeable/impervious interface, a conservative assumption for the quantification of water
429 stress and irrigation requirements. As a result, in the following analyses, we assume that
430 roots fully exploit the soil area below the permeable surfaces, but do not effectively extend
431 beyond canopy area, nor below the impervious surface (i.e., $r_R = \min\{r_B + r_P, r_C\}$; Fig.

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4 432 1b). It can be expected that the average rooting depth, Z_R , is generally smaller in an urban
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6 433 setting than under natural conditions, because of either the negative effect of soil compaction
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8 434 and low soil aeration or existing physical constraints (compacted or otherwise inhospitable
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10 435 layers, closed-bottom containers). A direct consequence of these limitations to rooting depth
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12 436 is that deeper planting soils may not fully compensate for narrow planting designs (Craul,
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14 437 1985).

14 438 To explore the effect of species selection and tree size, a sensitivity analysis is conducted
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16 439 on the tree transpiration rate under well watered conditions (see III below). As hinted at
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18 440 above, for set climatic conditions (solar irradiance, air humidity and temperature), tran-
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20 441 spiration rate per tree is a function of tree species (via maximum stomatal conductance)
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22 442 and tree canopy size (via total leaf area). Literature data suggest that the variability of
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24 443 stomatal conductance is rather small across species belonging to the same functional group
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26 444 and adapted to similar climatic conditions (see, e.g., Körner *et al.* (1979) for a synthesis).
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28 445 Hence, in the sensitivity analysis on tree water requirements below, it is assumed that larger
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30 446 trees have a higher transpiration rates, thus providing indications both regarding species
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32 447 selection (via the typical size of mature trees) and the effect of tree growth over time. In
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34 448 absence of more detailed information, in what follows, we assume that total daily transpira-
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36 449 tion scales with tree canopy volume, which in turn (for an idealized spherical canopy), scales
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38 450 as r_C^3 . Hence, higher total transpiration corresponds to higher r_C ; in turn, r_C potentially af-
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40 451 fects root lateral extension (being $r_R = \min\{r_C, r_B + r_P\}$). Conversely, because of potential
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42 452 urban-specific constraints on root ability to extend downward (Grabosky *et al.*, 2001), it is
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44 453 assumed that tree dimension does not significantly influence average rooting depth (i.e., in
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46 454 all the analyses, Z_R is kept constant also when varying T_{max}).

47 455 2. Planting design

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50 456 In the following simulations, it is assumed that the plant trunk is located within an area
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52 457 of bare soil of radius $r_B = 1$ m. The radius of the area influencing the tree water balance
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54 458 (through either direct infiltration or runoff) is $r_T = r_B + r_P + r_I$, a value that depends mainly
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56 459 on planting design and existence of curbs (Fig. 1). For non-isolated trees and in absence
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58 460 of pavement features impeding water free flow, r_T represents the semi-distance between
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60 461 adjacent trees, which in turn is set by desired tree density or level of canopy cover and

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4 462 shade. We explore the impact of permeable pavement ring size r_P (and consequently the
5 463 size of the impervious ring, $r_I = r_T - (r_B + r_P)$), both assuming set tree density (i.e., for
6 464 set r_T) and altering the fraction of permeable vs. impervious surfaces around the isolated
7 465 tree thus allowing the distance between adjacent tree to vary as well. In the first case, a
8 466 distance between adjacent trees of 15 m (corresponding to $r_T = 7.5$ m) is used as an example.
9 467 This value is in good agreement with typical municipal guidelines on street tree planting
10 468 and well balances the needs to achieve an adequate canopy cover and to exploit the areas
11 469 underneath for other uses. Furthermore, it is assumed that the bare soil area A_B does not
12 470 present extensive crusting, so that $\eta_B = 1$. We consider a permeable pavement that allows
13 471 the infiltration of a fraction of rainfall $\eta_P = 0.45$, while the impervious pavement contributes
14 472 to the tree soil water balance with runoff representing a fraction $\eta_I = 0.1$ of the precipitated
15 473 water.

16 17 18 19 20 21 22 23 24 25 26 27 474 3. *Rainfall forcing*

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30 475 With the idealization of rainfall occurrence as a marked Poisson process, rainfall pattern
31 476 is fully characterized by the average event depth α and the average event frequency λ .
32 477 We focus to the summer period (May-September at intermediate latitudes in the Northern
33 478 hemisphere), when trees are fully active, temperatures and atmospheric water demands tend
34 479 to be high, and hence the risk of water stress is highest. Accordingly, rainfall parameters
35 480 α and λ are averages for the same period rather than for the entire year. For a specific
36 481 location, these parameters can be inferred from daily rainfall records. In section III, we
37 482 explore the effect of the predicted intensification of extreme rainfall events and increased
38 483 frequency of dry spells by climate change (see e.g. Easterling *et al.* (2000)), by decreasing λ
39 484 while increasing α so that the total seasonal rainfall $R_{tot} = T_{seas}\alpha\lambda$ is maintained constant.
40 485 Additional locations may be investigated by altering R_{tot} .

41 42 43 44 45 46 47 48 49 50 51 52 53 486 4. *Rainfall interception*

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56 487 Because rainfall interception by urban canopies has not been experimentally character-
57 488 ized, we set rainfall interception threshold Δ at 3 mm, consistently with observations under
58 489 natural canopies (Daly *et al.*, 2008; Guswa, 2005; Helvey and Patric, 1965). Qualitative con-

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3 490 siderations suggest that this is a reasonable assumption for relatively dense urban canopies,
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5 491 while the question is more complex for individual trees. In fact, under optimal conditions,
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7 492 isolated trees may achieve higher leaf area index than forest trees, thanks to the light avail-
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9 493 ability from the sides, but harsh urban growing environments may limit leaf and branch
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11 494 production thus resulting in lower-than-natural interception rates. In alternative to natural
12
13 495 canopy data, species-specific interception thresholds can be inferred from empirical relation-
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15 496 ships linking leaf area index to canopy interception storage capacity (see e.g., Aston (1979);
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17 497 Thompson *et al.* (1981)).
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28 498 5. Irrigation parameters

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32 499 If irrigation is implemented, we assume that a deficit irrigation is performed for water
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34 500 conservation purposes. While the effects of water limitations are highly species-specific and
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36 501 deficit irrigation should account for these specific responses, Kopinga (1985) suggests that
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38 502 urban tree transpiration should be at least 75% of its well-watered value to maintain an
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40 503 acceptable vegetation health and aesthetic quality. Following this indication, in presence
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42 504 of irrigation, a deficit irrigation with intervention point $\tilde{s} = 0.75s^*$ is assumed (i.e., the
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44 505 minimum acceptable soil moisture level is set at $0.75s^*$). For traditional irrigation it is
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46 506 assumed that water applications are such that soil moisture level is restored to 80% of soil
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48 507 water holding capacity, i.e., the soil moisture target level is set to $\hat{s} = 0.8s_1$. While shallower
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50 508 irrigation applications may be more efficient for water conservation purposes, the assumed
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52 509 almost soil saturating irrigation application limits the frequency of required applications
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54 510 (Vico and Porporato, 2011b), with clear economic advantages when the water application
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56 511 is labor-intensive. Conversely, if a more sophisticated micro-irrigation system is in place,
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58 512 shallow and almost continuous water applications are possible. In this case, the irrigation
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60 513 event is idealized as a continuous application of water, fully balancing evapotranspiration
514 losses at \tilde{s} till the next (effective) rainfall event.

III. IMPACT OF PLANTING GEOMETRY, SPECIES SELECTION, AND CLIMATE ON TREE WATER STATUS AND IRRIGATION REQUIREMENTS

To provide useful indications for adequate and sustainable species selection and planting design under different climatic scenarios, in this section we explore the effect of tree transpiration rate under well watered-conditions, extension of the permeable pavement ring, planting density, and rainfall pattern on effective rainfall input (Fig. 2), soil moisture (Fig. 3), tree effective cooling capacity and dynamic water stress in absence of irrigation (Fig. 4), as well as on irrigation requirements (in terms of water volumes and application frequencies; Figs. 5 and 6). The Wolfram Mathematica codes used to produce the results presented in this paper are available from the authors upon request.

A. Effect of permeable pavement extension on soil moisture probability density function

In the proposed idealization of the problem (Fig. 1b), the dimension of the permeable pavement plays a complex role on soil moisture dynamics, through its impact on the amount of rainfall contributing to tree available soil water, as well as the lateral extension of the root for large trees (with potentially wider rooting zones for higher r_P) and hence overall soil water storage volume (Fig. 2). Consequently, for a set tree density (i.e., a given r_T ; black lines in Fig. 2), average infiltrable water volume \bar{R} and average soil moisture increase with the area of permeable pavement, till the point beyond which the enhanced infiltrated water cannot be fully exploited because it infiltrates beyond the rooting zone (Fig. 2b, black line). A similar pattern is observed when the distance between adjacent trees is allowed to increase linearly with r_P (i.e., r_I is kept constant, while r_T increases; Fig. 2a, grey dashed line), even though the decline in contributing water is less sharp when $r_P > r_C - r_B$ (Fig. 2b, grey line).

