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Scale effect of anaerobic digestion tests in fed-batch and semi-continuous mode for the technical and economic feasibility of a full scale digester / Ruffino, Barbara; Fiore, Silvia; Roati, Chiara; Campo, Giuseppe; Novarino, Daniel; Zanetti, Mariachiara. - In: BIORESOURCE TECHNOLOGY. - ISSN 0960-8524. - STAMPA. - 182:(2015), pp. 302-313. [10.1016/j.biortech.2015.02.021]

Availability: This version is available at: 11583/2592388 since: 2016-04-04T10:47:10Z

Publisher: Elsevier

Published DOI:10.1016/j.biortech.2015.02.021

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Accepted Manuscript

Scale effect of anaerobic digestion tests in fed-batch and semi-continuous mode for the technical and economic feasibility of a full scale digester

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PII: DOI: Reference:	S0960-8524(15)00190-X http://dx.doi.org/10.1016/j.biortech.2015.02.021 BITE 14593
To appear in:	Bioresource Technology
Received Date:	28 November 2014
Revised Date:	2 February 2015
Accepted Date:	7 February 2015



Please cite this article as: Ruffino, B., Fiore, S., Roati, C., Campo, G., Novarino, D., Zanetti, M., Scale effect of anaerobic digestion tests in fed-batch and semi-continuous mode for the technical and economic feasibility of a full scale digester, *Bioresource Technology* (2015), doi: http://dx.doi.org/10.1016/j.biortech.2015.02.021

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1	SCALE EFFECT OF ANAEROBIC DIGESTION TESTS IN FED-BATCH AND SEMI-
2	CONTINUOUS MODE FOR THE TECHNICAL AND ECONOMIC FEASIBILITY OF A
3	FULL SCALE DIGESTER
4	
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24 Abstract

25 Methane production capacity in mesophilic conditions of waste from two food industry plants was 26 assessed in a semi-pilot (6L, fed-batch) and pilot (300L, semi-continuous) scale. This was carried 27 out in order to evaluate the convenience of producing heat and electricity in a full scale anaerobic digester. The pilot test was performed in order to obtain more reliable results for the design of the 28 29 digester. Methane yield, returned from the pilot scale test, was approximately 80% of that from the 30 smaller scale test. This outcome was in line with those from other studies performed in different 31 scales and modes and indicates the success of the pilot scale test. The net electricity produced from 32 the digester accounted for 30% to 50% of the food industry plants' consumption. The available 33 thermal energy could cover from 10% to 100% of the plant requirements, depending on the energy MA demand of the processes performed. 34 35 36 Keywords Mesophilic anaerobic digestion 37

38 Fed-batch tests

39 Semi-continuous tests

40 Vegetable processing waste

41 Cost-benefit analysis

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42 **1. INTRODUCTION**

Italian food industry, with a sales volume of 130 G€, is the second most important industry after the 43 44 car manufacturing. The Italian food industry plants buy and transform 72% of domestic agricultural raw materials. Inevitably, the processes performed in food industries generate huge amounts of 45 agro-industrial waste. For Italy, Petruccioli et al. (2011) reported productions of 2.4 Mt/y of grape 46 pomace, 0.7 Mt/y of olive pomace and 0.2 Mt/y of both tomato pomace and soybean integuments. It 47 can be estimated that the transformation of 1 t of agricultural raw materials generates from 30 to 48 49 100 kg of organic waste. Agro-industrial waste, that includes both fruit and vegetable waste, may be stabilized by the 50 51 anaerobic digestion (AD) process (Khalid et al., 2011). AD is known as a more environmentally friendly and energy saving process for stabilizing high-strength organic waste, than other disposal 52 53 options like landfilling, incineration, and composting (Hosseini Koupaie et al., 2014; Traversi et al., 2013). In addition to biogas generation, AD benefits include enhancing nutrient characteristics of 54 the dewatered digestate used as fertilizers as well as pathogens reduction. Greenhouse emissions are 55 also decreased because the AD process diverts organic waste from landfills, thus preventing 56 uncontrolled methane and carbon dioxide emissions from decomposition (Xie et al., 2011). 57 58 In spite of the potential advantages of the AD process for the management of agro-industrial waste, at present, in Italy, most of it is sent to composting plants. This work aims to carry out a technical 59 60 and economic assessment of a more sustainable way of managing waste from food industry 61 processes, by evaluating the convenience of producing heat and electricity in an on-purpose made 62 anaerobic digester. The waste products considered in this work are generated in two food industry 63 plants located in the NW Italy.

64 In order to proceed to the technical and economic assessment of the digester, the potential

65 production of methane from the waste generated in the two plants was determined in both a semi-

66 pilot (6-L, fed-batch mode) and pilot (300-L, semi-continuous mode) scale in mesophilic

67 conditions. Values of the methane yield for waste originated from agro-industrial processes are

reported in several studies (Dareioti and Kornaros, 2014; Garcia-Peña et al., 2011; Jiang et al., 68 69 2012; Yang et al., 2013). However, only a couple of very recent papers focus on tests carried out on 70 fruit and vegetable waste fed as a single substrate to semi-continuous pilot scale reactors (Fiore et 71 al., 2013; Scano et al., 2014). In order to test the capacity of waste to produce methane in anaerobic conditions, different modes 72 73 (batch, semi-continuous, continuous) and several scales (from lab to full scale) can be used. 74 Traditional biochemical methane potential (BMP) tests have the undeniable advantage of taking up 75 small volumes (usually from 100 mL to 1,000 mL) and requiring very low efforts of personnel 76 during test developments (Angelidaki et al., 2009; Cavaleiro et al., 2013; Lisboa and Lansing, 2013; 77 Raposo et al., 2011; Triolo et al., 2012). However, the procedure does not assure reliable outcomes in the case of highly heterogeneous substrates and it is not suitable to highlight problems such as the 78 irreversible acidification of the reactor (Kolbl et al., 2014). 79 In this work a first estimate of the methane production capacity of waste was obtained using fed-80 batch tests in a semi-pilot scale. The authors preferred fed-batch tests to traditional one-fed essays 81 and employed larger vessels than usual (6 L instead of maximum 1 - 1.5 L) in order to test a 82 substrate with high heterogeneity using a solid procedure. The fed-batch mode allowed the authors 83 84 to feed higher amounts of substrate, guaranteeing the representativeness of the whole sample and avoiding possible inhibition in the methanogenesis phase due to the progressive acidification of the 85 86 reactor. In addition, because of the effect of particle size and surface area of the substrate on the performance of the AD process (Novarino and Zanetti, 2012; Palmwoski and Muller, 2000; 2003; 87 88 Zhang and Banks, 2013), the larger volume allowed the authors to test highly heterogeneous 89 substrates with realistic particle sizes. 90 The capacity of waste to produce methane obtained in the fed-batch mode was verified using a

91 semi-continuous test carried out in the 300-L digester. The pilot scale test was necessary in order to
92 obtain more reliable results on which to base the design of the full scale digester for the energy

93 valorization of the waste products. The results of tests in the pilot scale are of particular interest

- 94 especially for highly heterogeneous substrates. The outcomes of the experimentation were
- employed for a preliminary process and cost-benefit analysis of the full-scale AD digester for
- 96 treating the waste produced in the two food industry plants.
- 97

98 2. MATERIALS AND METHODS

99 2.1 Waste origin and characterization

100 Waste products considered in this work are generated in two food industry plants located in the NW 101 Italy. The first plant (Plant 1) transforms and preserves several kinds of vegetables, mainly carrots, 102 potatoes, onions, beetroots and celery with a capacity of 12,000 t/y. Processes carried out in Plant 1 103 generate around 350-400 t/y of organic waste. The organic waste from Plant 1 has a very high 104 heterogeneity, because it comes from both the peeling process of vegetables and grating and cloth-105 filtering phases in the wastewater treatment plant (WWTP) contained within the food industry plant. 106 The second food industry plant (Plant 2) specializes in the production of pesto sauce and other kinds 107 of sauces with an end product capacity of 5,000 t/y. Waste is made of residues of production and 108 sub-standard products, the main components of which are basil and sunflower oil. The monthly amount of organic waste is approximately of 25-30 t. Plants with a production capacity similar to 109 110 that of the plants studied, have electrical consumption in the order of 1,000 - 1,500 MWh/y and thermal consumption that ranges from $5 \cdot 10^2$ to 10^4 MWh/y, depending on the heat demand of the 111 112 operations performed.

