

Long-term retrospective analysis of the societal metabolism of cobalt in the European Union

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Long-term Retrospective Analysis of the Societal Metabolism of Cobalt in the European Union

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| Abstract: | <p>A proper understanding of the historical societal metabolism of cobalt (Co) in the European Union (EU) is lacking, resulting in a serious weakness in identifying potential sources of secondary materials, enhancing circularity, and improving forecasting assessments. A retrospective dynamic material flow analysis was performed to assess the stocks and flows of Co in the EU27, considering a number of Co-containing commodities and final metal products. Detailed results are provided for the year 2018, and the evolution of the stocks and flows over time is presented from 1955 to 2018. The results for 2018 indicate that the largest stock in the EU was the landfill stock, with around 166,000 metric tons of Co (35% of the total stock in that year), followed by the in-use stock, with around 157,000 metric tons of Co (33% of the total stock). From 1970 onwards, the in-use stock has doubled its size every 10 or 11 years and the landfill stock every eight years. A sensitivity analysis was performed with eight sets of parameters to determine their impact on the stocks in 2018. Six of them have a marginal impact on the stocks. The in-use stock and the stock of losses are mainly affected by the data uncertainty related to the end-use demand of Co and the lifetimes of the applications (between 4 and 22% in absolute terms for the in-use stock, and between 7 and 27% in absolute terms for the stock of losses).</p> |

1 Long-term Retrospective Analysis of the Societal Metabolism of Cobalt in the 2 European Union

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9 Abstract

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11 is lacking, resulting in a serious weakness in identifying potential sources of secondary materials,
12 enhancing circularity, and improving forecasting assessments. A retrospective dynamic material flow
13 analysis was performed to assess the stocks and flows of Co in the EU27, considering a number of Co-
14 containing commodities and final metal products. Detailed results are provided for the year 2018, and
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18 Co (33% of the total stock). From 1970 onwards, the in-use stock has doubled its size every 10 or 11
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21 stocks. The in-use stock and the stock of losses are mainly affected by the data uncertainty related to
22 the end-use demand of Co and the lifetimes of the applications (between 4 and 22% in absolute terms
23 for the in-use stock, and between 7 and 27% in absolute terms for the stock of losses).

24
25 **Keywords:** *Cobalt, Critical raw materials, Dynamic material flow analysis, Retrospective analysis,*
26 *Recycling*

27 28 List of abbreviations

| | |
|--------|--|
| ΣF | : Cumulative flow |
| BGS | : British Geological Survey |
| Co | : Cobalt |
| CRM | : Critical raw material |
| D | : Downcycled |
| DRC | : Democratic Republic of Congo |
| E | : Extrapolation |
| EoL | : End-of-life |
| EoL-RR | : End-of-life recycling rate |
| EoS | : End-of-service |
| EU | : European Union |
| dMFA | : Dynamic material flow analysis |
| GDP | : Gross domestic product |
| ICT | : Information and communications technology |
| MC | : Model calculation |

| | |
|--------|--|
| MFA | : Material flow analysis |
| MSA | : Material system analysis |
| NCA | : Battery cathode of nickel, cobalt, aluminium |
| NMA | : Battery cathode of nickel, manganese, aluminium |
| NMC | : Battery cathode of nickel, manganese, cobalt |
| OSR | : Old scrap ratio |
| OVC | : Output from the value chain |
| PCLR | : Pre-consumer loss rate |
| PET | : Polyethylene terephthalate |
| RESUCE | : Resource-Efficient Pathways towards Greenhouse-Gas- Neutrality |
| RPER | : Recycling process efficiency rate |
| SD | : Statistical data |
| SI | : Supporting information |
| SSP | : Self-sufficiency potential |
| T | : Tailings |
| TS-RIR | : Total scrap recycling input rate |
| UBA | : Federal Environment Agency of Germany |
| ULJAS | : Finnish customs databases |
| USGS | : US Geological Survey |
| W | : Waste |
| YoY | : Year-on-year |

29

30 1. Introduction

31 For the last decade, cobalt (Co) has been considered a critical raw material (CRM) for the European
32 Union (EU) since the first CRMs list was published in 2011. Its status has been affirmed in the three
33 updates of the list (in 2014, 2017, and 2020). CRMs are characterised by having high economic
34 importance for the EU and a high supply risk (European Commission, 2020a). In the case of Co, the
35 particularly high importance is mostly related to its strategic use in the production of Li-ion batteries,
36 which are key in the transition from fossil fuels to more sustainable sources of energy. The high risk is
37 due to a particularly high concentration of primary supply in the Democratic Republic of Congo (DRC),
38 a country considered politically unstable (World Bank, 2021b). In 2019, the DRC accounted for 71% of
39 the global production of Co, followed by Finland, New Caledonia, Russia, and Canada (USGS, 2020).

40 The EU's list of CRMs was developed as part of the Raw Materials Initiative, launched by the European
41 Commission in 2008. The initiative was based on three main pillars: ensuring a level playing field in
42 terms of access to resources in third countries, fostering a sustainable supply of raw materials from
43 European sources, and boosting resource efficiency and promoting recycling (European Commission,
44 2008). Later, in 2015, the Commission launched the EU action plan for the Circular Economy (European
45 Commission, 2015), which included a number of measures to maintain the value of products, materials,
46 and resources in the economy for as long as possible while minimising the generation of waste. CRMs
47 were one of the five priority areas defined in the action plan. In 2019, a new action plan on circular
48 economy was launched in March 2020, as part of the European Green Deal. This provides an action
49 plan to boost the efficient use of resources by moving to a circular economy, and to restore biodiversity
50 and cut pollution (European Commission, 2020c).

51 In this context, the European Commission has commissioned a high number of studies, aiming to better
52 understand the societal metabolism of CRMs, particularly Co. Many of these studies have been

53 developed with a lifecycle perspective using different tools, such as life cycle assessment and material
54 flow analysis (MFA) (RPA, 2012; BIO by Deloitte, 2015; Alves Dias et al. 2018; Matos et al., 2020).

55 MFA studies can be carried out using static or dynamic models. The former are based on linear
56 correlations and analyse the stocks and flows of a material within the system boundary for a year or at
57 a given point of time in the past. The latter describes the behaviour of a system over a time interval,
58 most applying lifetime distributions to the static flow of a material to analyse and forecast its stocks and
59 flows over time (Elshkaki, 2007). One example of dynamic modelling is the MaTrace model (Nakamura
60 et al., 2014; Godoy León et al., 2020), which tracks the fate of materials over time and across products.
61 It also explicitly considers the losses incurred during the conversion processes.

62 For the EU, one of the most well-established static MFA studies is the MSA (material system analysis)
63 study conducted by BIO by Deloitte (2015) on behalf of the European Commission in 2015. The study
64 consisted of a map of the flows through the economy, including the inputs and movements within it,
65 additions to stock, and end-of-life through either disposal or recovery. For Co, the assessment was
66 done for 2012. A recent MSA study was published by Matos et al. (2020), which included an updated
67 overview of the stocks and flows of Co in the EU for 2016 along with four other materials used in
68 batteries (lithium, manganese, natural graphite, and nickel).

69 Since the publication of the first MSA, several other studies have used its results to estimate the current
70 and future use of Co in the EU (Deetman et al., 2017; Monnet and Abderrahim, 2018; Bobba et al.,
71 2019; Tercero Espinoza et al., 2019, Godoy León et al., 2020), some using a dynamic approach.
72 However, little has been done in terms of retrospective analysis of stocks and flows of Co, which is
73 fundamental to properly understand the societal metabolism of the metal in the region. At the EU level,
74 to the best of the authors' knowledge, only Tercero Espinoza et al. (2019) estimated the historical Co
75 demand in the EU (between 2000 and 2017). Other studies, such as Zeng and Li (2015) and Sun et al.
76 (2019), have used dynamic MFA (dMFA) to trace Co stocks and flows in China and on a global scale,
77 respectively. The former assessed the period 2005–2013 and the latter 1995–2015. Recently, Liu et al.
78 (2021) developed a dMFA of critical metals for lithium-ion batteries in China for the period 2000–2018,
79 in which Co was also assessed. For other metals (e.g. copper), there are long-term retrospective dMFA
80 studies at the EU level (e.g. Ciacci et al. (2017), Soulier et al. (2018)), but that is not the case for Co. In
81 one of our recent studies (Godoy León et al., 2021), long-term statistical data on Co production and
82 trade at the EU level were analysed, acquiring and comparing data from different sources. The study
83 focused on the production and trade flows of Co-containing commodities, with a critical analysis of the
84 available data, but did not assess the full cycle of Co. Furthermore, it did not address other flows (e.g.,
85 flows of Co in final products and end-of-life flows), or how the stocks of Co have built up over time.

86 Therefore, there is a clear knowledge gap regarding the historical long-term societal metabolism of Co
87 in the EU. However, such knowledge is crucial to understand shifts and trends in cobalt flows along its
88 entire cycle, and to identify potential sources of secondary materials originating from earlier stock build-
89 ups. This, in turn, is required to enhance circularity in the economy. Furthermore, it is also essential to
90 improve forecasting assessments, which are key in discussions on managing the metal.

91 The objective of this work is to estimate long-term historical stocks and flows of Co in the EU27 in order
92 to fill in the above-mentioned knowledge gap. A long-term retrospective dMFA was carried out, building
93 on by the results of our aforementioned recent study (Godoy León et al., 2021). The model, data,
94 assumptions, and results are presented in the following.

95 96 2. Methodology

97 98 2.1 System boundaries

The model was applied for the EU from 1955 to 2018. The acquired data from our previous study (Godoy León et al., 2021) covered the period 1938-2018. However, the most complete dataset was for the period 1955-2018, which is why the present assessment was done for that timespan.

The 27 current member states were included in the analysis, including the period before they joined the EU. The analysed system considers the full life cycle of Co in the EU, from mining to recycling (see Figure 1), and takes into account the domestic production and international trade of a number of Co-containing commodities and final products (listed in the SI). These were established according to Matos et al. (2020). The commodities were classified as primary, secondary, semi-processed, and processed material. The final products containing Co were categorised in seven applications: batteries, catalysts, intentionally dissipative uses (modelled as pigments), hard metals, magnets, superalloys, and other uses (e.g. tool steels). Two sub-categories were studied for batteries: portable and mobility batteries. The focus was on Co-bearing rechargeable batteries. In the case of catalysts, three sub-categories were studied: for hydroprocessing, for hydroformylation, and for the production of polyester (PET) precursors. For intentionally dissipative uses, the focus was on pigments. A description of these applications can be found in Godoy León and Dewulf (2020).