For the case of set tree density, some examples of numerically generated soil moisture time series with no irrigation for three radii of permeable pavement and in presence of two irrigation strategies (and intermediate permeable pavement radius) are reported in Fig. 3a,c, along with the corresponding soil moisture pdf under stochastic steady state conditions (Eq. 7; Fig. 3b,d). As a consequence of the dependence of \bar{R} on r_P (Fig. 2b), there is an

intermediate dimension of the permeable pavement ring that maximizes soil water content. This is apparent in both the soil moisture dynamics and the corresponding pdfs, where the highest average soil moisture levels are obtained for r_P such that $r_B + r_P = r_C$ (which corresponds to $r_P = 2$ m in Fig. 3), while extremely low or high r_P result in very similar pdfs of soil moisture (Fig. 3b, dotted and solid lines). Assuming such intermediate r_P , Fig. 3 (bottom) illustrates the effect of irrigation applications (solid lines refer to micro-irrigation, dashed ones to traditional irrigation). Obviously, for both irrigation methods, the soil moisture process tends to spend more time at higher values than for rainfed conditions. Nevertheless, the almost soil saturating target level \hat{s} , imposed to traditional irrigation for practical reasons, causes wide fluctuations in soil moisture, mainly between the intervention point \tilde{s} and the target level \hat{s} , with excursions above the latter threshold caused by either very deep rainfall events (see the jump in soil moisture at around $t = 20$ d in the example reported in Fig. 3c) or precipitations immediately following an irrigation application (after $t = 120$ d in Fig. 3c). Conversely, the more sophisticated micro-irrigation results in the soil moisture process spending a finite amount of time at the intervention point \tilde{s} , while waiting for the next effective rainfall event. This fact is mirrored by the mixed pdf of soil moisture, consisting in a continuous part (solid line) and an atom of probability in \tilde{s} (solid bar), representing the non-zero probability that the soil moisture process is at \tilde{s} (Vico and Porporato, 2010).

B. Tree water stress with no irrigation

Tree water requirements, rainfall pattern, and fraction of permeable vs. impervious pavement around the tree nonlinearly affect tree cooling capacity, $\langle T \rangle / T_{max}$, and dynamic water stress, $\bar{\theta}$. For each T_{max} , there is an intermediated permeable area that maximize $\langle T \rangle / T_{max}$ and minimizes the value of $\bar{\theta}$, corresponding to $r_P = r_C - r_B$ (dashed line in Fig. 4a,d), which progressively increases with T_{max} (with which the canopy lateral extension grows with power 1/3; see II.F.4). Assuming a set tree density, for larger and more water-demanding trees (i.e., higher T_{max} ; Fig. 4a,d), dynamic water stress increases nonlinearly with decreasing permeable areas, in particular at low r_P .

The effect of shifts in the rainfall pattern is explored in Fig. 4b,f, where total rainfall over the growing season is held constant while increasing α and simultaneously decreasing

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4 574 λ . Low r_P results in low cooling capacities and high water stress levels regardless of rainfall
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6 575 pattern, because it limits effective root lateral extension and hence tree available water
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8 576 storage capacity. For medium-to-high r_P , both very low and very high rainfall frequencies
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10 577 are detrimental for cooling capacity and plant water status: infrequent but deep rainfall
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12 578 events enhance losses through runoff and deep percolation, thus reducing the amount of
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14 579 water available for plant transpiration; conversely, frequent but shallow rainfall events are
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16 580 mostly intercepted by the canopy, thus limiting soil water recharge. Permeable pavement
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18 581 extension $r_P = r_C - r_B$ and intermediate λ are ideal for tree ability to provide ecosystem
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20 582 services. The effect of rainfall pattern becomes less and less marked for permeable pavement
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22 583 extending beyond the canopy in particular regarding cooling capacity, because, in this case,
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24 584 soil water recharge from outer areas quickly tapers off with increasing distance.

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27 585 The assumption of set distance between adjacent trees is relaxed in Fig. 4c,f, where the
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29 586 combined effects of permeable and impervious pavement dimensions are explored for set
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31 587 species and climatic conditions. As expected, higher tree density (i.e., lower distances from
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33 588 the origin in Fig. 4c,f) has negative effects on tree cooling capacity and water status, because
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35 589 of the limited area that can be exploited for water collection and soil moisture recharge.
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37 590 More interesting is the quantification of the differential effect of an increase in either r_P
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39 591 or r_I . Similar reductions in cooling capacity and dynamic water stress can be achieved
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41 592 with a smaller increase in permeable pavement extension than in impervious pavement area,
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43 593 particularly at intermediate tree densities. Under these conditions, larger permeable areas
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45 594 allow a wider extension of roots as well as a more efficient collection of precipitated water,
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47 595 which could not be achieved with an equivalent amount of impervious surface, imposing
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49 596 limits to the amount of rainfall effectively contributing to meet tree water requirements.
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51 597 Hence, at intermediate tree densities (dashed line in Fig. 4c corresponds to $r_T = 7.5$ m),
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53 598 maximizing permeable areas may improve plant water status, while simultaneously limiting
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55 599 the negative effect caused by increased tree density. Conversely, extremely dense trees limit
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57 600 the lateral extension of non-competing roots and hence the buffering effects of soil volume
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59 601 regardless of the pavement permeability, while more isolated trees benefit the most from
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602 intermediate permeable area extension (the optimal width of the paved area corresponds to
603 $r_P = r_C - r_B$).

604 C. Irrigation requirements and optimal irrigation strategy

605 If irrigation is implemented, irrigation requirements (water volumes and application fre-
606 quency), as well as sustainability, cost, and feasibility of traditional vs. micro-irrigation
607 strongly depend on species selection, tree size, rainfall pattern, and planting geometry (Figs.
608 5 and 6).

609 Fig. 5a shows how the required volumes for traditional irrigation increase almost linearly
610 with transpiration rates T_{max} ; similar patterns are obtained for micro-irrigation, although
611 with lower total water requirements than for traditional irrigation, and with less steep in-
612 creases with increasing T_{max} (not shown). The pattern is more complex when altering rainfall
613 timing while maintaining rainfall totals (Fig. 5b). For any rainfall frequency, the minimum
614 water requirements are to be expected in connection with planting designs with $r_P = r_C - r_B$
615 (corresponding to those designs limiting tree water stress; Fig. 4). At this optimal perme-
616 able pavement width, the difference in terms of water requirements between micro-irrigation
617 and traditional irrigation is maximized, in particular for more frequent rainfall events, with
618 traditional irrigation requiring up to twice as much water as micro-irrigation (not shown). In
619 fact, with more frequent rainfall events, it becomes more likely that an irrigation application
620 is immediately followed by a rainfall event, the water input of which is then partially lost to
621 runoff and deep percolation because of the relatively high soil moisture at the event time.
622 Furthermore, infrequent but deep rainfall events may result in higher water requirements
623 than less intermittent rainfall patterns, despite the lower intercepted fraction of water typical
624 of the former rainfall pattern (Fig. 5b). Irrigation is often required during the inter-storm
625 periods to sustain plant transpiration and, when rainfall occurs, saturation-excess may re-
626 sult in the loss of a significant fraction of the precipitated water. Finally, trees planted at
627 high density (i.e., low $r_P + r_I$) will have higher average irrigation requirements than sparser
628 trees, because in the former case the almost continuous canopy enhances rainfall interception
629 and the precipitated water is split between adjacent trees. To facilitate soil water recharge
630 and limit irrigation water requirements, permeable areas should be maximized at high tree
631 density, while an intermediate $r_P = r_C - r_B$ should be sought for lower densities (Fig. 5c).

632 Required irrigation frequency for traditional irrigation determines its practical applica-
633 bility in the urban context or whether a more sophisticated system is necessary. Irrigation
634 frequency significantly increases with T_{max} , regardless of planting geometry (Fig. 6 a) and,

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4 635 for a set T_{max} , the minimum frequency occurs at $r_P = r_C - r_B$. For altered rainfall patterns
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6 636 the main determinant of irrigation frequency is the extension of the permeable area, with
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8 637 extremely high irrigation frequencies for very low r_P (i.e., extremely small effective rooting
9
10 638 volumes that cannot efficiently buffer plant water uptake). Conversely, at higher r_P , re-
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12 639 quired irrigation frequency is less sensitive to changes in rainfall frequency and extension of
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14 640 permeable pavement (Fig. 6b).