Two substrates coming from Plant 1 and Plant 2 were sampled, characterized and employed in the AD tests. The substrate from Plant 1 was a mixture of three waste products. The first waste product (named as "peeling") originated from the peeling process of vegetables. The second waste product (named as "screen") was separated from the grid screen positioned at the head of the WWTP contained within the food industry plant. The third waste product (named as "filtrate") was separated from the cloth filter positioned after the grid screen in the same WWTP. The produced

119 amounts of the three waste products, on dry basis, were the same. For this reason the mix to employ 120 for the tests was obtained by mixing identical amounts of each waste product (on dry basis). 121 The substrate from Plant 2 came from the production process of pesto sauce. According to the data supplied from the Plant, the principal components of this waste were basil and sunflower oil, in 122 123 approximately the same amount. Minor components were cashew nuts, pine nuts and Parmesan 124 cheese. 125 The physical and chemical characterization of substrates and the digestate, that resulted from the 126 digestibility tests, included pH, total solid (TS) and volatile solid (VS) content, and elemental 127 composition (C, H, N, S – this last analysis only for substrates). All parameters were determined 128 according to standard methods (APHA, AWWA, WEF, 2005). The elemental analysis was 129 performed by means of a CHNS-O Thermo Fisher Flash 2000 Analyzer EA 1112, assuming the oxygen content as the complementary fraction towards C, H, N, S contents. All the analyses were 130 131 carried out in three replicates, on representative and significant amounts of the samples as in Roati 132 et al. (2012). Determination of VS content and elemental composition were carried out on samples 133 dried at 105°C.

134 **2.2 Semi-pilot scale test – fed-batch mode**

135 Both substrates, that is the vegetable mix waste (VMW) and the waste from pesto sauce production (PSW, pesto sauce waste), were tested in a fed-batch mode. Tests were performed in mesophilic 136 137 conditions (35°C) in five replicates, using 6-L poly methyl methacrylate (PMMA) digesters placed 138 in a thermostatic bath. The anaerobic environment was prepared by filling digesters with water, then 139 replacing it with nitrogen. This procedure also ensures that the reactors were leak free. In order to 140 simulate real digestion conditions, pH value was not adjusted and no nutrients were added. Each 141 digester was manually mixed for 20–30 s once a day as in Ruffino et al. (2015). 142 The inoculums employed in the tests was prepared from 1,000 mL of digestate collected from the 143 anaerobic digesters of a municipal WWTP located in the same area of the two food industry plants.

144 The inoculums was progressively fed with amounts of primary sludge, coming from the same

145	WWTP, to reach a final volume of 3,000 mL. The inoculums was considered ready when its daily
146	biogas production was of less than 1% of the overall production recorded throughout the period of
147	preparation.
148	For the test with the VMW, the inoculums had a TS content of $3.26 \pm 0.31\%$, a VS/TS ratio of 64.3
149	$\pm 2.9\%$ and a pH value of 7.23 ± 0.01 (average value on five replicates \pm standard deviation). Each
150	digester was fed with six aliquots of the substrate. Each aliquot contained 15 g of TS and was
151	introduced into the digesters at days 0, 1, 2, 6, 8 and 9 from the beginning of the test. The final TS
152	content was 3% of the working volume (3 L).
153	For the test with the PSW, the inoculums had a TS content of 4.24 \pm 0.21%, a VS/TS ratio of 75.1 \pm
154	0.6% and a pH value of 7.18 \pm 0.01 (average value on five replicates \pm standard deviation). Each
155	digester was fed at days 0, 1, 5, 7, 8 and 12 with six aliquots of the substrate, each of them
156	contained 14 g of TS. Differences in fed frequency between test 1 and 2 were due to the limitation
157	in the maximum daily biogas collection capacity (10-12 L) and depended on the biogas production
158	rate. However, also for Test 2, the final TS content in the digester was approximately 3%.
159	For each replicate, the produced biogas was collected in two 5-L Tedlar® bags connected in
160	parallel. The characterization and measure of the biogas volume was carried out daily, throughout
161	the entire duration of the test. The characterization, that is the volumetric composition of the biogas
162	in terms of CH ₄ , CO ₂ , O ₂ and "balance" (the fraction made up of gases that are different from the
163	first three, that is mainly nitrogen, hydrogen, carbon monoxide and volatile organic compounds),
164	was obtained by flushing 500 mL of biogas through a biogas analyzer (Biogas Check, Geotechnical
165	Instruments Ltd). The volume was measured by replacing volumes of water with biogas and
166	referring the obtained value to standard conditions.
167	2.3 Pilot scale test – semi-continuous mode

168 The pilot scale test involved only the VMW from Plant 1. This substrate was chosen because of its

- 169 higher heterogeneity compared to the other waste.
- 170 The AD plant employed for the test was equipped with a 300-L reactor (240 L working volume), an

171 80-L gasometer and a system for on-line monitoring of the biogas composition. The temperature 172 was automatically monitored and regulated by means of a resistance temperature sensor (Pt 100) 173 positioned along the reactor and connected to the heating system. The biogas collection system included a condensation trap for moisture removal and a flow-meter. The pilot plant was controlled 174 175 by a PLC system that receives signal from the different sensors and drives the main electrical and 176 pneumatic pieces of equipment. Substrates inside the digester were mixed 15 minutes every hour by 177 a system of re-circulating biogas. The other physical and chemical parameters required for an 178 effective monitoring of the AD process (TS, VS content, pH and FOS-TAC values) were daily 179 determined on the sampled digestate. 180 The inoculums for the test was prepared by mixing 80 L of digestate, collected from the digesters of the WWTP as in Section 2.2, and 160 L of primary sludge from the same plant. The inoculums was 181 182 considered ready to start the test when its daily biogas production was of less than 1% of the overall 183 volume produced throughout the period of preparation. At the beginning of the semi-continuous test the inoculums had a TS content of $2.22 \pm 0.01\%$, a VS/TS ratio of 62.7 ± 0.2 and a pH value of 7.4 184 ± 0.1 (average value on 3 replicates \pm standard deviation). 185 The TS content of the substrate fed to the reactor was fixed to 3%, in order to make possible a 186 comparison between the performances of the two digestion systems (semi-pilot and pilot scale). For 187 the same reason the test was carried out in mesophilic conditions (35°C). The hydraulic retention 188 189 time (HRT) was fixed to 30 days. This value came from the results (trend of biogas and methane 190 production and hydrolysis constant) obtained for the VMW tested in a fed-batch modality, see 191 Section 3.2. 192 The volume of the fed substrate was of 8 L with a TS content of 240 g and resulting OLR of 1 kgTS/m³·d or 0.87 kgVS/m³·d. Due to the characteristics of the waste product (that is an average 193 194 TS content of approximately 8% and a VS/TS ratio of 86.8%, see Section 3.1), the daily feedstock 195 was prepared by mixing 2,200 g of the VMW with 5,800 mL of tap water. The AD test had an 196 overall duration of 54 days that were necessary for the loading of 30 volumes of substrate and the