2.2 Model, data, and assumptions

The model used to estimate the societal metabolism of Co in the EU27 is based on calculating material stocks and flows according to the law of conservation of mass. The model links the different phases of the life cycle of the metal through annually calculated material flows, considering a closed mass balance within one year and over time.

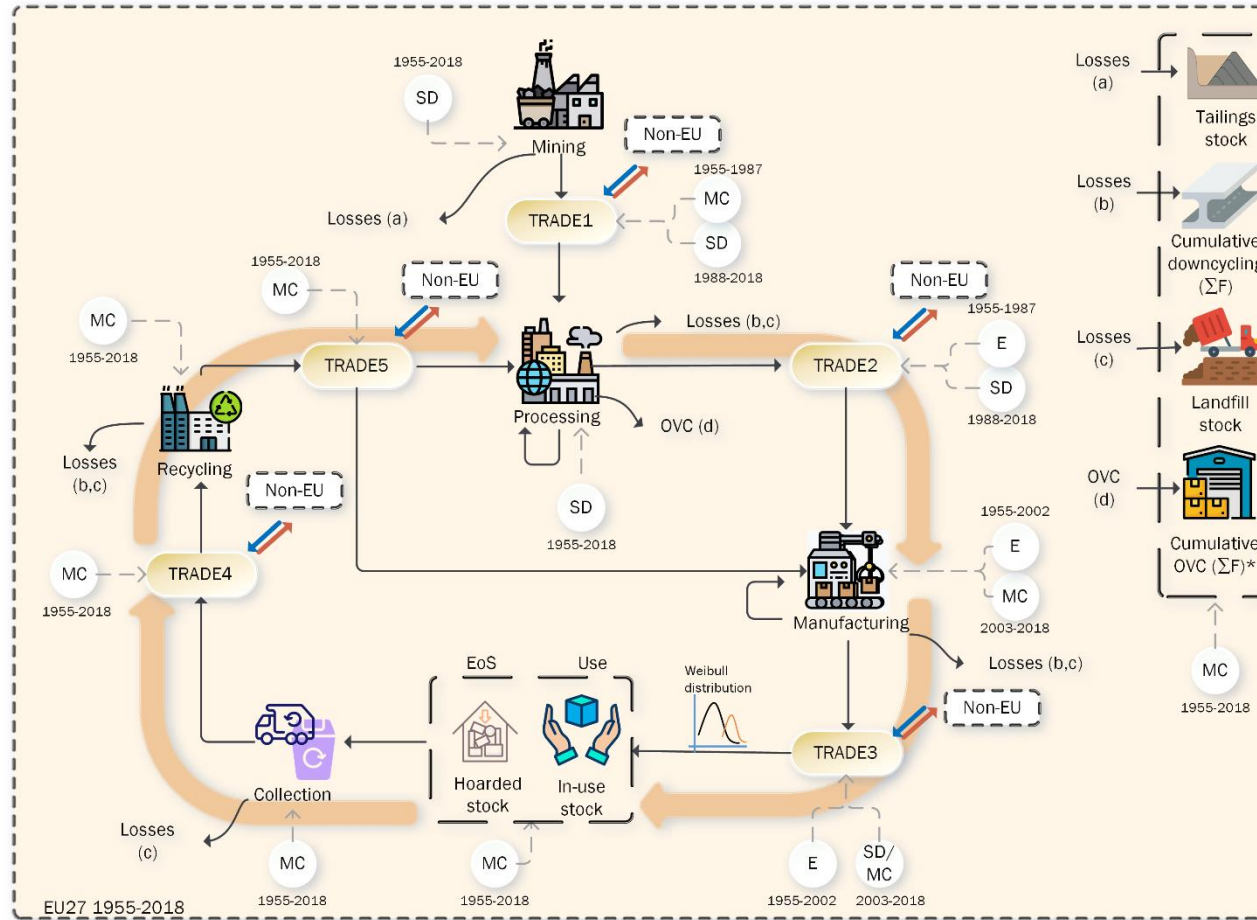
The input data of the model, specifically related to production and trade, included the sum of separately collected data for each of the 27 EU member states. Based on the available data, the model was developed for three periods: 1955–1987, 1988–2002, and 2003–2018. As can be seen in Figure 1 and Table 1, different types of input data for production and trade were required for each period. These were classified as statistical data (SD), model calculations (MC) (i.e. input data calculated through the model, such as when closing mass balances), and extrapolations (E) (i.e. extrapolated values from historical trends).

Table 1. Adaptations of the model according to the available data for the periods 1955–1987, 1988–2002, and 2003–2018 for each life cycle phase. E: Extrapolation, SD: Statistical data, MC: Model calculation, EoS: End-of-service. The trade refers to the numerated trade of Figure 1.

| Period | Mining | Processing | Manufacturing | Use | EoS | Collection and Recycling |
|-----------|----------------|------------------------------|------------------------------|----------------------------|-----------|--------------------------------|
| 1955-1987 | Production: SD | Trade1: MC Production: SD | Trade2: E Production: E | Trade3: E Stock: MC | Stock: MC | Production: MC Trade4–5: MC |
| 1988-2002 | Production: SD | Trade1: SD Production: SD | Trade2: SD Production: E | Trade3: E Stock: MC | Stock: MC | Production: MC Trade4–5: MC |
| 2003-2018 | Production: SD | Trade1: SD Production: SD | Trade2: SD Production: MC | Trade3: SD+MC Stock: MC | Stock: MC | Production: MC Trade4–5: MC |

For the first period (1955–1987) a bottom-up approach (i.e. based on data at the product level, which are more disaggregated; for example ‘Co ores and concentrates’) was used for the mining phase, as statistical data on production were available. For the processing phase, statistical data were available for production but not for trade, which had to be estimated through the model. For the manufacturing and use phases, the trade was obtained from extrapolations. The flows and stocks of the following life cycle phases were obtained using the applied model through a series of parameters, such as lifetimes, hoarding rate and time, and collection rate of the different applications.

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139 *Figure 1. Overview of the methodology applied to model the Co stocks and flows in the EU between 1955 and 2018. The short-dashed line rectangle represents the system*
 140 *boundaries; the long-dashed line rectangle encompasses stocks or cumulative flows. Black arrows represent mass flows. Grey dashed arrows represent input of data for the*
 141 *trade and the production/stock of the life cycle phases (E: Extrapolation, SD: Statistical data, MC: Model calculation), which is given per period. The losses ((a) Mining waste, (b)*
 142 *Downcycled material, (c) Processing, manufacturing, collection, and recycling waste) and the output from the value chain (OVC (d)) go to their respective stocks. The numerated*
 143 *trade relates to Table 1. Input data related to the parameters of the model, such as processes efficiencies, collection rates, and hoarding rate and time, are not explicit in the*
 144 *diagram, but they are part of the model calculations (MC). Stocks and losses are also calculated from the model. EoS: End-of-service. * Cumulative OVC refers to the stockpiling*
 145 *and/or confidential export aggregated over time. ΣF : cumulative flow.*

146 For the second and third periods (1988–2002, 2003–2018), a bottom-up approach was used for the
1 147 mining and processing phases. Data on production and trade were reasonably available, and it was
2 148 possible to estimate Co content due to the limited number of commodities in the EU. Regarding the
3 149 manufacturing and use phase, production and trade were extrapolated for the second period. For the
4 150 third period, a top-down approach (i.e. based on data at the EU level for more aggregated categories,
5 151 such as ‘Carbides whether or not chemically defined’) was applied due to the large number of finished
6 152 and semi-finished products and the lack of data related to their material content. For both periods, the
7 153 following steps were calculated according to the model, based on collecting and recycling practices.

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10 154 An exhaustive explanation of the bottom-up and top-down approaches is available in BIO by Deloitte
11 155 (2015).

12
13 156 In addition, different datasets for the model parameters (e.g. lifetimes, collection rates, scrap recoveries)
14 157 were used per period when data were available (see supporting information (SI)).

15
16 158 In the following, the model is explained per life cycle phase, along with its adaptations for each period.
17 159 Figure 1 presents a summary of the used methodology, indicating the statistical data, extrapolation, and
18 160 calculated data inputs.

20 161 2.2.1 Mining phase

21
22 162 For this phase, the produced material was obtained directly from the British Geological Survey (BGS,
23 163 2021) and the US Geological Survey (USGS, 2021). Due to the availability of data and the differences
24 164 between the two sources, the data were used as follows. Data from BGS were used between 1955 and
25 165 1959 (no data were available from USGS), and data from USGS were used for the period 1962–1969
26 166 (no data were available from BGS). For the years 1960 and 1961 none of the sources reports data. The
27 167 value for these years was assumed as the average value of the years 1959 and 1962. For the period
28 168 1970–2017, an average value between the data from BGS and USGS was used. For the year 2018,
29 169 only data from BGS were used due to lack of data from USGS for that year (see SI). For the calculation
30 170 of the waste produced in this phase, an extraction efficiency of 79% (BGS, 2021; BMLRT, 2021) was
31 171 used for the studied period. Based on the annual waste production (loss (a) in Figure 1), the cumulative
32 172 stock of tailings was calculated.

35 173 2.2.2 Processing phase

36
37 174 For the processing phase, data on the produced material (processed Co) were obtained directly from
38 175 the BGS and USGS databases. An average value was used for the years 1955 and 1956. For the period
39 176 1957–1962, data from BGS were used, and for the period 1963–2017, an average value between the
40 177 data from BGS and USGS was used. For 2018, only data from BGS were used.

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43 178 Regarding trade, the import and export of primary, secondary, and semi-processed Co were
44 179 considered. The data were extracted from the Eurostat and UN Comtrade databases (Eurostat, 2021a;
45 180 UN Comtrade, 2021). The data from these two sources are highly different for certain commodities
46 181 (Godoy León et al., 2021), which is why both datasets were analysed separately. In addition, the import
47 182 data in UN Comtrade were obtained with the member states as reporters and the export data with the
48 183 member states as partners (for further explanation, see Godoy León et al., 2021).

