

Evaluation of Fenton and modified Fenton oxidation coupled with membrane distillation for produced water treatment: Benefits, challenges, and effluent toxicity

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1 **Experimental evaluation of Fenton and modified Fenton**
2 **oxidation coupled with membrane distillation for**
3 **produced water treatment: Benefits, challenges, and**
4 **effluent toxicity**

5
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19

20 **ABSTRACT**

21 Membrane distillation is a promising technology to desalinate hypersaline produced waters.
22 However, the organic content can foul and wet the membrane, while some fractions may pass
23 into the distillate and impair its quality. In this study, the applicability of the traditional
24 Fenton process was investigated and preliminarily optimized as a pre-treatment of a synthetic
25 hypersaline produced water for the following step of membrane distillation. The Fenton
26 process was also compared to a modified Fenton system, whereby safe iron ligands, i.e.,
27 ethylenediamine-N,N'-disuccinate and citrate, were used to overcome practical limitations of
28 the traditional reaction. The oxidation pre-treatments achieved up to 55% removal of the
29 dissolved organic carbon and almost complete degradation of the low molecular weight toxic
30 organic contaminants. The pre-treatment steps did not improve the productivity of the
31 membrane distillation process, but they allowed for obtaining a final effluent with
32 significantly higher quality in terms of organic content and reduced *Vibrio fischeri* inhibition,
33 with EC₅₀ values up to 25 times those measured for the raw produced water. The addition of
34 iron ligands during the oxidation step simplified the process, but resulted in an effluent of
35 slightly lower quality in terms of toxicity compared to the use of traditional Fenton.

36

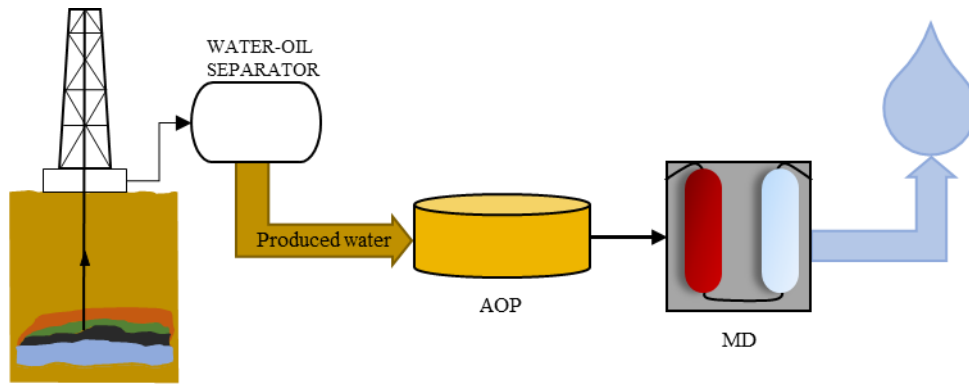
37 **Keywords:** membrane distillation; produced water; advanced oxidation; iron ligands;
38 ecotoxicity

39 **HIGHLIGHTS**

- 40 • Thermal and modified Fenton degraded target contaminants in produced water.
- 41 • The oxidative pre-treatment reduced organics in the membrane distillation effluent.
- 42 • Coupled oxidation and membrane distillation reduced the toxicity of the final effluent.
- 43 • Traditional Fenton pre-treatment provided the best effluent in terms of toxicity.
- 44 • Modified Fenton nearly degraded all target contaminants in hypersaline solutions.

45

46 **GRAPHICAL ABSTRACT**



47

48

49 **1. Introduction**

50 Despite the current transition to more sustainable sources of energy, oil and gas extraction
51 still plays a significant role in the energy sector and wastewater treatment is quickly emerging
52 as one of the most significant challenges of this industry. Indeed, the so-called produced
53 water (PW) is the largest waste stream generated in oil and gas extraction activities
54 (Ahmadun et al. 2009). Considering both onshore and offshore sites, the global PW
55 production has increased from around 150 million to around 300 million barrels per day from
56 1990 to 2015 (Ahmadun et al. 2009; Liu et al. 2021). The average water cut, namely, the
57 amount of water volume produced per oil volume, is roughly 3:1 (McCormack et al. 2001;
58 Ahmadun et al. 2009; Jimenez et al. 2018; Liu et al. 2021). As oilfields age during the current
59 energy transition, the water cut will also increase, together with wastewater treatment
60 difficulties (Igunnu and Chen 2014). PW is a highly complex matrix, rich in organic and
61 inorganic compounds, with widely diverse composition as a function of geological formation,
62 age of the oilfield, and type of hydrocarbon product being produced (Ahmadun et al. 2009;
63 Estrada and Bhamidimarri 2016). However, the major compounds are typically dispersed
64 oils, dissolved organics (e.g., phenols, benzene, toluene, xylenes), dissolved minerals (e.g.,
65 sodium chloride, calcium and magnesium salts), and natural organic matter (NOM) (Neff et
66 al. 1992; Ahmadun et al. 2009; Estrada and Bhamidimarri 2016; Jimenez et al. 2018; Al-
67 Ghouti et al. 2019; Kabyl et al. 2020; Cocha et al. 2021; Liu et al. 2021).