- subsequent complete depletion of the biogas produced. The feeding of the substrate and the
- 198 extraction of an equal volume of digestate were carried out in days from Monday to Friday,
- according to the procedure described as in the follow:
- 200 1. recording of volume and composition (methane, CO_2 , oxygen, balance and H_2S) of the biogas
- 201 produced since the previous recording;
- 202 2. extraction of 8 L of digestate to undergo to pH, TS, VS, FOS-TAC analysis;
- 203 3. loading of 8 L of substrate, prepared as previously described.
- 204
- **3. RESULTS AND DISCUSSION**

206 **3.1 Waste characterization**

- 207 The characteristics of the substrates involved in the digestibility tests are shown in Table 1.
- 208 The VMW had an average TS content of less than 10% b.w. (by weight), a VS/TS ratio close to
- 209 90%, an acidic pH value and a elemental composition of C: 45% b.w., H: 6-7% b.w., N: 2-3% b.w.,

19

- 210 S: 0.3-0.5% b.w., on dry basis, with a subsequent C/N ratio of approximately 15. A visual test
- revealed that the sizes of particles ranged from few millimeters (mainly residues from the peeling
- 212 process) to about 5 centimeters (pieces of vegetables and discarded mushroom caps).

213 The PSW was far more homogeneous than VMW. It had the texture of a sauce with a TS content of

- approximately 70% b.w. and a VS/TS ratio close to 100%. pH value was determined only with pH
- strips because of the high oily content of the waste product. For the same reason the determination
- of the elemental composition was not performed.
- 217 The molecular formula of the first substrate was obtained from its elemental composition. The
- 218 molecular formula of the second substrate was obtained from the average composition of a general
- vegetable ($C_{18}H_{30}O_{13.5}N$, as for the first substrate), that simulates basil, and sunflower oil
- 220 $(C_{17.9}H_{32.8}O_2)$. The ratio between basil, as dry matter, and sunflower oil was assumed equal to 1:1
- 221 b.w.
- 222

223 **3.2 Semi-pilot scale test**

Both fed-batch tests performed on the two waste products had an overall duration of 22 days.

- According to 1% criterion of German Guideline VDI 4630 (2006), the tests were stopped when the
- 226 daily biogas production dropped to values of less than 1% of the total gas production.
- 227 The daily evolution of the biogas produced by the VMW is shown in Figure 1. Figure 1a shows two

peaks at days 2-3 and 9-10 from the beginning of the test. In those days the daily biogas production

- rose to volumes respectively of 4.5 and 5.5 NL. The trend of the daily biogas production showed a
- strong dependence on the feeding rate. The total amount of the fed waste (90 g of dry substance in 6
- aliquots) determined an overall biogas production that was between 40 and 47 NL. The five
- replicates returned a biogas specific production of 0.554 ± 0.038 Nm³/kgVS added (average value)
- 233 on five replicates \pm standard deviation), a methane specific production of 0.294 \pm 0.029 Nm³

234 CH₄/kgVS added, with a subsequent average methane content of $53.0 \pm 1.8\%$ b.v. (by volume). The

composition of the biogas produced in the days in which the record of volume and composition was

carried out, is shown in Figure 1b. The average carbon dioxide content was of $32.6 \pm 1.5\%$. Biogas

did not show any traces of H_2S . The daily pH check did not highlight critical situations such as the

238 irreversible deviation of pH value toward acidic conditions (data not showed).

The digestate that resulted from the test carried out on the VMW had a TS content of $3.21 \pm 0.14\%$, a VS/TS ratio of $66.2 \pm 1.6\%$ and a pH value of 7.10 ± 0.02 . A TS depletion of 67.5% and a VS depletion of approximately 73% were calculated from the balance that involved the inoculums, the digestate and the fed substrate (1046.5 g wet substrate with a TS content of 8.60\%, 90 g TS, 78.1 g

243 VS).

With reference to the second substrate (PSW), the daily evolution of the biogas produced is shown in Figure 2. Figure 2a shows two peaks at days 10 and 13 from the beginning of the test. In those days the daily biogas production rose to volumes respectively of 9.5 and 8 NL. Unlike the case described above, the trend of the daily biogas production for the PSW did not show a strong

248 dependence on the feeding rate. The total amount of the fed substrate (84 g of dry substance in 6

aliquots) determined an overall biogas production that was between 82 and 92 NL. The five

- replicates of the semi-pilot scale test returned a biogas specific production of 1.08 ± 0.05
- 1251 Nm³/kgVS added (average value on five replicates ± standard deviation), a methane specific

production of 0.722 ± 0.033 Nm³ CH₄/kgVS added, with a subsequent average methane content of

- $66.9 \pm 0.2\%$ b.v. The composition of the biogas produced in the days in which the record of volume
- and composition was carried out is shown in Figure 2b. The average carbon dioxide content was of
- 255 29.8 \pm 2.1%. As for the previous substrate, the biogas produced did not show any traces of H₂S and
- the daily pH check did not highlight critical situations (data not showed).
- 257 With reference to the second tested substrate, the digestate had a TS content of $3.88 \pm 0.21\%$, a
- VS/TS ratio of 72.7 \pm 0.9% and a pH value of 7.24 \pm 0.02. A TS and VS depletion close to 100%
- was calculated from the balance that involved the inoculums, the digestate and the fed substrate
- 260 (121.1 g wet substrate with a TS content of 69.3%, 84 g TS, 82.3 g VS).
- A balance that involves carbon was carried out for the two substrates in order to compare the
- amount of carbon in the feedstock with the amount of carbon that was converted to methane or
- carbon dioxide (average amount on the five replicates). The results are detailed in Table 2. It can be
- seen that the AD process performed in a fed-batch mode converted 56% of the carbon from the
- VMW and 92% of the carbon from the PSW to methane and carbon dioxide. These results are in
- line with the results that concern TS and VS depletion.

267 In order to compare the obtained specific productions of biogas and methane with the maximum

- 268 production predicted by theoretical calculations, according to Buswell and Neave model
- 269 (Tchobanoglous et al., 1993), it is necessary to refer to the molecular formula of the two substrates,
- as reported in Table 1. According to this model, the VMW subjected to AD would give rise to a
- biogas specific production of 0.848 Nm³/kgVS added and a methane specific production of 0.424
- 272 Nm³/kgVS added (as listed in Table 3). The observed values were respectively only 65.3% and
- 69.3% of the values predicted by Buswell and Neave model. On the other hand, the theoretical
- biogas specific production of the PSW was of 1.14 Nm³/kgVS added, 5% more than the value

275 returned from the fed-batch test.

276 With reference to the afore-described results, the most important differences in the behavior of the

two substrates (VMW and PSW) were observed:

- in the specific production values of biogas and methane. The biogas specific production of the
- 279 VMW substrate was approximately 50% of that of the PSW substrate. On the other hand, the

280 methane specific production of the VMW was 40% of that of the PSW;

- in the response of the AD systems to the feedstock introduction. Reactors fed with the VMW
- substrate showed a fast production of biogas and methane. Reactors fed with the PSW
- substrate showed a production of biogas and methane more intense than that from the VMW,
- but only after a phase of initial delay;
- in the conversion rate of the fed carbon to methane and carbon dioxide, that was in the order of
 56% for the VMW substrate and greater than 90% for the PSW substrate;
- in the ratio between the biogas and methane specific production values obtained from the fed-

batch tests and those from Buswell and Neave model.