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50
51 184 Due to the lack of data on trade for the period 1963–1987, the net trade of all commodities was
52 185 estimated based on a closed mass balance and the calculation of the domestic production of secondary
53 186 Co from end-of-life (EoL) products. For the period 1988–2018, data for trade were obtained from both
54 187 of the above-mentioned databases. In addition, the Finnish customs databases (ULJAS, 2021) was
55 188 consulted for the trade of semi-processed Co.

56
57 189 As mentioned above, the input of domestic secondary material in this phase was estimated for each
58 190 year of the studied period. An imbalance was detected for some years, where the inputs to the
59 191 processing phase were higher than the outputs. The difference in output was called ‘output from the

value chain' (OVC). It consists of stockpiling, confidential exports (i.e. trade that is not publicly disclosed in the databases; see Godoy León et al., 2021), or a combination of both. Since the origin of this stream is uncertain, its accumulation over time is not considered a stock, and it is referred to only as a cumulative flow (see Figure 1).

For this phase, the production of processed material and waste was considered as well as the recovery of scrap and the downcycling of material. Downcycling, or non-functional recycling, refers to material that is incorporated into an associated large-magnitude material stream, ending up in low-end products where the original function is not required or as a contaminant (Buchert et al., 2009; Graedel et al., 2011; Zimmermann and Gößling-Reisemann, 2013; Zimmermann 2017).

Modified equations of the MaTrace model (Nakamura et al., 2014; Godoy León et al., 2020) were applied to estimate the flows of recovered scrap, waste, and downcycled material. These equations are based on the continuous internal reuse of scrap and depend on the processing efficiency and recovery rate of scrap. The original equations were extended to include primary material and trade of commodities. The derivation of the equations is presented in SI, along with the set of values used for the different parameters.

2.2.3 Manufacturing phase

Three sets of data were required for the manufacturing phase: the trade of processed material, the manufacturing share of the metal for each application, and the values for a number of parameters related to the calculation of waste and downcycled material.

Data from the Eurostat and UN Comtrade databases were considered for the trade. Due to the lack of data for the period 1955–1987, the net trade and manufacturing demand of these commodities were extrapolated. The extrapolation was performed using the available historical data from 1988 to 2018, considering exponential growth based on the conclusions of Krausmann et al. (2009). According to the authors, the 20th century was characterised by an exponential increase in global materials use (especially after WWII). This growth was driven by both population and economic growth, with a shift from the dominance of renewable biomass toward mineral materials (Krausmann et al., 2009). To validate this approach, the Co demand for manufacturing was compared to the GDP (gross domestic product) of the EU over time. The comparison showed similar results to those obtained considering exponential growth (see SI).

Regarding the manufacturing demand share of the different applications, no data were available for the EU before 2000. In the literature, the demand share was available for China (Zeng and Li, 2015), the USA (USGS, 2021), and the world (Sun et al., 2019). The share for China and the world were excluded, however, due to their high production of batteries, which are barely produced in the EU. The data available from the USA were selected as the most representative due to the similarity of its manufacturing sectors to those of the EU (National Research Council 1983; Sun et al., 2019) and their reasonably similar economic development (World Bank, 2021a). From 2000 onwards, the share was obtained from Roskill (2019), Tercero Espinoza et al. (2019), and Godoy León et al. (2020).

Finally, similarly to the processing phase, the recovered scrap, waste, and downcycled material produced during manufacturing were estimated using modified equations of the MaTrace model (Nakamura et al., 2014; Godoy León et al., 2020). The equations and set of values used for the different parameters are available in the SI.

2.2.4 Use phase.

For the use phase, the stock of in-use applications was modelled. To do so, the trade and lifetimes of the final products were required.

Regarding trade, for intentionally dissipative uses, hard metals, magnets, superalloys, and other uses, data were extracted from the Eurostat database, considering a number of finished or semi-finished

238 products (see SI), with data available from 2003 onwards. For catalysts, a net trade of zero was
1 239 assumed, as in Matos et al. (2020). In the case of portable and mobility batteries, the trade was
2 240 estimated based on data of the ProSUM database (RMIS, 2019; Huisman et al., 2020). The database
3 241 provides what was placed on the market from 2000 onwards.

5 242 Due to data gaps for most of the application before 2003, the final consumption of final products had to
6 243 be estimated for the period 1955–2002. Based on Sun et al. (2019) and Matos et al. (2020), the EU
7 244 was estimated to consume 30% of the global Co demand for final use in 2015. In addition, based on
8 245 the work of Sun et al. (2019) (who provided the Co demand for 1995, 2005, and 2015) and Krausmann
9 246 et al. (2009), the global Co demand for final use was extrapolated considering an exponential growth.
11 247 Subsequently, the Co demand for the EU was estimated based on the extrapolated global demand and
12 248 a share of 30%. To complement this assumption, the GDP of the EU was compared to the global GDP
13 249 (World Bank, 2021a). It was found that between 1966 (EU's GDP is not available before this year) and
14 250 2002, the EU's GDP corresponded to between 21 and 29% of the world's GDP, with an average value
15 251 of 26%.

17 252 Based on the end-use demand and the domestic manufacturing, the net trade of the applications was
18 253 estimated. It is also important to note that the UN Comtrade database was not consulted for the trade
20 254 of final products, as its available categories are more aggregated.

22 255 To estimate the in-use stock, a statistical probability distribution was used to model the lifetime of the
23 256 different applications. These distributions are characterised by a probability density function, which
24 257 describes the probability of events occurring over time. Here, the Weibull distribution was used, as it is
25 258 widely used in life data analysis (Weibull, 2021). The considered lifetimes and parameters related to
26 259 the Weibull distribution are available in the SI.

28 260 Two datasets were used for the lifetimes, one for the period 1955–1987 and another for the period
29 261 1988–2018. The ProSUM database was consulted for information on the waste generated in the EU27
30 262 from portable and mobility batteries, which was available from 2010 to 2018. Based on this information,
31 263 the values of the lifetime of batteries were tuned (comparing the reported waste in the database with
32 264 the values obtained by the model). Intentionally dissipative uses were modelled mainly as pigments,
34 265 which is why this application was not dissipated during the use phase but assumed to be landfilled after
35 266 its use.

37 267 Due to the lack of data from before 1955, it was assumed that the in-use stock before 1955 is negligible.

39 268 2.2.5 End-of-life phase

41 269 This phase is composed of three sub-phases: end-of-service (EoS), collection, and recycling. It is
42 270 assumed that EoS products are hoarded by the user (i.e. dead storage of a product that has reached
43 271 the end of its use, such as an old mobile phone kept in the attic (Wilson et al., 2017)) or are immediately
44 272 disposed for collection. The stock of EoS products, called hoarded stock was calculated based on the
45 273 Weibull distribution, hoarding rate, and hoarding period of each application. The annual addition of
46 274 hoarded products to the stock and the collection and disposal of EoS products were calculated
48 275 according to the equations of Godoy León et al. (2020). This phase also considers the trade of EoS
49 276 scrap.

51 277 Next, the functional recycling and downcycling of Co from EoS products were calculated. For functional
52 278 recycling, two routes were considered. The first involves a chemical process, through which batteries,
53 279 catalysts, and hard metals are recycled. Co metal or a Co compound that can return to the processing
54 280 phase are obtained from this route. The second route includes two different processes, where the
55 281 recycled material goes directly to manufacturing. The first is the Zn process, through which hard metals
56 282 are recycled. W-Co powder is obtained from this process, which is used in the production of new hard
57 283 metals. The second process is remelting, where superalloys are recycled to produce new superalloys.
58 284 The downcycling route is used for catalysts, magnets, superalloys, and other uses, and the recovered
60 285 metallic fraction (including Co) is predominantly used in steel production (Godoy León et al., 2020).

286 The distribution of material in the different recycling routes is specified by an allocation matrix (see SI).
1 287 An allocation matrix was used for the period 1955–2002 and for the period 2003–2018, according to the
2 288 available data.

3
4 289 Given the purpose of this study, the recycling processes were not analysed in detail, as the model only
5 290 requires efficiency of the whole phase (for further details, see Godoy León et al., 2020).

6
7 291 The mass balance of the system was closed by calculating the secondary material produced
8 292 domestically from EoL products:

- 9
10 293 • For the period 1955–1987, the trade of the processing phase was estimated assuming no trade
11 294 of secondary material (i.e. all domestic secondary material is used domestically, and there is
12 295 no import).
- 13
14 296 • For the period 1988–2018, the input of domestic secondary material was obtained based on
15 297 the mass balance between the production and trade of processed material. It was assumed
16 298 that only recycled Co produced through the chemical process is traded (i.e. the recycled Co
17 299 that returns to the manufacturing phase is used domestically). In case of obtaining a negative
18 300 value, it was assumed that part of the input to the processing phase was stockpiled and/or left
19 301 the value chain through confidential exports. In the event that the value was positive, the
20 302 amount was compared with the secondary material produced from EoL products. The
21 303 comparison made it possible to estimate the import or export of secondary Co.

22 304 23 305 2.2.6 Losses

24
25
26 306 The losses, shown in Figure 1, consist of the mining waste (loss (a)); the processing, manufacturing,
27 307 and recycling waste due to the inefficiencies of the processes (loss (c)); the collection waste due to non-
28 308 selective collection (loss (c)); and the downcycled material produced in processing, manufacturing, and
29 309 recycling (loss (b)). The mining waste accumulates in the tailings stock, and the waste produced by
30 310 inefficiencies and non-selective collection accumulates in the landfill stock. The downcycled material is
31 311 dispersed in the technosphere and does not accumulate in a specific stock. For this reason, the
32 312 accumulation of this stream was not called stock but cumulative flow. Nevertheless, it is considered part
33 313 of the lost stock, as the Co ending in this compartment is not currently being qualitatively recycled.