17 641 **D. Strategies for species selection and planting design**

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20 642 As apparent from the above results, the choice of tree species, size, and planting design
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22 643 requires considering several contrasting needs, the relevance of each one depending on the
23
24 644 specific location. On the one hand, it is necessary to consider the provisioning of ecosystem
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26 645 services by the street trees, and how they are altered by growing conditions: average transpi-
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28 646 ration per tree provides a measure of the ability to providing a cooling effect, while dynamic
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30 647 water stress provides an indication regarding the tree aesthetic quality and health. On the
31
32 648 other hand, for those species, planting designs, and climatic conditions where irrigation is
33
34 649 necessary to preserve adequate ecosystem service provisions, irrigation requirements needs
35
36 650 appropriate consideration.

37
38 651 To illustrate the usage of the proposed decision tool, we focus on the effect of permeable
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40 652 pavement dimensions on two trees, with 60 and 100 kg d⁻¹ tree⁻¹ water requirements (solid
41
42 653 and dot-dashed gray lines respectively in Figs. 4a,d, 5a, and 6a). Under rainfed conditions,
43
44 654 the less water demanding tree has an effective cooling capacity of 83% of its potential for
45
46 655 $r_P = 1.8$ m, and at least 70% for other permeable pavement dimensions. Similarly, the
47
48 656 dynamic water stress is lowest at such optimal r_P ($\bar{\theta} = 0.2$). Conversely, for the more
49
50 657 demanding (and larger) tree, maximum cooling is 54% of potential, at $r_P = 2.4$ m. Smaller
51
52 658 permeable pavement areas may reduce the cooling ability to 40% of potential, with dynamic
53
54 659 water stress levels reaching 0.55. While in most species a certain level of dynamic water
55
56 660 stress may not significantly limit aesthetic quality and longevity, it is not possible to provide
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58 661 a general threshold above which such ecosystem services can no longer be provided. In fact,
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60 662 the effects of medium-to-high dynamic stress on tree status strongly depend on the species-
663 specific response to water limitation. Some species may lose their leaves in response to
664 extended periods of water stress, other may sustain damages if exposed to frequent stress

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4 665 episodes. Hence, in most cases, the definition of the conditions under which a supplemental
5 666 irrigation should be implemented to preserve ecosystem services beyond cooling capacity
6
7 667 will require the evaluation of species specific response to water stress indicators, ranging
8
9 668 from the dynamic water stress, $\bar{\theta}$, to the length of periods of water stress, $\bar{T}^\downarrow(s^*)$ and their
10 669 frequency, $\nu^\downarrow(s^*)$. While clearly extremely relevant for planning, information on the species-
11
12 670 specific response to water stress tends to be difficult to access for urban planners, as recently
13
14 671 discussed for the case of Scandinavia by Sjöman and Nielsen (2010).

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16 672 When irrigation is necessary to preserve tree ecosystem services, the proposed model
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18 673 can provide quantitative information on irrigation requirements, as a function of tree water
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20 674 needs (Figs. 5 and 6). To assess irrigation feasibility, consideration needs to be given to
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22 675 required volumes and irrigation frequency. Maximum acceptable volumes depend on total
23
24 676 water allocation and number of trees to be watered. For the cases under scrutiny, the most
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26 677 demanding tree would require between 6 and 8 m³ per season, depending on planting geome-
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28 678 try (Fig. 5a), a figure that might not be acceptable under water shortage or when concerning
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30 679 a high number of plants. Furthermore, the previous results provide a quantitative basis to
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32 680 assess under which circumstances traditional irrigation is a feasible option or the more so-
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34 681 phisticated micro-irrigation may be needed. As discussed in Vico and Porporato (2011b),
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36 682 regardless of existing conditions, micro-irrigation has lower water requirements (not shown),
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38 683 thanks to its higher efficiency. The water savings associated to micro-irrigation may trans-
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40 684 late in an economic advantage when water has high costs. Nevertheless, micro-irrigation has
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42 685 high installation and maintenance costs, in particular in urban settings where damages may
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44 686 occur because of vandalism and pedestrian traffic. Traditional irrigation applied through
45
46 687 replenishment of tree bags or with direct irrigation with hose has low investment costs, but
47
48 688 high application costs, associated to labor costs (and higher water expenses, when the cost of
49
50 689 water is significant). Because of the relatively low required application frequency (Fig. 6a),
51
52 690 traditional irrigation is likely to remain the most economically viable option in most cases.
53
54 691 The only exceptions are trees with very high water demands or growing in locations with ex-
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56 692 tremely low rainfall inputs, when the installation and maintenance costs of micro-irrigation
57
58 693 are lower than the costs associated to high frequency applications of traditional irrigation,
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60 694 thus making the more sophisticated system advantageous also under the economic point of
695 view. Aside from the water conservation and economic questions, other aspects may play a
696 role in the choice of the most appropriate management strategies. In particular, in the ab-

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4 697 sence of occasional deep rainfall events, micro-irrigation may be responsible for salt build-up
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6 698 in the soil, especially when using saline water for irrigation or in areas with winter de-icing
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8 699 compound applications. Also, there are some indications that irrigation may result in the
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10 700 development of shallower root systems (Bijoor *et al.*, 2012), with possible implications for
11
12 701 tree stability and susceptibility to droughts. Finally, for long term planning, our framework
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14 702 can easily account for tree growth, which will be reflected on total tree water requirements,
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16 703 and can explicitly include changes in the rainfall pattern due to climate change scenarios,
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18 704 such as those investigated in Figs. 5b,e 5b, and 6b.

19 20 21 705 **IV. CONCLUSIONS**

22
23 706 A minimalist description of the soil water balance for urban trees was proposed, explicitly
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25 707 including rainfall unpredictability within a probabilistic framework, while still only requiring
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27 708 few, physically-based parameters characterizing rainfall pattern and vegetation response to
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29 709 water availability. The proposed model allows us to quantify the effect of species selection,
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31 710 tree size, and planting design on total average seasonal transpiration (and thus effective
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33 711 cooling capacity), tree water status (and thus health and aesthetic quality), and irrigation
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35 712 requirements. Hence, this minimalist description represents a first necessary step towards
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37 713 the definition of site-specific guidelines for species selection and planting design, to limit
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39 714 city water footprint while preserving street tree ability to provide ecosystem services. The
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41 715 planting design that maximizes cooling capacity while minimizing water stress occurrence
42
43 716 and irrigation requirements may be achieved by bare soil and permeable pavement with
44
45 717 combined area equal to the canopy extension. Because of the complex balance between root
46
47 718 lateral extension and efficient soil water recharge by precipitated water, denser trees benefit
48
49 719 more from permeable than impervious surfaces, while isolated trees benefit the most from
50
51 720 intermediate permeable area extensions. When irrigation becomes necessary to maintain
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53 721 the desired ecosystem services, small permeable areas and trees planted at high density
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55 722 require higher irrigation input to maintain low water stress than a more balanced design.
56
57 723 Because of its higher efficiency, micro-irrigation has lower total water requirements, and
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59 724 may be an adequate irrigation strategy for low permeable area extensions when the high
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725 required frequency of traditional irrigation may be unpractical and water savings by micro-
726 irrigation are the highest. Intermediate rainfall frequencies and event depths allow the

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3 727 minimization of water stress occurrence and severity, and irrigation requirements. Shifts
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5 728 from this rainfall regime, particularly towards deeper but less frequent rainfall events, have
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7 729 negative repercussions for tree cooling capacity and water stress, and enhance irrigation
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9 730 requirements, especially for trees surrounded by wide permeable areas. This is true also for
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11 731 trees with optimal or larger permeable pavements, even though the situation remains more
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13 732 positive than for narrow permeable zones. The results presented here can provide helpful
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15 733 indications for the definition of guidelines towards a sustainable design of urban vegetation
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17 734 under current and future climate scenarios. The predictive power of the proposed model
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19 735 would be greatly enhanced by a wider availability of data on plant water requirements under
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21 736 urban-specific growing conditions.

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41 42 43 746 **REFERENCES**

44
45
46 747 Aston, A R (1979), "Rainfall interception by eight small trees," *Journal of Hydrology* **42**, 383–396.
47
48 748 Balling, R C, P. Gober, and N. Jones (2008), "Sensitivity of residential water consumption to
49
50 749 variations in climate: an intraurban analysis of Phoenix, Arizona," *Water Resour. Res.* **44** (10),
51
52 750 W10401.
53
54 751 Beckett, K P, P. H. Freer-Smith, and G. Taylor (1998), "Urban woodlands: their role in reducing
55
56 752 the effects of particulate pollution," *Environmental Pollution* **99** (3), 347–360.
57
58 753 Berrang, P, D. F. Karnosky, and B. J. Stanton (1985), "Environmental factors affecting tree health
59
60 754 in New York City," *J. Arbor.* **11** (6), 185–189.