289 The afore-mentioned differences could be attributed to the composition and particle size distribution

290 of the two substrates. The VMW substrate was composed solely of vegetable matrix, rich in fiber

and lignin, therefore refractory to biological degradation. Due to the structure and particle sizes of

the VMW substrate, the fraction available for biological degradation must first be subjected to the

processes of disintegration and hydrolysis. It is therefore unlikely that an accumulation of reactionintermediates occurs and slows the course of the AD process.

295 On the other hand, the second substrate (PSW) showed a considerable oil content and very small

296 particle sizes. Fatty acids from oil and soluble COD liberated from the vegetable matrix, following

- the size reduction occurred in pesto sauce preparation, were readily available to AD. However,
- according to Batsone et al., 2002, the maximum removal rate of fatty acids $(6 d^{-1})$ is one order of
- 299 magnitude lower than that of sugars and amino acids (the simplest substances that compose
- 300 vegetable matrices, respectively equal to 30 and 50 d^{-1}). This may explain the delay observed in the

301 first phase of the test, that was due to the low removal rate of fatty acids. Moreover, the 302 composition of the second substrate may explain the higher values of specific production of biogas 303 and methane, compared to those observed for the VMW, and the high ratio between the experimental value and the theoretical value. These last observations came from the higher presence 304 305 of completely degradable substances (fatty acids, and available amino acids and sugars from the 306 vegetable matrix) in the PSW substrate than in the VMW. 307 The characterization of waste products from Plant 1 and Plant 2 was completed through the 308 evaluation of the disintegration rate constant (Batstone et al., 2002). The disintegration rate constant 309 describes the kinetics of the first of the two processes (i.e. disintegration and hydrolysis) that make 310 up the solubilization phase. In fact, according to Batstone et al., 2002, the disintegration process is commonly the limiting phase of the whole AD process when there is no accumulation of 311 intermediary products. According to Angelidaki et al., 2009, using the first part of the experimental 312 curve build for the determination of the ultimate methane production of a given substrate, it is 313 possible to define the constant for a first order model as in Equation 1: 314

 $315 \quad ln\frac{B_{\infty}-B}{B_{\infty}} = -kt \quad (1)$

where B_{∞} is the value of the ultimate biogas production and B is the biogas production at a given time, t. The method described as in Angelidaki et al., 2009 for the interpretation of results from BMP tests in a batch mode, was here adapted for the results obtained from tests carried out in a fedbatch mode. The parameter value for k was estimated using non linear least squares curve fitting of the net cumulative methane production (see Figure 3).

Curve fitting returned values of 0.382 d⁻¹ and 0.254 d⁻¹ respectively for the VMW and the PSW. However, as it can be seen in Figure 3, the model employed is suitable for describing the trend of the biogas production from the VMW, but it demonstrated not to be adequate for the PSW. In this second case, the disintegration process cannot be considered the limiting phase of the overall AD process because of the high amount of organic acids from sunflower oil. A possible accumulation of

organic acids in the first days of the test could have made the AD process slower than predicted bythe model.

The values found for the two substrates are in line with the values of the same parameter found for organic substrates that originates from vegetables, as reported in Galí et al., 2009; Giuliano et al., 2013; Shi et al., 2014. In those works the authors found k values that ranged from 0.2 to 0.4 d⁻¹ and Giuliano et al., 2013 reported values of 0.26 d⁻¹ and 0.34 d⁻¹ respectively for potatoes and onions when BMP tests were carried out in mesophilic conditions.

333 **3.3 Pilot scale test**

Only the VMW was involved in the pilot scale test. The daily production of biogas and the daily 334 335 specific production of biogas and methane are shown in Figure 4a. Values of production of the days 336 in which the recording of data (volume, biogas composition, see Figure 4b) was not performed (Saturdays, Sundays), were obtained dividing the value recorded the first day after the break by the 337 duration of the break. The diagram of Figure 4a shows that the biogas daily production had a 338 339 fluctuating trend, that ranged from 50 NL/d to 100 NL/d. The lowest values were registered in the two days that ensued the feeding stops. The highest values were registered when the loading of the 340 substrate was made. The value of the methane specific production, after showing considerable 341 fluctuations until around day 20 from the beginning of the test, stabilized at 0.223 Nm³/kgVS 342 343 added. The values of the two parameters FOS and TAC showed a trend that decreased from the 344 beginning of the test to the end. However, the ratio between FOS and TAC kept constant in the 345 range 0.2 - 0.3 (see Figure 5a), optimum for an AD process. The daily record of the TS content into 346 the digestate showed a considerable decrease in the first ten days from the beginning of the test, 347 and, then, a subsequent fluctuation around the constant value of 1.50% from day 10 (see Figure 5b). 348 At the end of the test the amount of TS removed was of 55% and the amount of VS removed in the 349 order of 65%. This value was only slightly lower than that found in the fed-batch test (73%). As 350 shown in Table 3 the methane specific production obtained in the semi-continuous test was

approximately 76% of that obtained in the fed-batch test and the difference in the average methanecontent between the two tests was of less than 20%.

353 The ratio between the methane specific production returned from the two tests (batch vs. semicontinuous) performed in this study was compared with that from other studies in order to verify its 354 355 consistency. Unfortunately, to the knowledge of the Authors, very few studies report the results of 356 tests carried out on different scales and modes on the same substrate. Moreover, some of them were 357 carried out on very small scale (Kafle and Kim, 2013; Zhang et al., 2013). In the cases reported and 358 discussed as in the follow, the amount of dry substrate fed to both batch and semi-continuous (or 359 continuous) reactors was similar, so as to make possible a comparison on the digestion 360 performances between the different modes and scales. Zhang et al. (2013) performed tests in both batch and semi-continuous modes employing reactors 361 with the identical active digestion volume of 1 L. The aim of their study was to test a dual solid-362 liquid system for the digestion in mesophilic conditions of food waste that contained considerable 363 amounts of organic substance in the liquid phase. The methane yields obtained in semi-continuous 364 mode was between 60% and 80% of the values obtained in the batch mode. 365 Kafle and Kim (2013) tested mixtures of apple waste and swine manure in both batch (1.8 L liquid 366 volume) and continuous modes (4.5 L liquid volume). The outcomes of the tests carried out in the 367 two modes on a mixture that contained 33% apple waste were compared. The methane yield from 368 369 the continuous test increased, from the first to the second phase of the experimentation, from 182 to 241 mL CH₄/gTCOD added. The methane yield registered in the second phase was approximately 370 371 90% of the value returned from the batch test (267 mL CH₄/gTCOD added). 372 Zupanĉiĉ and Jemec (2010) evaluated the capacity of several kinds of tannery waste to producing 373 methane, in both mesophilic and thermophilic conditions, using BMP tests carried out in 1,160 mL 374 glass vessels. Tests were repeated in semi-continuous mode in a 30-L digester obtaining specific 375 methane productivity values between 79 and 93% of those obtained in the BMP tests. The ratio 376 between the volume of the digester employed for the semi-continuous tests and that of the reactors