34 314 35 315 2.3 Indicators

36
37
38 315 Six indicators were calculated in order to characterise the Co cycle in the EU: total scrap recycling input
39 316 rate (TS-RIR), old scrap ratio (OSR), end-of-life recycling rate (EoL-RR), recycling process efficiency
40 317 rate (RPER), self-sufficiency potential (SSP), and pre-consumer loss rate (PCLR). The former four
41 318 come from UNEP (2011) and Tercero Espinoza and Soulier (2018), while the latter two come from
42 319 Eurostat (2021b) and Matos et al. (2021). These indicators measure different aspects of the metal cycle,
43 320 such as the amount of reused scrap. The descriptions and equations of the indicators are available in
44 321 the SI.

45 322 46 323 2.4 Sensitivity analysis

47
48
49 323 A sensitivity analysis was performed to assess the impact of the main assumptions of the model. These
50 324 were related to the following parameters or calculations: the manufacturing demand between 1955 and
51 325 1987, the manufacturing demand share between 1955 and 1999, the trade of catalysts between 2003
52 326 and 2018, the end-use demand between 1955 and 2002, the export of functionally recycled material
53 327 between 1955 and 1987, and the extraction efficiency for the complete studied period.

54
55 328 In addition, two other parameters were assessed due to their importance in other dMFA studies (Soulier
56 329 et al., 2018; Godoy León et al., 2020): the lifetime of applications and the collection rates. These are
57 330 key parameters required to model in-use stocks and flows of secondary materials and the available
58 331 information on these parameters bears considerable uncertainty.

Due to the lack of statistical measures of the uncertainty of the input values (i.e. the type of distribution they follow and/or the spread), a simple sensitivity analysis was performed where each parameter was changed individually over a certain range. The results were compared in terms of the in-use stock, OVC cumulative flow (stockpile and/or non-EU exports), hoarded stock, and lost stock (tailings, landfill, and downcycled).

Table 2 lists the assessed parameters or calculations, how they were established or calculated for the base case, and how they were modified in the sensitivity analysis.

*Table 2. List of parameters or calculations assessed in the sensitivity analysis. *Considering the main applications in 2000 (superalloys, hard metals, and intentionally dissipative uses). The difference in percentage is distributed equally among the other applications (when applicable, e.g. mobility batteries were not considered before 2000). **Maximum possible of 100%.*

| Parameter/Calculation | Base case | Range in the sensitivity analysis |
|---|--|-----------------------------------|
| Manufacturing demand (1955–1987) | Extrapolated assuming exponential growth | ±50% of the extrapolated value |
| Manufacturing demand share (1955–1999) | Same share as in the USA | ±50% for main applications* |
| Trade of catalysts (2003–2018) | Net trade of zero | ±50% net trade |
| End-use demand (1955–2002) | 30% of the global Co demand | ±50% of the base case |
| Trade of functionally recycled material (1955–1987) | Net trade of zero | ±50% net trade |
| Lifetime of applications (1955–1987) | See SI | ±50% of the base case |
| Collection rates (1955–1987)** | See SI | ±50% of the base case |
| Extraction efficiency (1955–2018) | 79% | ±25% of the base case |

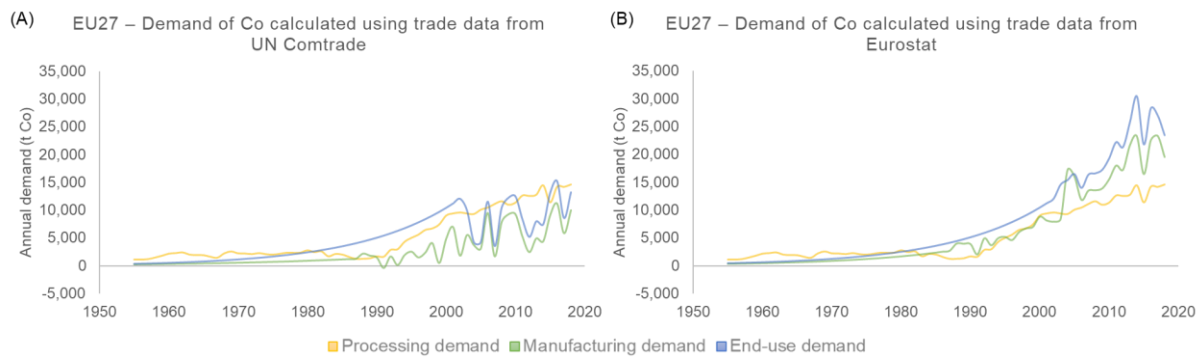
3. Results and discussion

In this section, we present the results obtained from the dMFA of Co in the EU27 from 1955 to 2018. Different aspects are analysed, such as the Co demand over time, the Co cycle for specific years, and how the cumulative stocks are built up over time. We finish this section by presenting the results and a discussion on the sensitivity analysis. The results are presented in metric tons (t).

3.1 Demand of cobalt over time

The evolution of the EU27 demand for Co over time was analysed, comparing the results obtained from the UN Comtrade and Eurostat databases separately. Figure 2 shows the demand for the processing, manufacturing, and end-use phases. Here, demand is understood as apparent consumption (also known as domestic material consumption, DMC), which is defined as ‘production+imports-exports’ (Fastmarkets RISI, 2021; United Nations, 2021).

The demand over time in both cases is similar for the processing phase, as in general the datasets from both sources do not present substantial differences (Godoy León et al., 2021). For the manufacturing and end-use phases, however, the differences are significant. For example, for 2018 the manufacturing and end-use demands obtained with the UN Comtrade dataset are 49 and 44% lower, respectively, than the demands obtained using the dataset from Eurostat. This is mostly explained by the substantial differences in the trade of processed material, as UN Comtrade reports considerably higher values for the export of that type of commodity (Godoy León et al., 2021). This is related to confidential exports from Finland, a topic that will be discussed in the next section.



364
365 *Figure 2. Annual Co demand for processing, manufacturing, and end use in the EU27 between 1955 and 2018,*
366 *according to the data obtained from UN Comtrade and Eurostat.*

367 For more information regarding the analysis and comparison of both datasets, we refer the reader to
368 our previous study (Godoy León et al., 2021).

369 In addition, some inconsistencies are found in the demands obtained with the UN Comtrade dataset.
370 For instance, the manufacturing demand is zero or even negative for some years. Moreover, the
371 processing demand for the period 2012–2016 is higher or almost equal to the end-use demand, which
372 conflicts with the findings of Matos et al. (2020). The demands obtained with the Eurostat dataset do
373 not present these inconsistencies.

374 Furthermore, the manufacturing demand obtained with both datasets is compared to the annual price
375 of Co (see SI). The demand obtained with the dataset from Eurostat is better correlated with the
376 fluctuations of the market. Based on these arguments, the following sections focus on the results
377 obtained with the Eurostat dataset. For the UN Comtrade dataset, the results are shown in the SI. This
378 decision is also supported by the recent study of Watari et al. (2020), in which data from the UN
379 Comtrade database were avoided due to inconsistent trade data between countries.

380 Focusing on the evolution of demand according to the Eurostat data (Figure 2(B)), it has increased
381 overall over time. The demand for processing increased at a high rate between 1990 and 2000, with an
382 average year-on-year (YoY) change of 21%, but for the period 2000–2018 it shows a relative stagnation,
383 with an average YoY change of 4%. For the manufacturing and end-use demands, a more constant
384 increase is observed. The manufacturing demand shows an average YoY change of 16% between
385 1990 and 2000 and of 9% between 2000 and 2018. Meanwhile, the end-use demand shows an average
386 YoY change of 7% between 1990 and 2000 and of 6% between 2000 and 2018. Nevertheless, the
387 studied demands (Figure 2) reveal a lower increase rate between 2000 and 2018 compared with the
388 previous decade. This will be further discussed in the next section.

389 3.2 Cobalt cycle in the EU

390 For the sake of simplicity, the Co cycle in the EU was analysed in detail for three years: 1988, 2003,
391 and 2018, as shown in Figure 3. These years were chosen to obtain a fair overview of the dynamics of
392 Co over time in the studied period, based on the data availability for the periods described in section
393 2.2. The full results are presented in the SI. For each year, the flows entering and leaving each life cycle
394 phase are presented along with the net trade and accumulated stocks. First, the cycle for the latest year
395 will be analysed in detail, followed by a comparison between the cycles of the three years.