- 1
2
3
4 755 Bijoor, N S, H. R. McCarthy, D. C. Zhang, and D. E. Pataki (2012), "Water sources of urban
5
6 756 trees in the los angeles metropolitan area," *Urban Ecosystems* **15** (1), 195–214.
- 7
8 757 Bowler, D E E, L. Buyung-Ali, T. M. M. Knight, and A. S. S. Pullin (2010), "Urban greening
9
10 758 to cool towns and cities: a systematic review of the empirical evidence," *Landscape and Urban*
11
12 759 *Planning* **97** (3), 147–155.
- 13
14 760 Brennan, D, S. Tapsuwan, and G. Ingram (2007), "The welfare costs of urban outdoor water
15
16 761 restrictions," *The Australian Journal of Agricultural and Resource Economics* **51**, 243–261.
- 17
18 762 Bush, S E, D. E. Pataki, K. R. Hultine, A. G. West, J. S. Sperry, and J. R. Ehleringer (2008),
19
20 763 "Wood anatomy constrains stomatal responses to atmospheric vapor pressure deficit in irrigated,
21
22 764 urban trees," *Oecologia* **156** (1), 13–20.
- 23
24 765 Chen, L, Z. Zhang, Z. Li, J. Tang, P. Caldwell, and W. Zhang (2011), "Biophysical control of
25
26 766 whole tree transpiration under an urban environment in Northern China," *J. Hydrol.* **402** (3-4),
27
28 767 388–400.
- 29
30 768 Clark, J R, and R. Kjelgren (1990), "Water as a limiting factor in the development of urban trees,"
31
32 769 *J. Arbor.* **16**, 203–208.
- 33
34 770 Clark, J R, N. P. Matheny, G. Cross, and V. Wake (1997), "A model of urban forest sustainability,"
35
36 771 *J. Arbor.* **23** (1), 17–30.
- 37
38 772 Coder, K (1996), *Cultural aspects of trees: traditions and myth*. (Coperative Extension Service,
39
40 773 Forest Resources Unit, University of Georgia, Athens, GA).
- 41
42 774 Craul, P J (1985), "A description of urban soils and their desired characteristics," *J. Arbor.* **11** (11),
43
44 775 330–339.
- 45
46 776 Craul, P J (1999), *Urban soils: applications and practices* (John Wiley & Sons, New York, NY).
- 47
48 777 Cregg, B M (1995), "Plant moisture stress of green ash trees in contrasting urban sites," *J. Arbor.*
49
50 778 **21** (6), 271–276.
- 51
52 779 Cregg, B M, and M. E. Dix (2001), "Tree moisture stress and insect damage in urban areas in
53
54 780 reation to heat island effects," *J. Arbor.* **27** (1), 8–17.
- 55
56 781 Daly, E, A. C. Oishi, A. Porporato, and G. G. Katul (2008), "A stochastic model for daily
57
58 782 subsurface CO₂ concentration and related soil respiration," *Adv. Water Resour.* **31** (7), 987–994.
- 59
60 783 DeGaetano, A T, and S. R. Hudson (2000), "Specification of soil volume and irrigation frequency
784 for urban trees," *J. Urban Plan. Dev. ASCE* **126** (4), 153–165.

- 1
2
3
4 785 Delcambre, A, and J. P. Rossignol (1999), "Characterization of moderate hydric stress on ash trees
5 786 (*Fraxinus americana*) in landscaped areas," in *International Symposium on Urban Tree Health*,
6
7 787 *Acta Horticulturae*, edited by M. Lemattre P. Lemaire F. Lemattre, pp. 353–360.
8
9 788 Drunasky, N, and D. K. Struve (2005), "*Quercus macrocarpa* and *Q. Prinus* physiological and
10 789 morphological responses to drought stress and their potential for urban forestry," *Urban forestry*
11 & *Urban Greening* **4**, 13–22.
12
13
14 791 Dwyer, J F, E. G. McPherson, H. W. Schroeder, and R. A. Rowntree (1992), "Assessing the
15 792 benefits and costs of the urban forest," *J. Arbor.* **18** (5), 227–234.
16
17
18 793 Dwyer, J F, D. J. Nowak, and M. H. Noble (2003), "Sustaining urban forests," *J. Arbor.* **29** (1),
19 794 49–55.
20
21
22 795 Dwyer, J F, D. J. Nowak, and G. W. Watson (2002), "Future directions for urban forestry research
23 796 in the United States," *J. Arbor.* **28**, 231–236.
24
25
26 797 Easterling, D R, G. A. Meehl, C. Parmesan, S. A. Changnon, T. R. Karl, and L. O. Mearns (2000),
27 798 "Climate extremes: observations, modeling, and impacts," *Science* **289** (5487), 2068–2074.
28
29
30 799 English, M, and S. N. Raja (1996), "Perspectives on deficit irrigation," *Agricultural Water Man-*
31 800 *agement* **32** (1), 1–14.
32
33 801 Fernandez-Juricic, E (2000), "Avifaunal use of wooded streets in an urban landscape," *Conserv.*
34 802 *Biol.* **14** (2), 513–521.
35
36
37 803 Ferrini, F, and A. Fini (2011), "Sustainable management techniques for trees in the urban areas,"
38 804 *Journal of Biodiversity and Ecological Sciences* **1** (1), 1–19.
39
40
41 805 Flückiger, W, and S. Braun (1999), "Stress factors of urban trees and their relevance for vigour
42 806 and predisposition for parasite attacks," *Acta Horticulturae* **496**, 325–334.
43
44
45 807 Foster, R S, and J. Blaine (1978), "Urban tree survival: trees in the sidewalk," *J. Arbor.* **4** (1),
46 808 14–17.
47
48
49 809 Gill, S E, J. F. Handley, A. R. Ennos, and S. Pauleit (2007), "Adapting cities for climate change:
50 810 The role of the green infrastructure," *Built Environment* **33** (1), 115–133.
51
52
53 811 Grabosky, J, N. Bassuk, L. Irwin, and H. Van Es (2001), "Shoot and root growth of three tree
54 812 species in sidewalks," *Journal of Environmental Horticulture* **19** (4), 206–211.
55
56
57 813 Guswa, A J (2005), "Soil-moisture limits on plant uptake: An upscaled relationship for water-
58 814 limited ecosystems," *Adv. Water Resour.* **28** (6), 543–552.
59
60

- 1
2
3
4 815 Hagishima, A, K. Narita, and J. Tanimoto (2007), "Field experiment on transpiration from isolated
5
6 816 urban plants," *Hydrol. Proc.* **21** (9), 1217–1222.
- 7 817 Hauer, R J, R. W. Miller, and D. M. Ouimet (1994), "Street tree decline and construction damage,"
8
9 818 *J. Arbor.* **20** (2), 94–97.
- 10 819 Helvey, J D, and J. H. Patric (1965), "Canopy and litter interception of rainfall by hardwoods of
11
12 820 Eastern United States," *Water Resour. Res.* **1** (2), 193–206.
- 13 821 Hilaire, R S, M. A. Arnold, D. C. Wilkerson, D. A. Devitt, B. H. Hurd, B. J. Lesikar, V. I. Lohr,
14
15 822 C. A. Martin, G. V. McDonald, R. L. Morris, D. R. Pittenger, D. A. Shaw, and D. F. Zoldoske
16
17 823 (2008), "Efficient water use in residential urban landscapes," *Hortscience* **43** (7), 2081–2092.
- 18 824 Home, R, N. Bauer, and M. Hunziker (2010), "Cultural and biological determinants in the evalu-
19
20 825 ation of urban green spaces," *Environment and Behavior* **42** (4), 494–523.
- 21
22 826 Imhoff, M L, P. Zhang, R. E. Wolfe, and L. Bounoua (2010), "Remote sensing of the urban heat
23
24 827 island effect across biomes in the continental USA," *Remote Sens. Environ.* **114** (3), 504–513.
- 25
26 828 Jenerette, G D, S. L. Harlan, W. L. Stefanov, and C. A. Martin (2011), "Ecosystem services and
27
28 829 urban heat riskscape moderation: water, green spaces, and social inequality in Phoenix, USA,"
29
30 830 *Ecological Applications* **21** (7), 2637–2651.
- 31
32 831 Jenerette, G D, and L. Larsen (2006), "A global perspective on changing sustainable urban water
33
34 832 supplies," *Glob. Planet. Change* **50** (3-4), 202–211.
- 35
36 833 Jorgensen, A, and P. H. Gobster (2010), "Shades of green: measuring the ecology of urban green
37
38 834 space in the context of human health and well-being," *Nature and Culture* **5** (3), 338–363.
- 39
40 835 Kalnay, E, and M. Cai (2003), "Impact of urbanization and land-use change on climate," *Nature*
41
42 836 **423** (6939), 528–531.
- 43
44 837 Kjelgren, R K, and J. R. Clark (1993), "Growth and water relations of *Liquidambar styraciflua* l.
45
46 838 in an urban park and plaza," *Trees-Structure and Function* **7** (4), 195–201.
- 47
48 839 Konijnendijk, C C (2000), "Adapting forestry to urban demands - role of communication in urban
49
50 840 forestry in Europe," *Landscape and Urban Planning* **52** (2-3), 89–100.
- 51
52 841 Kopinga, J (1985), "Research on street tree planting practices in the Netherlands," in *5th Annual*
53
54 842 *METRIA Conference* (Peensylvania State University, University Park, PA) pp. 72–84.
- 55
56 843 Körner, C H, J. A. Scheel, and H. Bauer (1979), "Maximum leaf diffusive conductance in vascular
57
58 844 plants," *Photosynthetica* **13** (1), 45–82.
- 59
60