377	for the batch tests was approximately of 30, a value of the same order of magnitude of that of this
378	study (80). For this reason, in spite of the kind of the tested substrates (fleshings, skin trimmings
379	and wastewater sludge), the results of the present study have to be considered consisted with those
380	by Zupanĉiĉ and Jemec.
381	Wang et al. (2014) investigated the anaerobic digestion performances of kitchen waste and
382	fruit/vegetable waste from lab-scale (working volume of 1.5 L) to pilot-scale (2 and 4 m^3),
383	including batch, single-phase and two-phase experiments. The methane yields obtained on a pilot-
384	scale (0.64 LCH ₄ /gVS) was about 80% of that obtained on a lab-scale (0.725 L CH ₄ /gVS) under
385	similar VS loading values.
386	Moreover, a study performed in the same apparatus of the present study and in the same operating
387	conditions, on a mix of rice hull and rice bran, returned a ratio of approximately 0.81 between the
388	methane specific production obtained on a semi-continuous mode (0.312 Nm ³ /kgVS added) and the
389	methane specific production obtained on a fed-batch mode (0.386 Nm ³ /kgVS added).
390	On the contrary, Jard et al. (2012) obtained results which are in contrast with those of the above-
391	discussed studies. They tested the biogas production of two kinds of microalgae (Palmaria Palmata
392	and Saccharina Latissima) in both batch (500 mL) and semi-continuous reactors (3 L). Methane
393	yields obtained in the semi-continuous reactor were around 25% higher than those from the BMP
394	essays. According to the authors, AD typically produces higher yields of methane when run as a
395	continuous rather than a batch process due to the progressive adaptation of the microbial
396	population.
397	In spite of the results of this last case, usually batch tests returned methane yields higher than those
398	obtained in a semi-continuous mode. This is obviously due to the longer contact time between
399	substrate and microbial population, that allows microorganisms to completely degrade substrates,
400	and to the better mixing conditions achieved in batch reactors (especially if of small size). It can be
401	concluded that tests in a batch (or fed-batch) mode may provide a good estimate of the capacity of a
402	substrate to be degraded in anaerobic conditions and are useful to assess if further tests with higher-

403 scale reactors may be justified. Batch tests performed on a scale higher than usual assure reliable 404 outcomes in the case of high heterogeneity of the substrate, thus guaranteeing the representativeness 405 of the whole sample and avoiding the necessity of reducing sizes. However, the assessment of a full 406 scale reactor for waste energy valorization requires solid methane yield values that can be obtained 407 only from a pilot scale test. In agreement with previous literature studies, the outcomes of this work 408 demonstrated that a successful pilot scale test returned methane yields between 70 and 80% of the 409 values from batch tests.

410 **3.4. Energy valorization**

411 The results obtained in the tests described in Sections 3.2 and 3.3 were employed for a preliminary 412 technical and economic evaluation of a full-scale AD reactor for the energy valorization of the 413 waste products generated in Plant 1 and Plant 2. The methane specific production for the VMW was assumed equal to 0.223 Nm³/kgVS added, the value returned by the semi-continuous test. For the 414 PSW the methane specific production was assumed equal to 0.578 Nm³/kgVS added, that is 80% of 415 416 the value from the fed-batch test. According to the remarks set out in Section 3.3, this value could be considered acceptable for the design of a full scale plant. The two feeding substrates are 417 generated at a constant rate throughout the year. Their characteristics are also nearly constant. 418 419 The full scale plant is based on a single-stage reactor that operates in mesophilic condition with an HRT of 30 days. HRT was set equal to the value employed in the semi-continuous test. Because the 420 421 dry substance fed to the digester is supposed not to exceed 10% (the highest value allowed in an AD 422 wet process), the PSW has to be diluted with water. From a balance that takes into account the 423 amount and the TS content of the PSW and the TS content allowed in the substrate fed to the digester, the yearly amount of water required for dilution was equal to approxiametey 2,200 m^3 , that 424 425 is 6 L/kg waste. The produced biogas feeds a combined heat and power (CPH) unit based on an 426 internal combustion engine integrated with a heat recovery section. The assumptions concerning the 427 electrical and thermal efficiencies of the CHP unit were based on typical values for commercial 428 units. The main data for the design of the full scale digester are listed in Table 4.

429 The main outcomes that result from the energy balance applied to the digester are listed in Table 5. 430 The overall volume of the digester was calculated taking into account the daily amount of fed 431 substrate, the HRT and the digester filling coefficient. The density of the fed substrate (10% TS) was fixed equal to 1,000 kg/m³, then the daily amount of the fed substrate resulted to be equal to 432 approximately 8 m^3 and the overall volume of the digester of 300 m^3 . 433 434 From the lower heating value (LHV) of the methane generated from the AD process, a first estimate 435 of the daily methane production suggested a CHP unit with a power output of 60 kW. The gross 436 electrical and thermal yields were calculated starting from biogas production and electrical and 437 thermal efficiencies of the CHP unit. The net electrical energy production was obtained by 438 subtracting the internal electrical energy consumption from the gross electrical energy production. 439 The internal electrical energy consumption was assumed to be equal to 10% of the gross production. The available thermal energy was obtained by subtracting the thermal energy requirements of the 440 441 AD process and the heat losses with the outside from the gross thermal energy production. The 442 thermal energy requirements of the AD process were calculated as the amount of heat necessary to raise the temperature of the fed substrate (yearly average, 15°C) to the temperature of the AD 443 process (35°C). The ratio between the thermal energy requirement of the AD process and the gross 444 445 thermal energy production resulted to be of about 10%. Taking into account the geometry of the full-scale digester, the materials employed for construction, the temperature of the fed mixture 446 447 (15°C), soil (15°C) and outside (exterior environment, as monthly averages as reported in the UNI 448 10349 rule), it may be assumed that the amount of thermal energy to balance the thermal losses is of 449 the same order of magnitude of the thermal energy requirement of the AD process (i.e. $7 \cdot 10^4$ kWh/y). 450

The annual net electrical energy production of the digester, equal to 470 MWh, was enough to cover from 30% to 50% of the electrical consumption of a food industry plant with a size similar to that of the plants that produced the waste products considered in this study. On the other hand, the

454	available thermal energy (500 MWh/y) may cover from 10% to near 100% of the plant
455	requirements depending on the energy demand of the processes performed.
456	On the full-scale digester designed as described above, a preliminary economic analysis was carried
457	out. Two different scenarios that refer to the use of the electrical energy generated only by the AD
458	process were considered. In the first scenario the electricity was employed for the internal energy
459	requirements of the food industry plant and no public subsidies were obtained. In the second
460	scenario the electricity was dispatched to the national grid, in agreement with the current Italian
461	legislation (DM 6/07/2012) on energy production from renewable sources. In fact, the most recent
462	Italian policy supports the use of biological by-products deriving from agricultural and food
463	industries as substrate for AD power plants, especially of small sizes (from 1 to 300 kWe). The fee-
464	in tariff of 236 €/MWh is the maximum allowed for the electricity generated in AD power plants.
465	With reference to both scenarios, the total annual costs (C_T) born by the plant for the production of
466	electricity from the AD process was calculated as in (2)
467	$C_{\rm T} = {\rm TFC} \cdot {\rm CCR} + {\rm C}_{\rm O\&M}$ (2)
468	where:

- TFC, total fixed costs, are the costs for the implementation of the digester power plant;
- CCR, capital charge rate, calculated as in (3), where (i) is the annual interest rate and (n) the
 operating lifetime;
- C_{0&M}, annual operating and maintenance costs.
- 473 $CCR = \frac{i}{1 (1 + i)^{-n}}$ (3)
- For the economic assessment carried out in this work, the annual interest rate was fixed to 7% and the operating lifetime to 20 years (as stated by DM 6/07/2012). The TFC was calculated as the sum of the costs of digester, CPH unit and installation. Plant unit capital cost investment were fixed to $6,000 \notin$ /kWe. Since the electrical power of the plant was of 60 kWe and the costs for installation

were fixed to $100,000 \notin$, the resulting TFC was approximately of $450,000 \notin$ and, consequently, the annual depreciation $42,500 \notin$ (see Table 6).