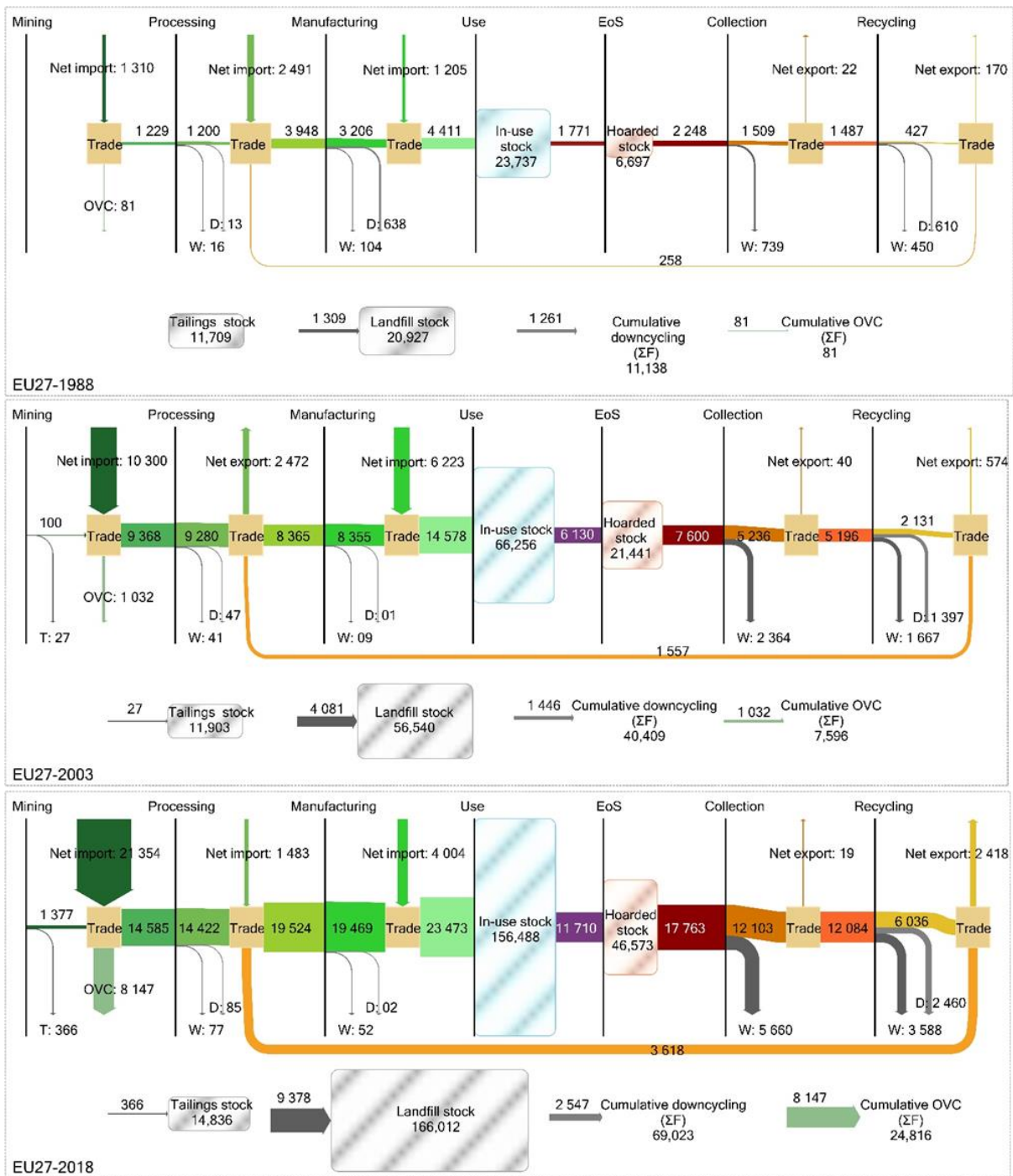


Figure 3. Co cycle in the EU27 in 1988 (A), 2003 (B), and 2018 (C). Values are shown in metric tons (t). Collection includes what is released after hoarding and what is immediately collected at the end-of-service (EoS). D: downcycled, OVC: output from the value chain, T: Tailings, W: waste. Differences of 1 metric ton can be encountered due to approximations. Green is used to represent pre-consumer flows, and red to orange is used to represent post-consumer flows. ΣF: cumulative flow.

3.2.1 Cobalt cycle in 2018

In 2018 (Figure 3(C)), around 1,700 t of primary Co were mined in the EU (specifically in Finland), and about 1,400 t were extracted as Co concentrates. The cycle of this year was characterised by net imports to the processing, manufacturing, and use phases and net exports from the collection and recycling phases. Overall, 24,400 t of Co were imported to the EU, mainly as semi-processed material. Domestically, around 14,400 t of Co were embedded in processed material and 19,500 t in final products. Around 23,500 t of Co were added to the in-use stock, which was about 157,000 t of Co in

410 2018. In addition, 46,600 t of Co were stored in the hoarding stock. About 18,000 t of Co were disposed,
411 of which 68% (12,000 t) were selectively collected for recycling. From the recycling phase (which also
412 includes the pre-treatment phase), 6,000 t were functionally recycled, 2,500 t were downcycled, and
413 3,600 t were disposed as waste. From the functional recycled material, 60% was used domestically,
414 returned to the manufacturing phase, and the rest was exported.

415 Finally, for 2018 there was an output from the value chain of 8,100 t of Co, which plausibly was
416 stockpiled or left the EU through confidential exports. Considering this flow, by 2018 there were almost
417 25,000 accumulated metric tons of Co whose fate is unclear.

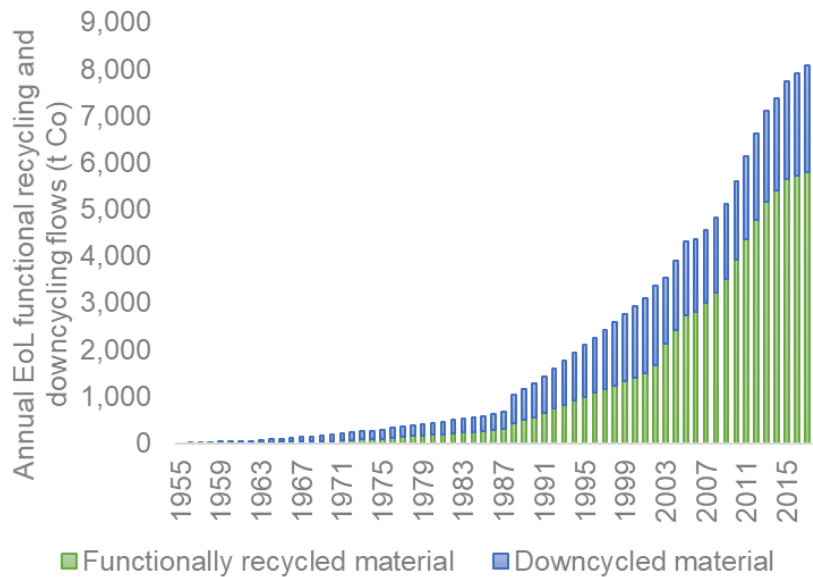
418 3.2.2 Evolution of the cobalt cycle

419 A number of points arise from Figure 3. First, it is clear that in the last 30 years the Co stocks and flows
420 in the EU have increased substantially. For instance, from 1988 to 2003 the inputs to the processing,
421 manufacturing, and use phases increased by 6.6, 1.1, and 2.3 times, respectively. From 2003 to 2018,
422 these inputs also increased considerably, but in general at a lower rate (56, 133, and 61%, respectively).
423 For the phases related to the end of life, a similar behaviour is observed. In particular, the functionally
424 recycled material used domestically increased by 13 times between 1988 and 2018. Second, the EU
425 has become more reliant on imports of materials, especially for the processing phase. The net import
426 to this phase increased by 6.9 times between 1988 and 2003 and by 1.1 times between 2003 and 2018.
427 In addition, the import of semi-processed Co rose steeply at the end of the 1990s and in the middle of
428 the 2010s. At the end of the 1990s, this was probably due to the higher demand of processed Co for its
429 use in rechargeable batteries (Shedd, 1997). A plausible reason for the increase in the middle of the
430 2010s was the construction of gigafactories for battery production in different parts of Europe (Climate
431 Home News, 2020). This occurred in the context of the European Battery Alliance (EBA), aiming to
432 boost the production of electric cars in the continent (EBA, 2021). This indicates that, in the last decade,
433 Co has been stockpiled in the EU. Indeed, Figure 3(C) shows that the OVC stream was 8,100 t of Co
434 in 2018, which was calculated based on the imbalance between the inputs to the processing phase and
435 final production, creating an excess of Co material. This imbalance could be partly explained by
436 stockpiling, to avoid possible uncertainties in the Co market. However, another plausible reason is the
437 existence of confidential exports of semi-processed material with the commodity code CN81052000,
438 which are not reported in Eurostat and are aggregated in the ULJAS database for all countries. From
439 the UN Comtrade database, it is possible to determine that Japan, Russia, and the USA are the main
440 importers of the commodities with the code CN81052000. All these countries have Co refining
441 capabilities, and the ratio of value imported versus mass is lower or equal to the price of refined Co.
442 Therefore, it can be expected that some of the volumes imported from Finland to these three countries
443 are of intermediate Co commodities, that is, 'semi-processed materials'. More details regarding the
444 evolution of the trade are available in the SI and in a previous study (Godoy León et al., 2021).

445 Third, in line with higher inputs and outputs within the system, the Co losses have also increased. A
446 decrease in the losses is only observed between 1988 and 2003 for the manufacturing phase, due to
447 improvements in the efficiency of the processes. The larger losses are for the collection and recycling
448 phases, where the total losses increased by 6.7 and 4.7 times, respectively, between 1988 and 2018.
449 While the model takes into account higher collection rates over time, the collection losses still increase
450 due to higher consumption and subsequent disposal of EoL products. This means that efforts to
451 increase the circularity of Co in the EU should be focused on both differentiated collection and recycling.

452 Finally, regarding the functional recycling and downcycling of Co, both of corresponding flows have
453 increased in time. However, as Figure 4 shows, the share of functionally recycled material has
454 increased over time, while the share of downcycled material has decreased. For instance, in 1960 the
455 share was 10% of recycled Co and 90% of downcycled Co (without considering the produced waste).
456 By 2000, the share was 50/50, and in 2018 the share was 71% of recycled Co and 29% of downcycled
457 Co. This change is due to a shift in the end-use demand, as for some applications the corresponding

458 EoL products are more downcycled than others (e.g. magnets are fully downcycled while superalloys
 459 can follow both routes, mostly depending on the Co price).



460
 461 *Figure 4. Annual flow of functionally recycled and downcycled Co from EoL products between 1955 and 2018.*

462 Returning to Figure 3, it is noteworthy that between 2003 and 2018 the flows of Co in general increased
 463 at a lower rate compared to the previous 15 years. This could indicate saturation of Co in the EU. For
 464 example, Tsiropoulos and colleagues indicate that the electronics market has gradually slowed, and it
 465 is expected to saturate further in the coming years (Tsiropoulos et al., 2018). However, the demand per
 466 capita (see SI) on average has increased at a similar rate; between 1988 and 2003 the average annual
 467 growth rate of the demand per capita was 7.2%, and between 2003 and 2018 it was 7.4%. Furthermore,
 468 when comparing the end-use demand to GDP (known as resource intensity, i.e. how much material is
 469 required to create one unit of monetary value (Krausmann et al., 2018), the resource intensity has
 470 increased over time, from 0.7 g Co per €1,000 of GDP in 1995 to 2.1 g Co per €1,000 of GDP in 2018.
 471 This means that the EU's economy has become more intense in the use of Co over time.

472 To complement the discussion, Figure 5 shows the evolution over time of the manufacturing and use
 473 of applications in the EU27. The values correspond to the average value per decade. Once again, both
 474 the manufacturing and use skyrocket starting in the 1990s. In both cases, the main applications are
 475 intentionally dissipative uses, hard metals, and superalloys. In addition, Figure 5(A) shows that in the
 476 last eight years the manufacturing of magnets and catalysts represents the highest increase (on
 477 average 4.5 and 3.3 times, respectively), followed by mobility batteries and intentionally dissipative uses
 478 (on average 2.5 times). Regarding the use (Figure 5(B)), the highest increase is seen for mobility
 479 batteries and magnets (on average 11 and 6.3 times, respectively), followed by catalysts and
 480 intentionally dissipative uses (on average 3.3 and 2.3 times, respectively).