- 1
2
3
4 845 Kuo, F E, and W. C. Sullivan (2001), "Environment and crime in the inner city: does vegetation
5 reduce crime?" *Environment and Behavior* **33** (3), 343–367.
6
7 847 Laio, F, A. Porporato, L. Ridolfi, and I. Rodriguez-Iturbe (2001), "Plants in water-controlled
8 ecosystems: active role in hydrologic processes and response to water stress - II. Probabilistic
9 soil moisture dynamics," *Adv. Water Resour.* **24** (7), 707–723.
10
11 849 Lindsey, P, and N. L. Bassuk (1991), "Specifying soil volumes to meet the water needs of mature
12 urban street trees and trees in containers," *J. Arbor.* **17** (6), 141–149.
13
14 850 Litvak, E, H. R. McCarthy, and D. E. Pataki (2011), "Water relations of coast redwood planted
15 in the semi-arid climate of southern California," *Plant Cell Environ.* **34** (8), 1384–1400.
16
17 852 Litvak, Elizaveta, Heather R. McCarthy, and Diane E. Pataki (2012), "Transpiration sensitivity
18 of urban trees in a semi-arid climate is constrained by xylem vulnerability to cavitation," *Tree
19 Physiology* **32** (4), 373–388.
20
21 854 Lohr, V I, C. H. Pearson-Mims, J. Tarnai, and D. A. Dillman (2004), "How urban residents rate
22 and rank the benefits and problems associated with trees in cities," *J. Arbor.* **30** (1), 28–35.
23
24 855 Maas, J, R. A. Verheij, P. P. Groenewegen, S. de Vries, and P. Spreeuwenberg (2006), "Green
25 space, urbanity, and health: how strong is the relation?" *J. Epidemiol. Community Health* **60** (7),
26 587–592.
27
28 856 MacDonald, DH, N. D. Crossman, P. Mahmoudi, L. O. Taylor, M. D. Summers, and P. C. Boxall
29 (2010), "The value of public and private green spaces under water restrictions," *Landscape and
30 Urban Planning* **95** (4), 192–200.
31
32 857 McCarthy, H R, and D. E. Pataki (2010), "Drivers of variability in water use of native and
33 non-native urban trees in the greater los angeles area," *Urban Ecosystems* **13**, 393–414.
34
35 858 McHale, M R, I. C. Burke, M. A. Lefsky, P. J. Peper, and E. G. McPherson (2009), "Urban forest
36 biomass estimates: is it important to use allometric relationship developed specifically for urban
37 trees?" *Urban Ecosyst.* **12**, 95–113.
38
39 859 McPherson, E G (2001), "Sacramento's parking lot shading ordinance: environmental and economic
40 costs of compliance," *Landscape and Urban Planning* **57** (2), 105–123.
41
42 860 McPherson, E G, and R. A. Rowntree (1989), "Using structural measures to compare twenty-two
43 U.S. street tree populations," *Landscape Journal* **8** (1), 13–23.
44
45 861 McPherson, E G, J. R. Simpson, Q. F. Xiao, and C. X. Wu (2011), "Million trees Los Angeles
46 canopy cover and benefit assessment," *Landscape and Urban Planning* **99** (1), 40–50.
47
48
49
50
51
52
53
54
55
56
57
58
59
60

- 1
2
3
4 876 Milly, P C D (2001), "A minimalist probabilistic description of root zone soil water," *Water Resour.*
5
6 877 *Res.* **37** (3), 457–463.
- 7 878 Morgenroth, J, and G. D. Buchan (2009), "Soil moisture and aeration beneath pervious and
8
9 879 impervious pavements," *Arboriculture & Urban Forestry* **35** (3), 135–141.
- 10 880 Niinemets, U, and J. Penuelas (2008), "Gardening and urban landscaping: significant players in
11
12 881 global change," *Trends Plant Sci.* **13** (2), 60–65.
- 13
14 882 Nowak, D J, D. E. Crane, and J. C Stevens (2006), "Air pollution removal by urban trees and
15
16 883 shrubs in the United States," *Urban forestry & Urban Greening* **4**, 115–123.
- 17
18 884 Nowak, D J, J. R. McBride, and R. A. Beatty (1990), "Newly planted street tree growth and
19
20 885 mortality," *J. Arbor.* **16** (5), 124–130.
- 21
22 886 Parés-Franzi, M, D. Sauri-Pujol, and E. Domene (2006), "Evaluating the environmental perfor-
23
24 887 mance of urban parks in mediterranean cities: An example from the Barcelona metropolitan
25
26 888 region," *Environmental Management* **38** (5), 750–759.
- 27
28 889 Pataki, D E, C. G. Boone, T. S. Hogue, G. D. Jenerette, J. P. McFadden, and S. Pincetl (2011a),
29
30 890 "Socio-ecohydrology and the urban water challenge," *Ecohydrology* **4** (2), 341–347.
- 31
32 891 Pataki, D E, H. R. R. McCarthy, E. Litvak, and S. Pincetl (2011b), "Transpiration of urban forests
33
34 892 in the Los Angeles metropolitan area," *Ecological Applications* **21** (3), 661–677.
- 35
36 893 Pauleit, S (2003), "Urban street tree plantings: indentifying the key requirements," *Proceedings*
37
38 894 *of the Institution of Civil Engineers-Municipal Engineer* **156** (1), 43–50.
- 39
40 895 Payton, S, G. Lindsey, J. Wilson, J. R. Ottensmann, and J. Man (2008), "Valuing the benefits of
41
42 896 the urban forest: a spatial hedonic approach," *J. Environ. Plan. Manag.* **51** (6), 717–736.
- 43
44 897 Peters, E B, J. P. McFadden, and R. A. Montgomery (2010), "Biological and environmental
45
46 898 controls on tree transpiration in a suburban landscape," *J. Geophys. Res. Biogeosci.* **115**, G04006.
- 47
48 899 Porporato, A, E. Daly, and I. Rodriguez-Iturbe (2004), "Soil water balance and ecosystem response
49
50 900 to climate change," *Am. Nat.* **164** (5), 625–632.
- 51
52 901 Porporato, A, F. Laio, L. Ridolfi, and I. Rodriguez-Iturbe (2001), "Plants in water-controlled
53
54 902 ecosystems: active role in hydrologic processes and response to water stress - III. Vegetation
55
56 903 water stress," *Adv. Water Resour.* **24** (7), 725–744.
- 57
58 904 Quattrochi, D A, and M. K. Ridd (1998), "Analysis of vegetation within a semi-arid urban environ-
59
60 905 ment using high spatial resolution airborne thermal infrared remote sensing data," *Atmospheric*
906 *Environment* **32** (1), 19–33.