480 The annual operating and maintenance costs were assumed to be equal to 4% of the TFC, that is 18,000 €. The total labor costs were fixed to $15,000 \in$, considering 0.5 workers employed on the 481 plant. Costs for utilities, that is raw materials supply, treating and restoring water and waste 482 483 disposal, were assumed to be equal to 4% of the value of the electrical energy (see EER, Table 6), if 484 dispatched to the national grid as required by Scenario 2. The income from the sale of electrical 485 energy, as in Scenario 2, was calculated multiplying the annual net electrical energy produced (469 486 MWh) by the fee-in tariff of 236 €/MWh and resulted of approximately 110,700 €/y. The total 487 amount of the annual operating costs was then equal to approximately 37,500 €. With reference to the first scenario, the equivalent Cost of Energy (CE), as in (4), was calculated 488 489 through the ratio between the total annual costs (C_T) and the annual net electrical energy 490 production:

 $491 \quad CE = \frac{C_T}{E_N} \quad (4)$

The economic analysis applied to the first scenario returned an equivalent electrical energy
production cost for the full-scale AD reactor of 172 €/MWh, which is lower than the purchase cost
of electrical energy for industrial consumers (200 €/MWh, Scano et al., 2014).

The second scenario takes into account the option of delivering the electrical energy produced by the AD reactor to the national grid, according to the current Italian legislation (DM 6/07/2012). In this case, the economic analysis was carried out with the same assumptions (interest rate, operating lifetime, TFC, operating and maintenance costs) employed for the assessment of Scenario 1. The equivalent cost of energy was calculated as in (5),

500
$$CE = \frac{c_T}{E_N} + C_{IC} - T_{FI}$$
 (5)

where the ratio C_T/E_N is the equivalent cost of electrical energy produced in the AD power plant (see Scenario 1), C_{IC} is the purchase cost of electrical energy for industrial consumers (200 \notin /MWh)

- and TFI is the fee-in tariff (236 €/MWh). The equivalent cost of energy for Scenario 2 was then of
 134 €/MWh.
- 505 The economic analysis for Scenario 2 was concluded with the evaluation of the Net Present Value
- 506 (NPV, after 20 years) that resulted of 302,000 €, the Pay-Back Time of 8.5 years and the ratio
- 507 between the present TFC and the NPV of 0.67. This means that the expenses born for the
- 508 construction and installation of the AD power plant were fully recovered after 8.5 years. Moreover,
- the revenues from the dispatch of electricity to the national grid made possible to reconstitute
- approximately 70% of the sum of money at the end of the useful lifetime for the implementation of
- a new plant (digester, CHP unit and installation).
- 512

513 **5. CONCLUSIONS**

- 514 The pilot scale test returned a methane yield for the VMW of approximately 80% of that obtained in
- the smaller scale. This outcome was in line with those of other studies performed in different scales
- and modes and indicates the success of the pilot scale test. The net electricity produced from the
- 517 digester accounted for 30% to 50% of the food industry plants' consumption and had a cost of 172
- 518 €/MWh. The available thermal energy could cover from 10% to nearly 100% of the plant
- requirements, depending on the heat demand of the processes performed.
- 520

521 AKNOWLEDGEMENTS

- The financial support of the European Community and Regione Piemonte through POR-FESR
 (2007-2013) project SAFEFOOD CONTROL is gratefully acknowledged.
- The Authors wish also to thank two anonymous reviewers for their useful suggestions to improvethe quality of the manuscript.

526 **REFERENCES**

- Angelidaki I., Alves M., Bolzonella D., Borzacconi L., Campos J.L., Guwy A.J., Kalyuzhnyi S., Jenicek P., van Lier J.B., 2009. Defining the biomethane potential (BMP) of solid organic wastes and energy crops: a proposed protocol for batch assays. Water Sci. Technol., 59, 927 – 934.
- APHA, AWWA, WEF, 2005. Standard methods for the examination of water and wastewater,
 21st ed. American Public Health Association, American Water Works Association and Water
 Environment Federation, Washington DC, ISBN: 978-0875530475.
- Batstone D.J., Keller J., Angelidaki I., Kalyuzhnyi S.V., Pavlostathis S.G., Rozzi A., Sanders
 W.T.M., Siegrist H., Vavilin V.A., 2002. The IWA Anaerobic Digestion Model No 1 (ADM1).
 Water Sci. Technol. 45, 65-73.
- 4. Cavaleiro A.J., Ferreira T., Pereira F., Tommaso G., Alves M.M., 2013. Biochemical methane potential of raw and pre-treated meat-processing wastes. Bioresource Technol., 129, 519-525.
- 5. Dareioti M.A., Kornaros M., 2014. Effect of hydraulic retention time (HRT) on the anaerobic
 co-digestion of agro-industrial wastes in a two-stage CSTR system. Bioresource Technol., 167, 407-415.
- 542 6. Fiore S., Roati C., Ruffino B., Marchese F., Zanetti M.C., 2013. Laboratory and pilot tests for
 543 the anaerobic digestion of wine waste. Proceedings of Sardinia 2013, S. Margherita di Pula,
 544 Cagliari, Italy, 30/09/2013-4/10/2013, 12 pp.
- 545 7. Galí A., Benabdallah T., Astals S., Mata-Alvarez J., 2009. Modified version of ADM1 model
 546 for agro-waste application. Bioresource Technol., 100, 2783-2790.
- Sarcia-Peña E.I., Parameswaran P., Kang D.W., Canul-Chan M., Krajmalnik-Brown R., 2011.
 Anaerobic digestion and co-digestion processes of vegetable and fruit residues: process and microbial ecology. Bioresource Technol., 102, 9447-9455.
- 9. Giuliano A., Bolzonella D., Pavan P., Cavinato C., Cecchi F., 2013. Co-digestion of livestock
 effluents, energy crops and agro-waste: Feeding and process optimization in mesophilic and
 thermophilic conditions. Bioresource Technol., 128, 612-618.
- 10. Hosseini Koupaie E., Barrantes Leiva M., Eskicioglu C., Dutil C., 2014. Mesophilic batch
 anaerobic co-digestion of fruit-juice industrial waste and municipal waste sludge: Process and
 cost-benefit analysis. Bioresource Technol., 152, 66-73.
- Jard G., Jackowiak D., Carrère H., Delgenes J.P., Torrijos M., Steyer J.P., Dumas C., 2012.
 Batch and semi-continuous anaerobic digestion of *Palmaria palmata*: Comparison with *Saccharina latissima* and inhibition studies. Chem. Eng. J. 209, 513-519.
- Jiang Y., Heaven S., Banks C.J., 2012. Strategies for stable anaerobic digestion of vegetable
 waste. Renew Energy, 44, 206-214.
- 13. Kafle G.K., Kim S.H., 2013. Anaerobic treatment of apple waste with swine manure for biogas
 production: Batch and continuous operation. Appl. Energ. 103, 61-72.
- 14. Khalid A., Arshad M., Anjum M., Mahmood T., Dawson L., 2011. The anaerobic digestion of
 solid organic waste (review). Waste Manage, 31, 1737-1744.
- 565
 15. Kolbl S., Paloczi A., Panjan J., Stres B., 2014. Addressing case specific biogas plants tasks:
 566
 567
 568
 568
 569
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 560
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 560
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- 16. Lisboa M.S., Lansing S., 2013. Characterizing food waste substrates for co-digestion through
 biochemical methane potential (BMP) experiments. Waste Manage., 33, 2664-2669.
- 17. Novarino D., Zanetti M.C., 2012. Anaerobic digestion of extruded OFMSW. Bioresource
 Technol., 104, 44-50.
- 18. Palmowski, L.M., Muller, J.A., 2000. Influence of the size reduction of organic waste on their
 anaerobic digestion. Water Sci. Technol. 41, 155–162.