481 There are different reasons for this pattern. Recent advances in defence and aerospace have escalated
 482 the global demand for magnets (Sinha et al., 2017), which could explain the noticeable increase in both
 483 manufacturing and use of this application in the EU. In the case of superalloys, the increase in both
 484 manufacturing and use has mainly been due to the increasing demand for superalloys in aerospace. In
 485 2017, the global commercial and military aerospace markets were estimated to account for 67% of the
 486 total superalloy demand. Since 2010, the global demand for superalloys has increased, not only due to
 487 the higher production rate of aircrafts but also due to the so-called new-generation engines. However,
 488 the strong aerospace demand has been offset by a sharp drop in the demand for industrial gas turbines
 489 used in power generation due to the global rise in renewable energy from solar and wind generation
 490 (Darton Commodities Limited, 2018). Although the demand for superalloys in aerospace is forecasted

to further increase in the coming decades (Airbus, 2017; Darton Commodities Limited, 2018), there are some factors that could cause the demand to remain constant or even decrease, specifically legal and environmental drivers related to aviation (Monnet and Abderrahim, 2018). Direct emissions from aviation account for about 3% of the EU's total greenhouse gas emissions and more than 2% of the global emissions. Furthermore, to date aviation emissions are already around 70% higher than in 2005, and it is predicted that by 2050 they could grow by over 300% (European Commission, 2020b). As some specialists have already indicated (Capoccitti et al., 2010; Bock and Burkhardt, 2019), improving technological or logistical aspects, such as fuel efficiency, engine technology, and traffic management, might not be enough to stop this trend. Hence, if governments enforce stronger measures for climate change mitigation, a decrease in demand could occur.

In the same context of climate change mitigation, the demand for batteries has changed and is expected to change further. The increasing use of mobility batteries is in line with the EU's policy objective of reducing greenhouse gas emissions from transport. Based on the Paris Agreement's goals and the EU Batteries Alliance, different studies and reports forecast that the Co demand for this sector will continue increasing at both the EU and global levels (Alves Dias et al., 2018; Deetman et al., 2018; IEA, 2018; Neef and Thielmann, 2018; Tercero Espinoza et al., 2019). According to recent data, in 2030 the use of Co for mobility in the EU will be between 29,000 and 53,000 t of Co, of which the EU will manufacture 20,000 to 47,000 t. This means that between 2018 and 2030 the demand will increase between 3,700 and 8,900% for manufacturing and between 3,600 and 6,700% for use. However, based on the research of new battery chemistries, other scenarios foresee a decrease of the demand of Co in batteries. For example, the Federal Environment Agency (UBA) of Germany assumed in its RESUCE (Resource-Efficient Pathways towards Greenhouse-Gas-Neutrality) study (Günther et al., 2019) that lithium-sulphur batteries will fully replace lithium-ion batteries containing Co from 2040 onwards. Lithium-sulphur batteries are already in the market, and are one of promising candidates for next-generation energy storage device due to the sulphur cathode material with low cost and nontoxicity (Zhou et al., 2019). Another cathode chemistry was recently developed (Li et al., 2020) based on nickel, manganese, and aluminium, abbreviated NMA. The Co-free cathode material presents overall attractive electrochemical properties, which were benchmarked against NMC and NCA for lithium-ion batteries. According to the authors, this research opened a new space for cathode material development for next-generation high-energy, Co-free lithium-ion batteries.

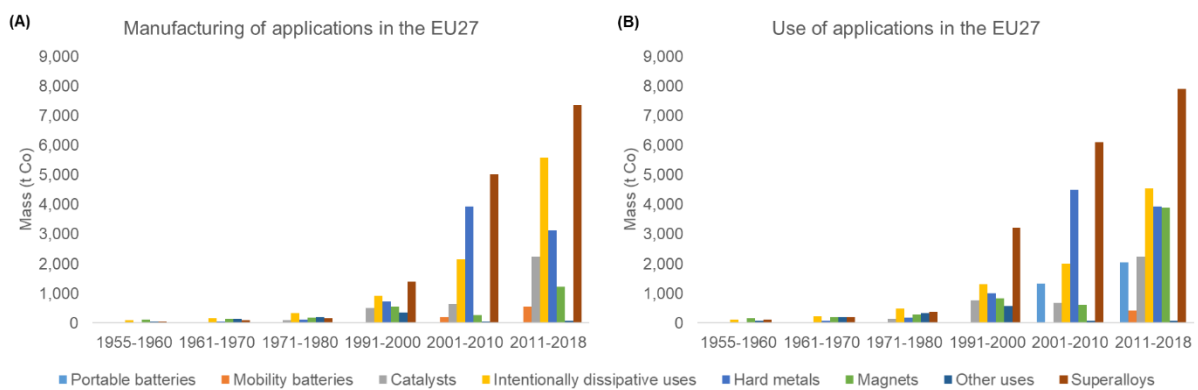
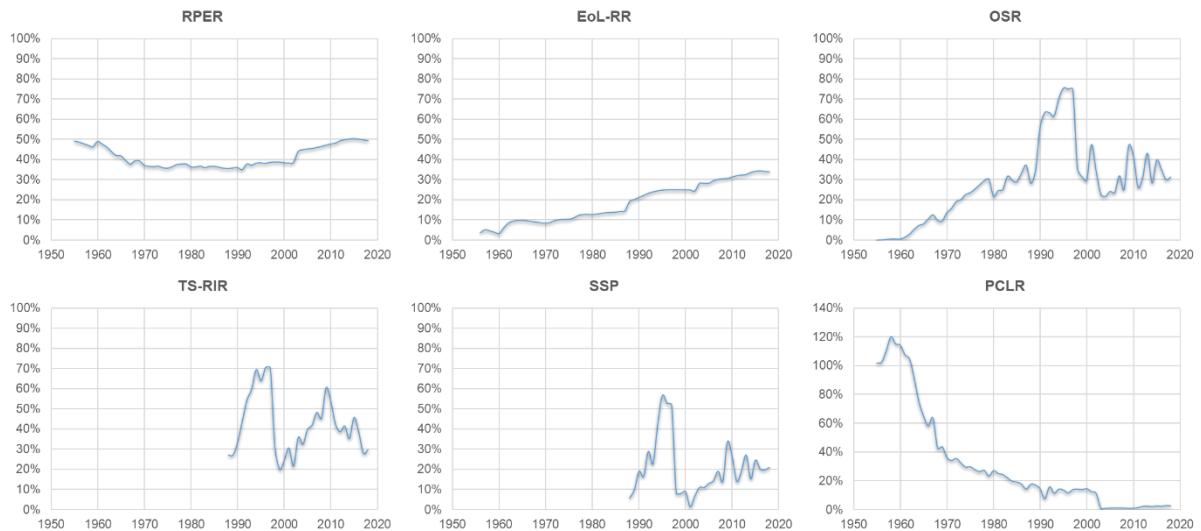


Figure 5. (A) Manufacturing of applications of Co in the EU. (B) Use of applications of Co in the EU.

The evolution of the societal metabolism of Co can also be analysed through comparison of the calculated indicators, which are presented in Figure 6. For the indicators TS-RIR and SSP, the results are shown from 1988 onwards due to data availability related to the trade of commodities.

The RPER has fluctuated between 35 and 50%, showing the lowest values between 1970 and 1990. This behaviour is mainly due to the temporal evolution of the demand of final products, which affects the amount of scrap recovered from the processing and manufacturing phases, and the amount of scrap recovered from EoL products. From 2002 onwards, the modelling assumes a higher recovery of scrap

530 from the manufacturing phase and a lower rate of downcycling for superalloys, which results in higher
 531 values of the RPER after 2000. The EoL-RR has increased steadily over time, meaning that more
 532 secondary Co has been recovered from domestically collected and imported EoL products.
 533 Nevertheless, there is room for improvement to increase the circularity of Co, as in 2018 the EoL-RR
 534 was estimated at 34%.



535
 536 *Figure 6. Value for the set of indicators in percentage (%). Recycling process efficiency rate (RPER), end-of-life*
 537 *recycling rate (EoL-RR), old scrap ratio (OSR), total scrap recycling input rate (TS-RIR), self-sufficiency potential*
 538 *(SSP), and pre-consumer loss rate (PCLR).*

539 Regarding the OSR, an increasing trend is observed between the 1960s and end of the 1990s. After
 540 the 2000s, the OSR fluctuates, with an average value of 32%. This behaviour is explained by the change
 541 in the demand of final products, shifting to applications from which more new scrap can be recovered
 542 during the processing and manufacturing phases. The TS-RIR shows an increasing trend between the
 543 end of the 1980s and the end of the 1990s and between the beginning of the 2000s and the end of the
 544 2000s. The behaviour of this indicator is mainly influenced by trade dynamics, as it depends on the
 545 import of processed and semi-processed material. In addition, as with RPER and OSR, it is affected by
 546 the change in the demand of final products. The SSP also shows an increasing trend between the end
 547 of the 1980s and the end of the 1990s and between the beginning of the 2000s and the end of the
 548 2000s. Afterwards it fluctuates, with an average value of 21%. In 1995, it reached its peak at 57%. High
 549 values of this indicator are due to a high return of recycled Co from EoL products to the processing and
 550 manufacturing phases. The input of primary material is generally negligible.

551 Finally, it is observed that the PCLR has decreased over time. However, even though efficiency
 552 improvements were made to the model when possible, the main reason for this decrease is an
 553 increasing end-use demand. This result has to be considered carefully since the EU also relies on
 554 imports in the use phase, but the PCLR only reflects domestic pre-consumer losses, and it does not
 555 consider the losses occurring outside the EU. The value of the PCLR in the 1950s is higher than 100%.
 556 This is related to the export of processed material. In that period, the processing losses were high due
 557 to the production of processed Co that was subsequently exported, with the total losses surpassing the
 558 amount used domestically in final products.