- 1
2
3
4 907 Raupp, J M, A. B. Cumming, and E. C. Raupp (2006), "Street tree diversity in Eastern North
5 America and its potential for tree loss to exotic borers," *Arboriculture & Urban Forestry* **32** (6),
6 297–304.
7
8
9 910 Rees, W, and M. Wackernagel (2008), "Urban ecological footprints: why cities cannot be
10 sustainable—and why they are a key to sustainability," in *Urban Ecology*, edited by J. M. Marzluff,
11 E. Shulenberg, W. Endlicher, M. Alberti, G. Bradley, C. Ryan, U. Simon, and C. ZumBrunnen
12 (Springer US, Berlin) pp. 537–555.
13
14
15 914 Reichwein, S (2003), *Root growth under pavements - results of a field study*, Second International
16 Symposium on plant health in urban horticulture, Berlin, Germany, 27-29 August, 2003.
17
18
19 915
20 916 Richards, N A (1983), "Diversity and stability in a street tree population," *Urban Ecology* **7**,
21 159–171.
22
23
24 918 Roberts, B R, and V. M. Schnipke (1994), "The relative water demand of five urban tree species,"
25 *J. Arbor.* **20** (3), 156–159.
26
27
28 920 Rodriguez-Iturbe, I, A. Porporato, L. Ridolfi, V. Isham, and D. R. Cox (1999), "Probabilistic
29 modelling of water balance at a point: The role of climate, soil and vegetation," *Proc. R. Soc.*
30 *Lond. A* **455**, 3789–3805.
31
32
33 923 Rodriguez-Iturbe, R, and A. Porporato (2004), *Ecohydrology of water-controlled ecosystems - Soil*
34 *moisture and plant dynamics* (Cambridge University Press, Cambridge, MA).
35
36
37 925 Sæbø, A, T. Benedikz, and T. B. Randrup (2003), "Selection of trees for urban forestry in the
38 Nordic countries," *Urban forestry & Urban Greening* **2**, 101–114.
39
40
41 927 Salvador, R, C. Bautista-Capetillo, and E. Playan (2011), "Irrigation performance in private urban
42 landscapes: a study case in Zaragoza (Spain)," *Landscape and Urban Planning* **100** (3), 302–311.
43
44
45 929 Schenk, H J, and R. B. Jackson (2002), "The global biogeography of roots," *Ecol. Monogr.* **72** (3),
46 311–328.
47
48
49 931 Shashua-Bar, L, D. Pearlmutter, and E. Erell (2009), "The cooling efficiency of urban landscape
50 strategies in a hot dry climate," *Landscape and Urban Planning* **92** (3-4), 179–186.
51
52
53 933 Shooshtarian, S, A. TehraniFar, A. Ghani, and M. Kiani (2011), "Effects of irrigation frequency
54 regimes on morphological and physiological characteristics of six sedum species," *African Journal*
55 *of Agricultural Research* **6** (26), 5694–5700.
56
57
58 936 Sjöman, H, and A. B. Nielsen (2010), "Selecting trees for urban paved sites in Scandinavia - A
59 review of information on stress tolerance and its relation to the requirements of tree planners,"
60

- 1
2
3
4 938 Urban forestry & Urban Greening **9** (4), 281–293.
- 5
6 939 Sjöman, Henrik, Johan Östberg, and Oliver Bühler (2012), “Diversity and distribution of the
7
8 940 urban tree population in ten major nordic cities,” Urban Forestry & Urban Greening **11** (1),
9
10 941 31–39.
- 11
12 942 Suleiman, M K, N. R. Bhat, S. Jacob, R. R. Thomas, and G. D’Cruz (2011), “Performance of
13
14 943 selected native plants under deficit irrigation,” World Journal of Agricultural Sciences **7** (1),
15
16 944 19–25.
- 17
18 945 Swanwick, C, N. Dunnett, and H. Woolley (2003), “Nature, role and value of green space in towns
19
20 946 and cities: an overview,” Built Environment **29** (2), 94–106.
- 21
22 947 Takagi, M, and K. Gyokusen (2004), “Light and atmospheric pollution affect photosynthesis of
23
24 948 street trees in urban environments,” Urban Forestry & Urban Greening **2**, 167–171.
- 25
26 949 Thompson, N, I. A. Barrie, and M. Ayles (1981), *The Meteorological Office Rainfall and Evapo-*
27
28 950 *ration Calculation System (MORECS)*, Tech. Rep. (Met. Office).
- 29
30 951 Thorsson, S, F. Lindberg, J. Björklund, B. Holmer, and D. Rayner (2011), “Potential changes in
31
32 952 outdoor thermal comfort conditions in Gothenburg, Sweden due to climate change: the influence
33
34 953 of urban geometry,” Int. J. Climat. **31** (2, Sp. Iss. SI), 324–335.
- 35
36 954 United Nations, (2005), *United Nations Population Division, World Urbanization Prospects, Urban*
37
38 955 *and Rural Population Estimates and Projections* (United Nations Publications Board, New York,
39
40 956 NY).
- 41
42 957 Čermák, J, J. Hruška, M. Martinková, and A. Prax (2000), “Urban tree root systems and their
43
44 958 survival near houses analyzed using ground penetrating radar and sap flow techniques,” Plant
45
46 959 Soil **219** (1-2), 103–116.
- 47
48 960 Vico, G, and A. Porporato (2010), “Traditional and microirrigation with stochastic soil moisture,”
49
50 961 Water Resour. Res. **46**, W03509.
- 51
52 962 Vico, G, and A. Porporato (2011a), “From rainfed agriculture to stress-avoidance irrigation: I. A
53
54 963 generalized irrigation scheme with stochastic soil moisture,” Adv. Water Resour. **34** (2), 263–271.
- 55
56 964 Vico, G, and A. Porporato (2011b), “From rainfed agriculture to stress-avoidance irrigation: II.
57
58 965 Sustainability, crop yield, and profitability,” Adv. Water Resour. **34** (2), 272–281.
- 59
60 966 Viswanathan, B, A. Volder, W. T. Watson, and J. A. Aitkenhead-Peterson (2011), “Impervious
967
968 and pervious pavements increase soil CO₂ concentrations and reduce root production of American
sweetgum (*Liquidambar styraciflua*),” Urban forestry & Urban Greening **10** (2), 133–139.