19. Palmowski, L.M., Muller, J.A., 2003. Anaerobic degradation of organic materials significance 575 576 of the substrate surface area. Water Sci. Technol. 47, 231–238. 20. Petruccioli M., Raviv M., Di Silvestro R., Dinelli G., 2011. Agriculture and Agro-Industrial 577 578 Wastes, Byproducts and Wastewaters: Origin, Characteristics and Potential in Bio-Based-579 Compounds Production. Reference Module in Earth Systems and Environmental Sciences, from Comprehensive Biotechnology (Second Edition), vol. 6, pp. 531-545. 580 21. Raposo, F., Fernandez-Cegri, V., De la Rubia, M.A., Borja, R., Beline, F., Cavinato, C., 581 582 Demirer, G., Fernandez, B., Fernandez-Polanco, M., Frigon, J.C., Ganesh, R., Kaparaju, P., Koubova, J., Mendez, R., Menin, G., Peene, A., Scherer, P., Torrijos, M., Uellendahl, H., 583 584 Wierinck, I., de Wildep, V., 2011. Biochemical methane potential (BMP) of solid organic 585 substrates: evaluation of anaerobic biodegradability using data from an international interlaboratory study. J. Chem. Technol. Biotechnol. 86, 1088–1098. 586 587 22. Roati C., Fiore S., Ruffino B., Marchese F., Novarino D., Zanetti M.C., 2012. Preliminary Evaluation of the Potential Biogas Production of Food-Processing Industrial Wastes. American 588 589 Journal of Environmental Sciences, 8, 291-296. 23. Ruffino B., Campo G., Genon G., Lorenzi E., Novarino D., Scibilia G., Zanetti M.C., 2015 590 Improvement of anaerobic digestion of sewage sludge in a wastewater treatment plant by means 591 of mechanical and thermal pre-treatments: Performance, energy and economical assessment. 592 593 Bioresource Technol., 175, 298-308. 24. Scano E.A., Asquer C., Pistis A., Ortu L., Demontis V., Cocco D., 2014. Biogas from anaerobic 594 digestion of fruit and vegetable wastes: Experimental results on pilot-scale and preliminary 595 performance evaluation of a full-scale power plant. Energ. Convers. Manage., 77, 22-30. 596 597 25. Shi X.S., Yuan X.Z., Wang Y.P., Zeng S.J., Qiu Y.L., Guo R.B., Wang L.S., 2014. Modeling of the methane production and pH value during the anaerobic co-digestion of dairy manure and 598 599 spent mushroom substrate. Chem. Eng. J., 244, 258-263. 600 26. Tchobanoglous, G., Theisen H., Vigil S.A., 1993. Integrated solid waste management. McGraw-Hill International Editions, pp. 681-682. 601 27. Traversi D., Bonetta S., Degan R., Villa S., Porfido A., Bellero M., Carraro E., Gilli G., 2013. 602 Environmental Advances Due to the Integration of Food Industries and Anaerobic Digestion for 603 604 Biogas Production: Perspectives of the Italian Milk and Dairy Product Sector. Bioenerg. Res., 6, 605 851-863. 28. Triolo J.M., Pedersen L., Qu H., Sommer S.G., 2012. Biochemical methane potential and 606 607 anaerobic biodegradability of non-herbaceous and herbaceous phytomass in biogas production. Bioresource Technol., 125, 226-232. 608 29. VDI Standard, 2006. VDI 4630 Fermentation of organic materials. Characterization of the 609 610 substrate, sampling, collection of material data, fermentation tests, pp. 92, available on line at 611 <http://www.vdi.eu/guidelines, last accessed 1/27/2015>. 30. Wang L., Shen F., Yuan H., Zou D., Liu Y., Zhu B., Li X., 2014. Anaerobic co-digestion of 612 613 kitchen waste and fruit/vegetable waste: Lab-scale and pilot-scale studies. Waste Manage., 34, 2627-2633. 614 615 31. Xie, S., Lawlor, P.G., Frost, J.P., Hu, Z., Zhan, X., 2011. Effect of pig manure to grass silage ratio on methane production in batch anaerobic co-digestion of concentrated pig manure and 616 617 grass silage. Bioresour. Technol. 102, 5728-5733. 32. Yang Y.Q, Shen D.S., Li N., Xu D., Long Y.Y., Lu X.Y., 2013. Co-digestion of kitchen waste 618 619 and fruit-vegetable waste by two-phase anaerobic digestion. Environ. Sci. Pollut. Res. 20, 2162-620 2171. 621 33. Zhang, Y., Banks, C.J., 2013. Impact of different particle size distributions on anaerobic 622 digestion of the organic fraction of municipal solid waste. Waste Manage. 33, 297–307. 623 34. Zhang C., Su H., Tan T., 2013. Batch and semi-continuous anaerobic digestion of food waste in 624 a dual solid-liquid system. Bioresource Technol., 145, 10-16.

- 625 35. Zupanĉiĉ G.D., Jemec A., 2010. Anaerobic digestion of tannery waste: semi-continuous and
- anaerobic sequencing batch reactor processes. Bioresource Technol., 101, 26-33.

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629 Table Legends

- Table 1. Physical and chemical characterization of the substrates employed in the tests
- 631 Table 2. Results of the carbon balance (substrate biogas)
- Table 3. Comparison of the main results obtained from the two test modes (FB, fed-batch; SC,
- 633 semi-continuous)
- Table 4. Main design parameters for the full scale digester
- Table 5. Main performance outcomes of the full scale digester
- Table 6. Outcomes of the economic analysis applied to the two scenarios
- 637

638 Figure Legends

- Figure 1. Trend of the daily biogas production (1a) and biogas composition b.v. for the VMW (1b)
- Figure 2. Trend of the daily biogas production (2a) and biogas composition b.v. for the PSW (2b)
- Figure 3. Theoretical and experimental trend of the daily biogas production. Theoretical daily
- biogas production due to each feedstock. (3a) VMW, (3b) PSW
- Figure 4. Semi-continuous test: trend of the daily biogas production and biogas and methane
- specific production (4a). Biogas composition b.v. for the vegetable mix waste (4b)
- Figure 5. Semi continuous test: trend of pH, FOS-TAC, OLR (5a). Trend of TS and VS/TS (5b)
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				acron or an		projea m.		
	тс	VS/TS		С	ц	N	8	Molecular
Sample	(%)	ratio	pН	(%)	$(\%)_{1}$	(%)	(%)	formula
	(<i>ne</i>)wei	(%) _{dry}		(<i>ic)</i> diy	(<i>ic)</i> diy	(70)uly	(/0)uly	iorinaia (dry
Peeling	6.29	85.8	4.5	-				
Screen	10.4	85.0	5.0	45.2	6.32	2.93	0.316	$C_{18}H_{30}O_{13.5}$
Filtrate	9.08	89.6	6.0					
Pesto	69.3	98.0	acidic		N.A.			$C_{48.5}H_{86}O_{17}$