559 3.3 Anthropogenic and lost stocks over time

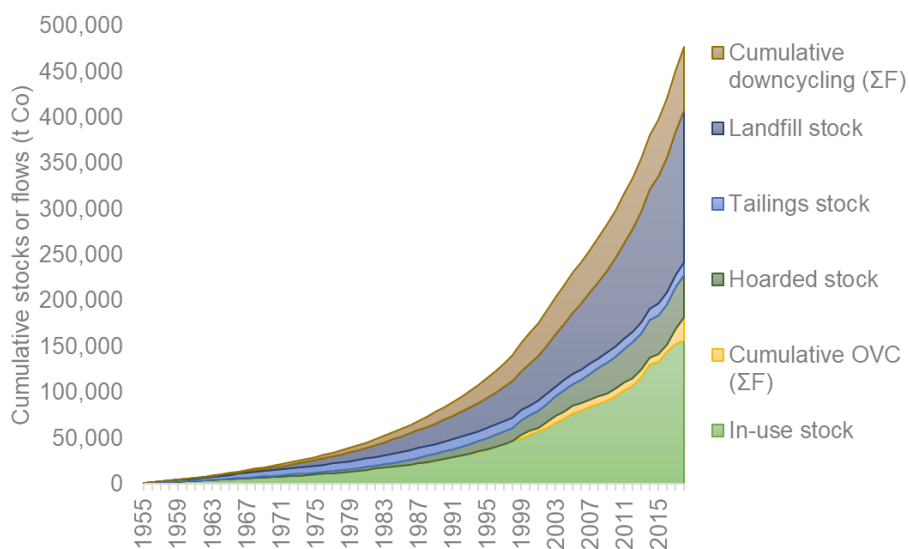
560 In this section, the cumulative stocks and flows of Co are discussed. Figure 7 presents the accumulation
 561 between 1955 and 2018 for six different destinations of Co: in-use (stock), OVC (cumulative flow),
 562 hoarded (stock), tailings (stock), landfill (stock), and downcycled (cumulative flow) (see section 2.2 for
 563 an explanation of each).

564 In 2018, the landfill stock was the largest, with around 166,000 t of Co, representing 35% of the total
 565 accumulation in that year. The main contribution to this stock comes from the non-selective collection;
 566 in 2018, almost 100,000 t of Co were stocked in landfills due to this stream. The second biggest stock
 567 or cumulative flow was the in-use stock, which was around 157,000 t of Co in 2018 (33% of the total
 568 accumulation). The smallest stock or cumulative flow in 2018 was the tailings stock and the cumulative
 569 output from the value chain, representing 3 and 5% of the total accumulation in that year, respectively.

570 The in-use stock increased at an average rate of 33% per year between 1955 and 1960, 11% per year
 571 between 1961 and 1970, and about 7% per year from 1971 until 2018. In other words, after 1970 the
 572 in-use stocks doubled in size every 10 or 11 years.

573 The cumulative OVC (which refers to stockpiled Co and/or Co that has left the EU through confidential
 574 exports) started to build up in the late 1980s, but only at the end of the 1990s did it become substantial,
 575 with around 4,000 t of Co in 1999. In 2006, it doubled (about 8,000 t of Co), and in 2017 and 2018 a
 576 high value chain output was estimated, increasing the cumulative flows to 17,000 and 25,000 t of Co,
 577 respectively. Here it is important to mention that the model does not consider 'de-stockpiling' since, to
 578 the best of the author's knowledge, there is no information available.

579 The hoarded stock doubled its size every 8 to 11 years between 1970 and 2010, growing at an average
 580 value of 8% per year. However, between 2010 and 2018 its average growth rate decreased to 3.6%
 581 per year. This decrease was due to a shift in the end-use demand for applications with lower hoarding
 582 rates and/or hoarding times. Between 2012 and 2018, the end-use demand for mobility batteries,
 583 catalysts, intentionally dissipative uses, and magnets increased, with a decrease in the demand for
 584 portable batteries, hard metals, and superalloys. The latter are indeed the applications with the highest
 585 hoarding rates and/or hoarding times.



586
 587 *Figure 7. Cumulative stocks or flows of Co in the EU27 from 1955 until 2018. OVC: output from the value chain.*
 588 *ΣF: cumulative flow.*

589 The tailings stock has evolved according to the periods in which primary material was produced
 590 domestically (for more details, see Godoy León et al., 2021), but in general after 1970 it increased at a
 591 rate between 1 and 6% per year. Only between 1955 and 1967 did it show higher growth rates, between
 592 11 and 61% per year.

593 Regarding the landfill stock, it exhibits a similar behaviour to that of the in-use stock; it increased at a
 594 higher rate between 1955 and 1960 (49% in average per year), decreasing between 1961 and 1970
 595 and stabilising after 1971 (25% on average per year between 1961 and 1970 and 9% on average per
 596 year between 1971 and 2018). This means that the landfill stock doubled in size on average every eight

597 years. Finally, the cumulative downcycling flow, which refers to the aggregated flow of Co lost by
598 downcycling, increased at an average rate of 41% per year between 1955 and 1960, 18% per year
599 between 1961 and 1970, 10% per year between 1971 and 2002, and 4% per year between 2003 and
600 2018. The lowest rate after 2002 was due to the higher recovery of Co from functional recycling,
601 specifically from EoL superalloys.

602 The main products of the in-use stock in 2018 were intentionally dissipative uses, superalloys, magnets,
603 and hard metals: 29, 26, 19, and 15% of the total stock in-use, respectively. The high volume of stock
604 observed for products containing superalloys is consistent with the higher lifetime characteristic of these
605 products. One important consequence of this result is that, for more than a quarter of the actual in-use
606 stock, Co could be retained in use within products with lifetimes as high as 30 years. Regarding
607 batteries, the high expected demand for electric vehicles will increase their in-use stock in the near
608 future. According to RMIS (2019), the stock of Co in mobility batteries will more than double in 2021,
609 compared to the values reported in 2016. In contrast, a high volume of the in-use stock (29%) is
610 embedded in products in dissipative uses, which means that part of this amount will be lost from the Co
611 cycle after use (some dissipative uses are landfilled after use, while others are dissipated in the
612 environment).

613 It is important to keep in mind that the presented stocks and cumulative flows do not consider inputs
614 before 1955. Nevertheless they would not likely contribute significantly to the current stocks and
615 cumulative flows, mainly due to the magnitude of the Co flows before the 1990s.

616 In terms of circularity, it was observed that the efforts should be focused on differentiated collection and
617 recycling. In addition, the in-use and hoarded stocks and related flows provide information about how
618 much Co will be available in the coming years. For example, the amount of Co in the hoarded stock
619 (about 47,000 t of Co) at the current collection and recycling rates could be sufficient to cover the need
620 for processed Co for a whole year. In terms of the EoL flows, in 2018 the hoarded and the non-
621 selectively collected Co could have covered around 70% of the processed Co demand, increasing the
622 self-sufficiency of the EU from 21 to 80%. In both examples, it is assumed that the recycled Co has
623 sufficient quality for all the applications. These results are relevant for policy decision-making, especially
624 regarding the new Circular Economy action plan of the EU (European Commission, 2020c). One of the
625 pillars of the plan is circularity in production processes, which promotes the use of digital technologies
626 for tracking, tracing, and mapping of resources. In addition, the plan is focused in key product value
627 chains, such as ICT (information and communications technology), and batteries and vehicles.

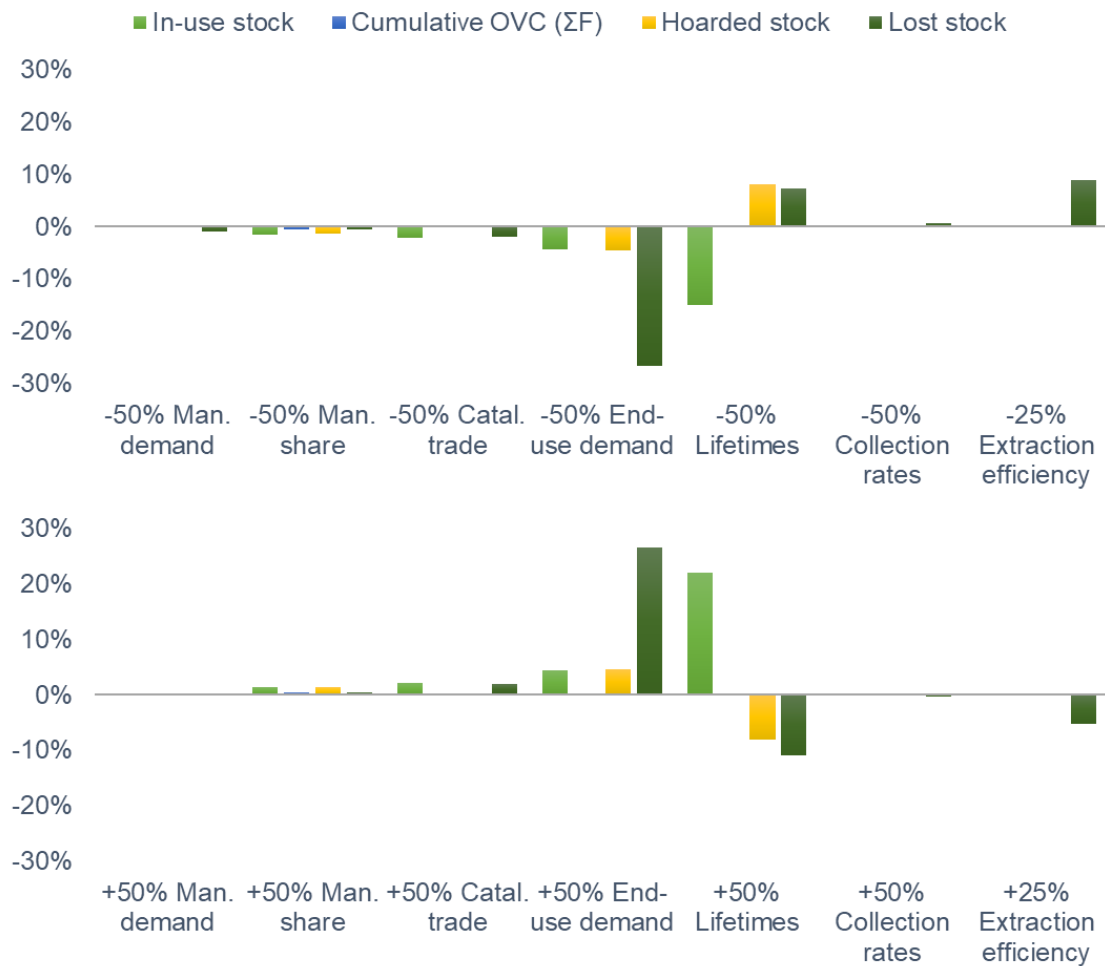
628 3.4 Limitations and sensitivity analysis

629 The calculation of the presented stocks and flows is subject to limitations. For instance, production–
630 inventory problems are not considered. Here, it is assumed that the amount produced is equal to the
631 demand. This aspect is more related to the management of single companies, but it could be considered
632 in future research to determine how it affects the stocks and flows of Co.