- 1
2
3
4 969 Walker, J S, N. B. Grimm, J. M. Briggs, C. Gries, and L. Dugan (2009), "Effects of urbanization
5
6 970 on plant species diversity in central arizona," *Frontiers in Ecology and the Environment* **7** (9),
7
8 971 465–470.
- 9
10 972 Wang, H, Z. Ouyang, W. Chen, X. Wang, H. Zheng, and Y. Ren (2011), "Water, heat, and airborne
11
12 973 pollutants effects on transpiration of urban trees," *Environmental Pollution* **159**, 2127–2137.
- 13
14 974 Whitlow, T H, N. L. Bassuk, and D. L. Reichert (1992), "A three-year study of water relations of
15
16 975 urban street trees," *J. Appl. Ecol.* **29** (2), 436–450.
- 17
18 976 Wolf, K L (2005), "Business district streetscapes, trees, and consumer response," *J. For.* **103** (8),
19
20 977 396–400.
- 21
22 978 Xiao, Q F, and E. G. McPherson (2002), "Rainfall interception by Santa Monica's municipal urban
23
24 979 forest," *Urban Ecosyst.* **6**, 291–302.
- 25
26 980 Young, R F (2010), "Managing municipal green space for ecosystem services," *Urban Forestry &*
27
28 981 *Urban Greening* **9** (4), 313–321.
- 29
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4 Figure 1 a) Examples of street trees planting design, including the usage of a tree bag for irrigation
5 purposes, and b) schematic representation of the geometry of the problem for the circular symmetric
6 case. (Photo credits: S. Manzoni, A. Porporato, G. Vico)
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17 Figure 2 a) Geometry of the problem and b) average volumetric rainfall input \bar{R} for varying perme-
18 able radial extensions r_P with fixed and decreasing tree density (black and gray lines respectively).
19 Panel a) depicts lateral root extension r_R (solid line), radial extension beyond the rooting zone r_{nR}
20 (dotted line), radial extension of the permeable and impervious pavements, r_P and r_I (dot-dashed
21 and dashed lines respectively).
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44 Figure 3 Examples of numerically generated soil moisture time series (a) in absence of irrigation and
45 (c) in presence of micro- and traditional-irrigation, and corresponding probability density functions
46 of soil moisture (b,d). In (a,b) r_P increases from 0 (absence of permeable pavement; solid line)
47 to 2 m (dashed line), to 4 m (dotted line), while $r_I = r_T - (r_P + r_B)$ decreases accordingly. In
48 (c,d) solid line refer to micro-irrigation, dashed line to traditional irrigation (in both cases $r_P = 2$
49 m). Other parameters are $Z_R = 0.5$ m, $n = 0.43$, $T_{max} = 70$ kg d⁻¹ tree⁻¹, $s^* = 0.28$, $s_1 = 0.62$,
50 $\alpha = 12$ mm, $\lambda = 0.2$ d⁻¹, $\Delta = 3$ mm, $\kappa_C = 0.6$, $r_B = 1$ m, $r_T = 7.5$ m, $\eta_P = 0.45$, $\eta_I = 0.1$. In
51 c,d), the irrigation parameters are $\tilde{s} = 0.75s^*$ and $\hat{s} = 0.8s_1$; the atom of probability (solid bar in
52 d) is not to scale.
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4 Figure 4 a-c) Dependence of tree effective cooling capacity with respect to potential, $\langle T \rangle / T_{max}$,
5 and (d-f) dynamic water stress $\bar{\theta}$ (bottom) on planting geometry, species selection, and rainfall
6 pattern: a, d) effect of permeable pavement dimension r_P and tree transpiration requirements
7 T_{max} (r_C varies along with T_{max} from 2.5 to 3.8 m as $r_C = 0.73T_{max}^{\frac{1}{3}}$, where the constant is
8 chosen so that $T_{max} = 70 \text{ kg d}^{-1} \text{ tree}^{-1}$ for a $r_C = 3 \text{ m}$ canopy); b, e) effect of r_P and rainfall
9 frequency and depth, with constant total rainfall over the growing season $R_{tot} = 341 \text{ mm}$ (i.e.,
10 $\alpha = R_{tot}(T_{seas}\lambda)^{-1}$ decreases from 48 to 5 mm as λ increases); and c, f) effect of tree density and
11 extension of permeable and impervious pavement (r_P and r_I respectively). The growing season
12 length is assumed to be $T_{seas} = 142 \text{ d}$. For the definition of dynamic stress, $q = 1$, $k = 1$. All
13 the other non-varying parameters are as in Fig. 3. In (a,d), the thick dashed line represent the
14 permeable pavement extension such that $r_P + r_B = r_C$, while the gray horizontal lines represent
15 the low and high water demanding trees discussed in III.D (solid and dotdashed lines repectively).
16 In (c,f), the thick dashed line indicate the parameter combinations for which $r_T = 7.5 \text{ m}$ (i.e., the
17 case explored in the other panels).
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32 Figure 5 Seasonal water requirements V_t for traditional irrigation as a function of a) permeable
33 pavement dimension r_P and tree transpiration requirements T_{max} , b) r_P and rainfall frequency and
34 depth (with constant $R_{tot} = 341 \text{ mm}$), and c) extension of permeable and impervious pavement, r_P
35 and r_I respectively. In a), the thick dashed line representst the permeable pavement extension such
36 that $r_P + r_B = r_C$, while the gray horizontal lines represent the low and high water demanding
37 trees discussed in III.D (solid and dotdashed lines repectively). In c), the dashed line corresponds
38 to $r_T = 7.5 \text{ m}$ (i.e., to the case explored in the other panels). All the other non-varying parameters
39 are as in Fig. 4.
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49 Figure 6 Required application frequency $\nu^{\downarrow}(\tilde{s})$ for traditional irrigation as a function of a) per-
50 meable pavement lateral extension r_P and tree transpiration requirements T_{max} and b) r_P and
51 rainfall frequency λ (with constant $R_{tot} = 341 \text{ mm}$ and variable α). In a), the thick dashed line
52 represent the permeable pavement extension such that $r_P + r_B = r_C$, while the gray horizontal
53 lines represent the low and high water demanding trees discussed in III.D (solid and dotdashed
54 lines repectively). All the other non-varying parameters are as in Fig. 4.
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Table I Main tree parameters for several species belonging to the most common genera in North American cities (as shown by data presented by Raupp *et al.* (2006) and McPherson and Rowntree (1989)); additional data mainly on transpiration for less common species can be found in Bush *et al.* (2008); Chen *et al.* (2011); Hilaire *et al.* (2008); Litvak *et al.* (2011); McCarthy and Pataki (2010))

Species	City (location) [†]	T_{max} (kg d ⁻¹ tree ⁻¹)	A_C (m ²)	DBH (cm)	Ref
<i>Acer campestre</i> L. ^a	Brno, Czech Rep. (S)	64.6	29.8	22.4	Čermák <i>et al.</i> (2000)
<i>Acer campestre</i> L. ^b	Brno, Czech Rep. (S)	140.7	58.9	40.3	Čermák <i>et al.</i> (2000)
<i>Fraxinus pennsylvanica</i>	Lincoln, NE (LP)	49.4 ^c	17.5 ± 1.36	15.3 ± 0.49	Cregg (1995)
<i>Fraxinus pennsylvanica</i>	Lincoln, NE (SP)	40.8 ^c	18.4 ± 1.24	13.5 ± 0.48	Cregg (1995)
<i>Fraxinus pennsylvanica</i>	Lincoln, NE, Campus (P)	70.1 ^c	22.4 ± 4.14	17.4 ± 1.63	Cregg (1995)
<i>Fraxinus pennsylvanica</i>	Saint Paul (P)	69.1 ± 25.8	75.9 ± 11.14	38.6 ± 11.4	Peters <i>et al.</i> (2010)
<i>Fraxinus pennsylvanica</i>	CTC Minneapolis, MN (P)	92.5 ± 34.5	101.6 ± 59.8	41.1 ± 2.0	Peters <i>et al.</i> (2010)
<i>Gleditsia tricanthos</i> ^d	Los Angeles Arboretum, CA (P)	89.9 ± 23.6	-	45.2 ± 5.5	Pataki <i>et al.</i> (2011b)
<i>Quercus rubra</i>	CTC Minneapolis, MN (P)	-	85.1 ± 30.1	42.9 ± 7.7	Peters <i>et al.</i> (2010)
<i>Ulmus parvifolia</i> ^d	Los Angeles Police Ac., CA (P)	67.7 ± 15.8	-	28.9 ± 2.5	Pataki <i>et al.</i> (2011b)
<i>Ulmus pumila</i>	Lauderdale Saint Paul, MN (P)	107.1 ± 38.1	90.8 ± 70.6	50.8 ± 12.6	Peters <i>et al.</i> (2010)
<i>Ulmus thomasii</i>	Saint Paul, MN (P)	80.4 ± 28.6	68.1 ± 36.3	44.4 ± 7.0	Peters <i>et al.</i> (2010)

[†] Location is P for parks and other extensively green areas, S for trees growing along streets or isolated in parking lots, LP or SP for large or small planter respectively. ^a Shaded tree with nearby sidewalk, with $r_R=3.6$ m; ^b Exposed tree with garden, with $r_R=5.3$ m; ^c Originally expressed in mmol m⁻² s⁻¹ and transformed in kg d⁻¹ tree⁻¹ assuming a 12-hr day and a leaf area index of 4 (in agreement with Cregg (1995)); ^d Irrigated

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Figure 1

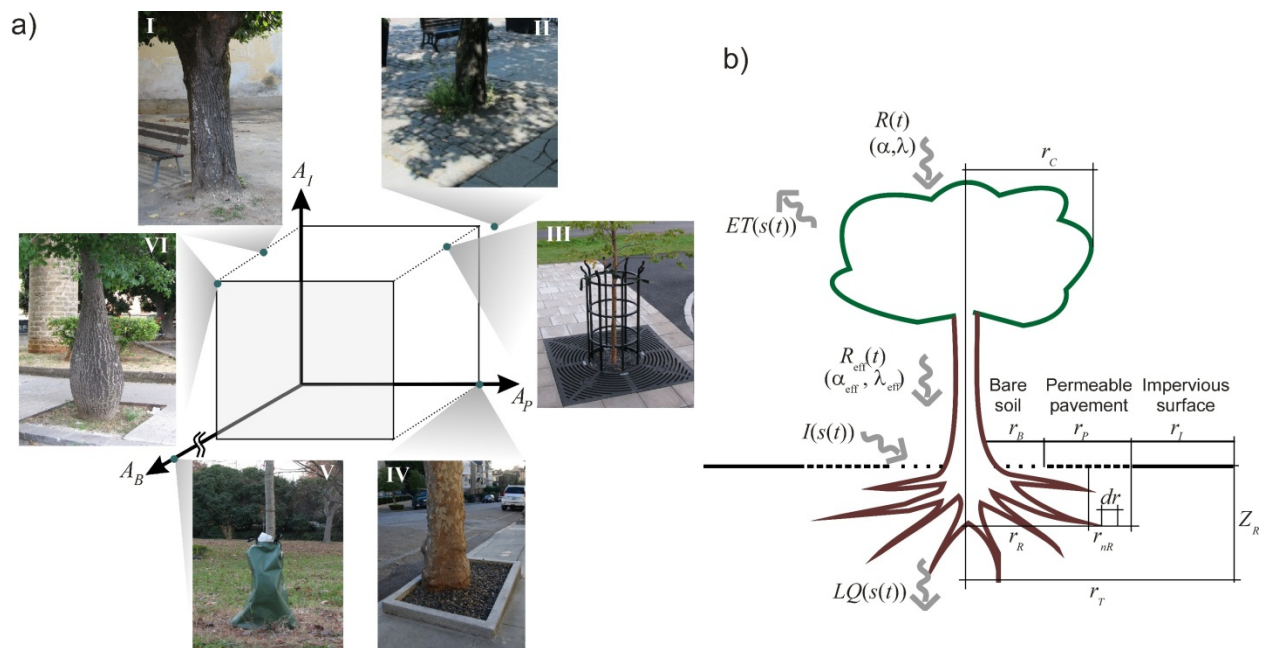
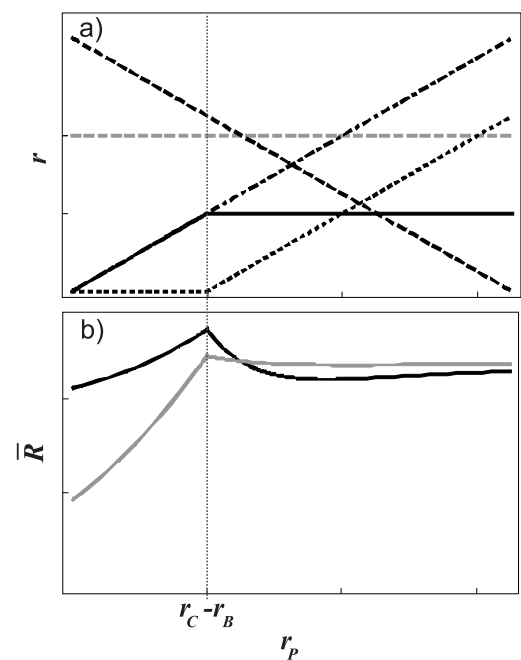