Table 2. Results of the carbon balance (substrate – biogas) 662

	Substrate	C amount (g) feedstock	C (moles) feedstock	C (moles) CH ₄ biogas	C (moles) CO ₂ biogas	C conversion (%)
	Vegetable mix waste	35.4	2.95	1.02	0.63	56
	Waste from pesto sauce production	50.0	4.16	2.64	1.18	92
663						0
664 665						
					0	•
					6	
		0				
	0					

Table 3. Comparison of the main results obtained from the two test modes (FB, fed-batch; SC, 666

semi-continuous) 667

		Buswell and	Ead batab	Semi-	SC vs.
		Neave model	reu-batch	continuous	FB ratio
	Vegetable mix waste				
	Biogas specific production (Nm ³ /kg VS)	0.848	0.554 ± 0.038	0.503	0.91
	Methane specific production (Nm ³ /kg VS)	0.424	0.294 ± 0.029	0.223	0.76
	Average methane content (%)	50	53.0 ± 1.8	44.3	0.84
	Waste from pesto sauce production				
	Biogas specific production (Nm ³ /kg VS)	1.14	1.08 ± 0.05		-
	Methane specific production (Nm ³ /kg VS)	0.713	0.722 ± 0.033		-
	Average methane content (%)	63	66.9 ± 0.2		-
668 669					

670 Table 4. Main design parameters for the full scale digester

		u.m.	Waste from Plant 1 (VMW)	Waste from Plant 2 (PSW)
	Waste generation	t/v	400	360
	Waste TS content	<u>% b.w.</u>	10.0	69.3
	Waste VS/TS ratio	-	0.850	0.980
	Methane specific production	Nm ³ /kg VS	0.223	0.578
	Fed substrate TS content	% h w	10.0	10.0
	HRT	davs	30	30
	Digester filling coefficient	-	0.8	0.8
	CHP electrical efficiency	-	0.35	0.35
	CHP thermal efficiency	_	0.42	0.42
	Process temperature	°C	35	35
	Temperature of the fed substrate	<u>°C</u>	15	15
671	Temperature of the fed substrate	C	15	15
			A	

Table 5. Main performance outcomes of the full scale digester 673

Working digester volumeOverall digester volumeAverage biogas productionMethane productionBiogas primary energy productionGross electrical energy productionGross thermal energy productionElectrical energy consumptionThermal energy productionNet electrical energy productionAvailable thermal energyCHP power output	m ³ m ³ /y Nm ³ /y MWh/y MWh/y MWh/y MWh/y MWh/y MWh/y MWh/y	$ \begin{array}{r} 238 \\ 297.5 \\ \hline 2.28 \cdot 10^5 \\ \hline 1.49 \cdot 10^5 \\ \hline 1490 \\ 521 \\ 625 \\ \hline 52.1 \\ 135 \\ \end{array} $
Overall digester volumeAverage biogas productionMethane productionBiogas primary energy productionGross electrical energy productionGross thermal energy productionElectrical energy consumptionThermal energy consumptionNet electrical energy productionAvailable thermal energyCHP power output	m ³ Nm ³ /y Nm ³ /y MWh/y MWh/y MWh/y MWh/y MWh/y MWh/y	$ \begin{array}{r} 297.5 \\ 2.28 \cdot 10^5 \\ 1.49 \cdot 10^5 \\ 1490 \\ 521 \\ 625 \\ 52.1 \\ 135 \\ \end{array} $
Average biogas productionMethane productionBiogas primary energy productionGross electrical energy productionGross thermal energy productionElectrical energy consumptionThermal energy consumptionNet electrical energy productionAvailable thermal energyCHP power output	Nm³/yNm³/yMWh/yMWh/yMWh/yMWh/yMWh/yMWh/yMWh/yMWh/y	$ \begin{array}{r} 2.28 \cdot 10^{5} \\ 1.49 \cdot 10^{5} \\ \hline 1490 \\ 521 \\ 625 \\ 52.1 \\ 135 \\ \end{array} $
Methane productionBiogas primary energy productionGross electrical energy productionGross thermal energy productionElectrical energy consumptionThermal energy consumptionNet electrical energy productionAvailable thermal energyCHP power output	Nm³/yMWh/yMWh/yMWh/yMWh/yMWh/yMWh/yMWh/y	$ \begin{array}{r} 1.49 \cdot 10^{5} \\ 1490 \\ 521 \\ 625 \\ 52.1 \\ 135 \\ \end{array} $
Biogas primary energy productionGross electrical energy productionGross thermal energy productionElectrical energy consumptionThermal energy consumptionNet electrical energy productionAvailable thermal energyCHP power output	MWh/y MWh/y MWh/y MWh/y MWh/y MWh/y	1490 521 625 52.1 135
Gross electrical energy production Gross thermal energy production Electrical energy consumption Thermal energy consumption Net electrical energy production Available thermal energy	MWh/y MWh/y MWh/y MWh/y MWh/y	521 625 52.1 135
Gross thermal energy production Electrical energy consumption Thermal energy consumption Net electrical energy production Available thermal energy	MWh/y MWh/y MWh/y MWh/y	625 52.1 135
Electrical energy consumption Thermal energy consumption Net electrical energy production Available thermal energy	MWh/y MWh/y MWh/y	52.1 135
Thermal energy consumption Net electrical energy production Available thermal energy CHP power output	MWh/y MWh/y	135
Net electrical energy production Available thermal energy CHP power output	MWh/y	
Available thermal energy		469
CHD nower output	MWh/y	490
	kW	60

Table 6. Outcomes of the economic analysis applied to the two scenarios 676

	1	Scenario 2	
Electrical energy emplo	oyed in the plant	Electrical energy dispatched to the	National grid
Total Fixed Costs, TFC	450.000€	Total Fixed Costs, TFC	450.000 €
Annual Installment, AI	42,500 €		
Annual Operating and Mainte	nance	Electrical anarou fac in tariff	226 60,000
costs, $C_{O\&M}$		Electrical energy lee-initarini	250 €/IVI W I
Labor	15,000€	Net electrical energy production	469 MWh/y
• Maintenance (4% TFC)	18,000€	Electrical Energy Revenues, EER	110,700 €/y
• Utilities (4% EER, scena	rio 2) 5,000 €		
Total annual CO&M	37,500€	Net cash flow (EER $- C_{O\&M}$)	73,200 €/y
Total annual costs ($Ct = AI + 0$	C _{0&M}) 80,500 €	Net Present Value (after 20 years), NPV	302,000 €
Net electrical energy production	on 469 MWh/y	Pay-back time	8.5 y
Cost of Energy, CE	172 €/MWh	NPV/TFC	0.67
	4		





Figure 1. Trend of the daily biogas production (1a) and biogas composition b.v. from VMW (1b)

681





Figure 2. Trend of the daily biogas production (2a) and biogas composition b.v. from PSW (2b)

684





685

Figure 3. Theoretical and experimental trend of the daily biogas production. Theoretical daily
biogas production due to each feedstock. (3a) VMW, (3b) PSW.





Figure 4. Semi-continuous test: trend of the daily biogas production and biogas and methane

691 specific production (4a). Biogas composition b.v. for the vegetable mix waste (4b)

692





694 Figure 5. Semi continuous test: trend of pH, FOS-TAC, OLR (5a). Trend of TS and VS/TS (5b)

695

696

697 Highlights

- 698 CH₄ production capacity of vegetable waste was assessed on different scales and modes
- 699 CH₄ yield from the pilot scale test was about 80% of that from the smaller scale test
- Produced net electricity was from 30% to 50% of the food industry plant consumption 700
- , giren 701 Available thermal energy can cover from 10% to near 100% of the plant requirement