633 Other limitations are related to the assumptions considered in the study. For example, the tailings were
634 calculated using a constant value of 79% for mining extraction, which is a considerable assumption
635 since the ore grade and extraction efficiency can vary over time according to the characteristics of the
636 mine and available technology. Due to data unavailability, the in-use flows and stock were calculated
637 using extrapolated values for the manufacturing and end-use demand, assuming exponential growth
638 between 1955 and 1987 and between 1955 and 2002, respectively; this is in addition to assumptions
639 related to the share of manufacturing demand and the trade of catalysts. Similarly, the EoL flows and
640 related stocks and cumulative flows were estimated assuming relatively constant values of the lifetime
641 and the collection rate of the applications. All of these assumptions may have a low or high impact on
642 the final results. This impact was examined through a sensitivity analysis (Figure 8).

643 Eight sets of parameters were analysed (see Table 2). Figure 8 presents the results when the value of
644 the parameters changes by $\pm 50\%$ ($\pm 25\%$ for the extraction efficiency) compared with the base case.

645 The results indicate the variation of the in-use stock, cumulative OVC, hoarded stock, and lost stock
 1 646 (including the cumulative downcycling flow) in 2018. The results related to the net trade of functionally
 2 647 recycled material are not shown in Figure 8 since the value of that parameter does not influence stocks
 3 648 in 2018 but rather the trade of primary and semi-processed Co between 1955 and 1987. Between these
 4 649 years, the trade of these commodities varies on average 3.8% compared with the base case (3.8%
 5 650 higher when there is a net export of recycled Co, and 3.8% lower when there is a net import of recycled
 6 651 Co).



652
 653 *Figure 8. Variation in the percentage of the in-use stock, cumulative OVC (output from the value chain), hoarded*
 654 *stock, and lost stock in 2018 when changing the value of seven sets of parameters. Above, the value of the*
 655 *parameters decreased by 50% (25% for extraction efficiency); below, the value of the parameters increased by*
 656 *50% (25% for extraction efficiency). Man. demand: Manufacturing demand, Man. share: Manufacturing demand*
 657 *share, Catal. trade: Net trade of catalysts. ΣF: cumulative flow.*

658 Regarding the other seven sets of parameters, those with the highest influence on the 2018 stocks are
 659 the end-use demand of Co and the lifetimes of the applications. The rest of the assessed parameters
 660 have a minor impact on the stocks (less than 10% in absolute terms). The end-use demand of Co mainly
 661 affects the lost stock, which decreases (increases) by 27% compared with the base case when the
 662 value of the end-use demand decreases (increases) by 50%. A higher Co demand by end-users implies
 663 higher flows of EoL Co. Indeed, it is observed that the losses that increase the most are those related
 664 to non-selective collection, recycling waste, and downcycled material in the recycling phase. Since the
 665 value of the end-use demand changes between 1955 and 2002, its effect on the in-use stock in 2018
 666 is marginal (around 4%); by 2018, many of the applications that entered the use phase during that
 667 period have already reached their EoL.

668 Related to the lifetime of the applications, when their value decreases by 50%, the in-use stock in 2018
669 decreases by 15%, and the lost stock increases by 7%; when their value increases by 50%, the in-use
670 stock increases by 22%, and the lost stock decreases by 11%. If the lifetime of the applications
671 increases, they remain in the in-use stock longer, making it bigger over time; in addition, fewer EoL
672 products are produced, decreasing the EoL losses (and the other way around). Even though the lifetime
673 values only change between 1955 and 1987, the change is strong enough to influence the in-use stock
674 in 2018. This means that the in-use stock is more sensitive to a change in the lifetimes than to the end-
675 use demand.

676 A comparison of our findings with those of previous studies can help us to understand the accuracy of
677 the data. As indicated in section 1, Bio by Deloitte (2015) and Matos et al. (2020) assessed the stocks
678 and flows of Co in the EU28 in 2012 and in the EU27 in 2016, respectively. When comparing our results
679 with those of BIO by Deloitte, some differences are found in the processing and manufacturing phases.
680 According to that study, in 2012 the EU28 produced around 15,000 t of refined Co and manufactured
681 around 11,000 t of Co in final products. Our results show that in 2012 the EU27 produced around 12,400
682 t of refined Co and manufactured 17,800 t of Co in final products. However, for the use and collection
683 phases, the differences are smaller. For example, the end-use demand and the collection of EoL
684 products each show a difference of about 6% between both studies. When compared with the results
685 of Matos et al. (2020), the differences are lower for the processing and manufacturing phases (0.1%
686 and 7%, respectively) but higher for the use and collection phases. For 2016, Matos et al. (2020)
687 estimated that the EU27 consumed around 33,000 t of Co in final products, and 20,000 t of Co were
688 collected in EoL products. Our calculations show that in 2016 the EU27 consumed 28,300 t of Co in
689 final products, and 16,600 t were collected.

690 There are various reasons for these differences. First, the geographical coverage is different, as BIO
691 by Deloitte developed the study for the EU28 whereas this study was developed for the EU27. Second,
692 the data used in the three studies are different; for instance, here we used analysed and complemented
693 statistical data on the production and trade of the different commodities and applications. In BIO by
694 Deloitte and Matos et al., the considered statistical data were not scrutinised. In addition, both previous
695 studies were developed using a static approach whereas we used a dynamic approach. This made it
696 possible to introduce lifetime distributions and estimate stocks and flows considering changes in the
697 past. For example, in BIO by Deloitte and Matos et al., the calculation of the collection of EoL products
698 depended on constant annual growth rates of consumption and on the manufacturing and trade flows
699 of the applications of the assessed year. In other words, it was assumed that the production and trade
700 of the applications in the past follow the same trend as in the present. Here, the dynamics over time
701 were considered in the calculation.

702 Despite these differences, the values in this research, in general, vary less than 20% from those of the
703 previous studies, which is acceptable given the data availability and quality. However, there is an
704 important discrepancy in the estimated downcycled Co. On one hand, BIO by Deloitte did not consider
705 any downcycling of the metal. On the other hand, Matos et al. estimated the downcycled flow as 4,000
706 t of Co, which is almost 85% higher than the flow estimated here. This difference mainly comes from
707 the estimation of EoL magnets, which is 75% higher in Matos et al. In the authors' opinion, this flow was
708 overestimated in Matos et al. due to the applied methodology. Consequently, it is recommended that in
709 future research EoL flows should be estimated using a historical dynamic approach.

711 4. Conclusions

712 A dynamic material flow analysis (dmFA) was applied for long-term retrospective assessment of the
713 cobalt (Co) societal metabolism in the EU27. The assessment was done between 1955 and 2018.

714 It was found that the Co stocks and flows have increased substantially over time. For example, between
715 1988 and 2003 the inputs to the processing phase increased 6.6 times. Between 2003 and 2018, the

716 inputs to the different phases also increased considerably, but generally at a lower rate. Nevertheless,
717 when comparing the end-use demand per capita it can be seen that, on average, it increased at a
718 similar rate between 1988 and 2003 (7.2%) and between 2003 and 2018 (7.4%). In addition, the growth
719 rates indicate that the Co demand is growing at a considerably faster pace as GDP and as a
720 consequence, the EU's economy has become more intense in the use of Co over time, from 0.7 g Co
721 per €1,000 of GDP in 1995 to 2.1 g Co per €1,000 of GDP in 2018.

722 The EoL stocks and flows have also increased over time. In particular, the flows of functionally recycled
723 and downcycled Co have increased significantly, although the share of functionally recycled material
724 has increased, while the share of downcycled material has decreased. This change has been mainly
725 due to shifts toward some applications for which EoL products are often downcycled. In 2018, the share
726 was 71% of functionally recycled Co and 29% of downcycled Co.

727 According to the results, the landfill stock was the largest one in 2018, with around 166,000 t of Co,
728 accounting for 35% of the total stock in that year. The main input to this stock was the non-selective
729 collection, which in 2018 had stocked about 100,000 t of Co in landfills. In terms of circularity, landfills
730 might be both a big opportunity and a challenge for the recovery of Co in the coming years. According
731 to Dewulf et al. (2021), in the best scenario material could start to be recovered from landfills in the
732 coming decades. If the entire landfilled stock of Co could be identified, accessed, and recovered, and
733 assuming a hypothetical efficiency of 50%, it could cover the EU27 material demand for about four
734 years (assuming a demand of 20,000 t of Co, as in 2018). In a similar vein, in 2018 the hoarded and
735 non-selectively collected Co could have covered around 70% of the processed Co demand, increasing
736 the self-sufficiency of the EU from 21 to 80%. Therefore, it is clear that enhancing selective collection
737 and switching from downcycling to recycling is key for increasing the circularity of Co, reducing supply
738 risks at the same time.

739 This work is novel in multiple ways. First, it is the first long-term retrospective dMFA of Co developed
740 for the EU. Second, it offers an updated view of the metabolism of Co in the EU27 for 2018. Third, the
741 calculation of stocks in previous assessments was based on assumptions and applied values of market
742 levels some years before the assessment. Here, historical long-term statistical data were applied. In
743 addition, data from the Eurostat and UN Comtrade databases were used separately, comparing the
744 results when using one or the other, and substantial differences were identified.

745 The results of our research can be useful for policy makers, industries, and researchers. They establish
746 how the stocks have built up over time and where the main reserves of secondary Co in the EU are
747 located, which could be exploited in the coming years. In addition, they indicate where efforts should
748 be made to increase the circularity of Co, decreasing dependency on imported material. Finally, the
749 results could be used in a prospective dMFA, modelling different scenarios to determine the potential
750 future of Co in the EU in the coming decades.

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