Figure 2



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Figure 3

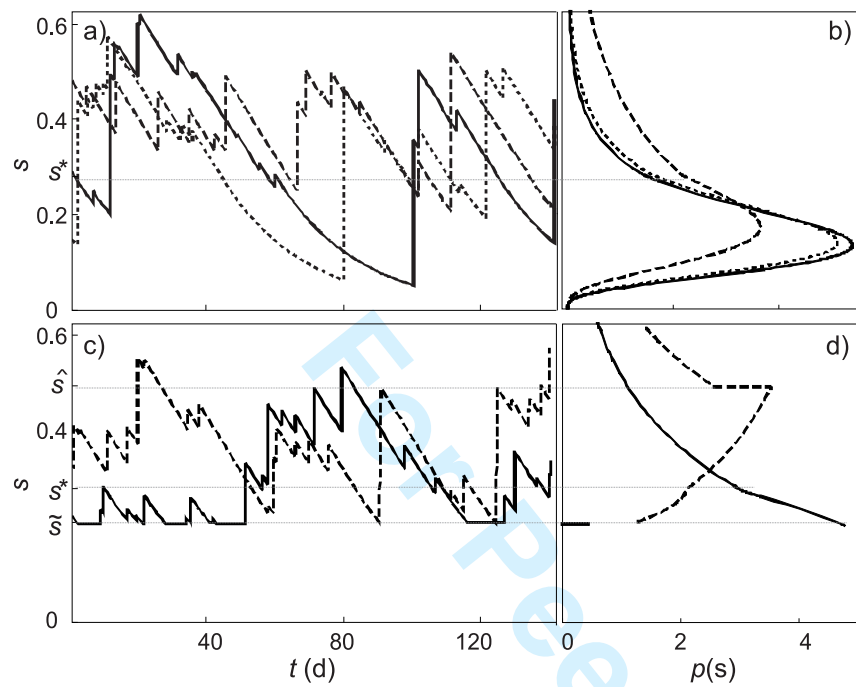


Figure 4

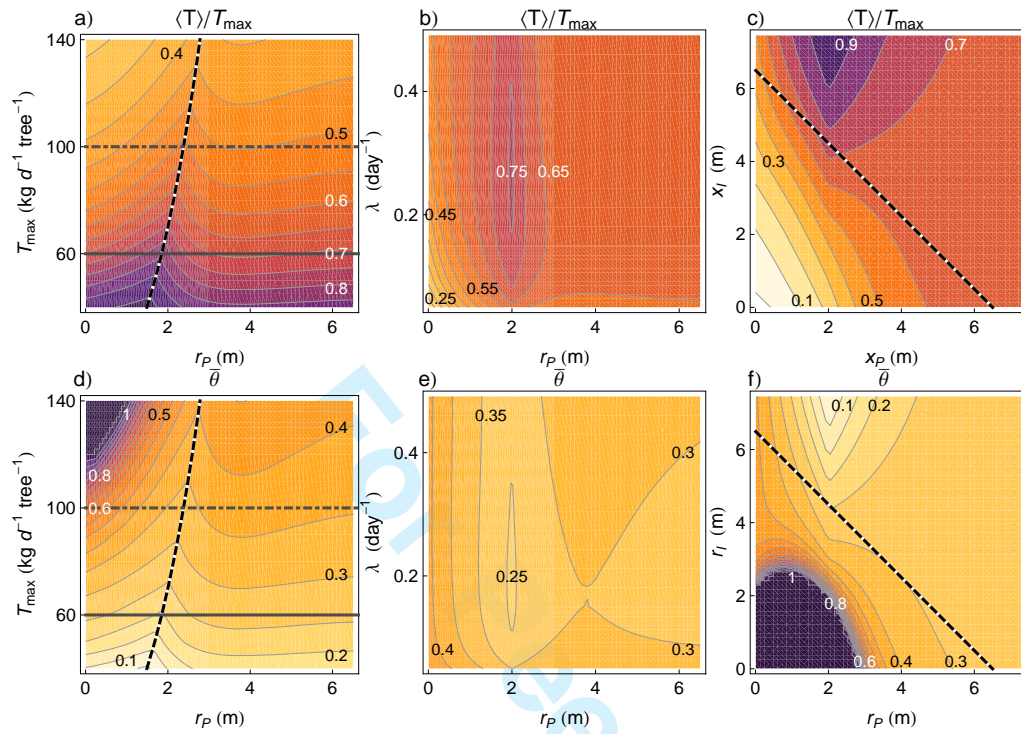


Figure 5

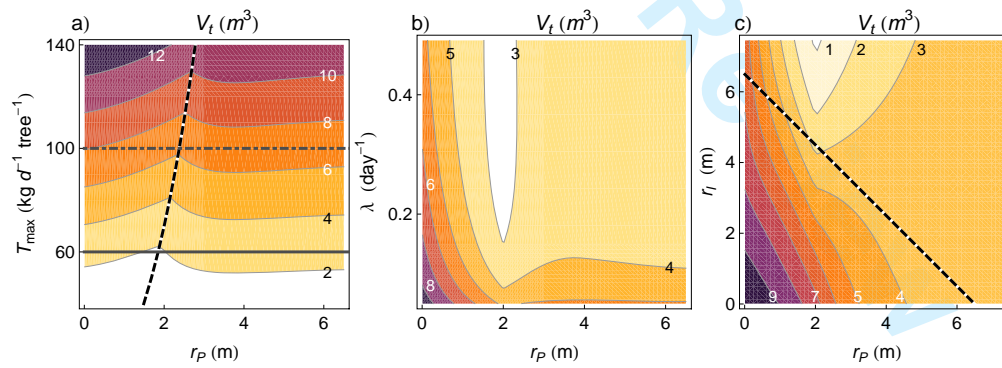
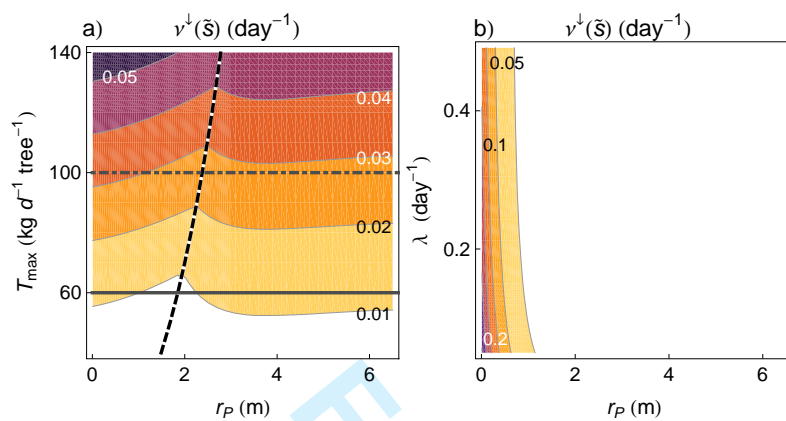


Figure 6



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