68 The oil and gas industry faces increasing pressure to limit its environmental footprint
69 (Mohammad-Pajooch et al. 2018). Expensive treatment trains for a multi-contaminated water,
70 as well as water scarcity and increasing international attention to environmental issues, are
71 the drivers pushing this industry to use water more sustainably, bringing along the concepts
72 of water reuse and safe water discharge. As a result, innovative, environmentally focused,
73 and reliable methods of meeting water treatment demands, capable of operating in this

74 specific application are being developed (Estrada and Bhamidimarri 2016; Mohammad-
75 Pajoo et al. 2018; Liu et al. 2021). These technologies should be versatile to meet the
76 requirements of low cost and compactness, the latter characteristic being especially important
77 in offshore activities (Kabyl et al. 2020; Liu et al. 2021).

78 One of the target parameters that needs abatement to allow for water reuse or safe
79 discharge is salinity, which presents an average value of 100,000 ppm in PW (Estrada and
80 Bhamidimarri 2016; Coxa et al. 2021). Such salt concentration can plug the reinjection well
81 or may be toxic if PW is discharged in the environment without desalination (Kleinitz et al.
82 2001; Aquiliina 2012; Ariono et al. 2016; Chen et al. 2016; Canedo-Arguelles et al. 2019; Liu
83 et al. 2021). Membrane distillation (MD) is a promising emerging technology, capable to
84 extract high-quality effluents from hypersaline solutions using low-grade energy and with
85 relatively low capital cost, due to the absence of high pressure and high temperature
86 components (Howell 2004; Shaffer et al. 2013; Lin et al. 2014; Chen et al. 2017; Han et al.
87 2017). In recent studies, MD was successfully tested on hypersaline PW (Han et al. 2017;
88 Ricceri et al. 2019). However, since the MD membranes are highly hydrophobic (PTFE and
89 PVDF membranes are generally used), this process may present important practical
90 limitations in the presence of a large and broad content of organic compounds, such as for
91 typical PW (Estrada and Bhamidimarri 2016; Gonzalez et al. 2017). Organics may either
92 induce wetting phenomena or freely pass through the hydrophobic membrane and end up in
93 the final effluent (Franken et al. 1987; Kargbo et al. 2010; Chen et al. 2017; Wang et al.
94 2018). Wetting phenomena in MD occur when the transmembrane pressure (ΔP) exceeds the
95 liquid entry pressure (LEP), according to eq. (1), thus allowing the contaminated feed water
96 to pass undisturbed through the porous membrane (Franken et al. 1987; Ricceri et al. 2019;
97 Horseman et al. 2021).

$$98 \quad \Delta P \geq LEP \quad (1)$$

99 LEP is defined according to eq. (2):

$$100 \quad LEP = -\frac{2B\gamma\cos\theta}{r} \quad (2)$$

101 where γ is the feed water surface tension, θ is the intrinsic contact angle between the feed
102 water and the solid membrane material, r is the equivalent pore radius, and B is a geometric
103 factor accounting for the noncylindrical nature of the membrane pore geometry ($B = 1$ for
104 perfectly cylindrical pores). A low surface tension γ reduces also $\cos\theta$, thus facilitating
105 membrane pore wetting. The degradation of toxic organic compounds in the feed solution
106 both increases its surface tension and thwart their interaction with the hydrophobic
107 membrane, potentially allowing for a more efficient MD process and a more effective
108 management of the PW.

109 The list of available technologies and processes counts a plethora of options to remove or
110 partially degrade organic compounds (Adewumi et al. 1992; Ahmadun et al. 2009; Estrada
111 and Bhamidimarri 2016; Chang et al. 2019; Chang et al. 2019; Shang et al. 2019; Liu et al.
112 2021; Tang et al. 2021). Activated carbon adsorption and sand filtration are low-cost
113 treatment processes; however, they produce harmful waste since they do not degrade the
114 toxic organic compounds. Biological treatment is not effective toward biorecalcitrant organic
115 compounds, such as benzene, toluene, xylenes (BTX), and requires large plants with long
116 retention times, hence not available for offshore platforms (Ayed et al. 2017).
117 Electrochemical, photocatalytic, and ozone-based oxidations are growing rapidly, but they
118 are still currently associated with high capital costs and with difficulties in practical
119 implementation (Dalmacija et al. 1996; Bessa et al. 2001; Ma and Wang 2006; Ahmadun et
120 al. 2009; Shokrollahzadeh et al. 2012; Ricceri et al. 2019; Cocha et al. 2021). Among
121 advanced oxidation processes (AOPs), the Fenton reaction involves the use of iron sulfate
122 and hydrogen peroxide to generate highly reactive hydroxyl radicals able to oxidize almost

123 all the organic compounds (Haber et al. 1934; Miklos et al. 2018; Coha et al. 2021). The main
124 reaction is as follows:



126 The Fenton process is a promising candidate to treat PW both onshore and offshore since it is
127 versatile, characterized by high kinetics also at room temperature and capable to remove
128 organics from a multi-contaminated matrix. This method has some limitations, mainly the
129 need for acidic pH to avoid iron hydroxide precipitation, and the production of sludge once
130 neutral pH is restored (Diya'uddeen et al. 2012). A modified Fenton process encompassing
131 the addition of an iron ligand helps overcoming these precise challenges (Chahbane et al.
132 2007; Farinelli et al. 2019; Messele et al. 2019; Farinelli et al. 2020). However, the literature
133 lacks reports about the application of Fenton processes carried out in presence of iron ligands
134 to treat PW.

135 The first objective of this work is to evaluate a coupled system including Fenton (or
136 modified Fenton) pre-oxidation and MD to desalinate PW and to allow for an easy
137 management of the final effluent. This sequence is applied to treat a synthetic PW that
138 mimics the effluent from primary treatment, which typically includes de-oiling and flotation
139 or sedimentation. A thermal Fenton reaction is first applied as a potential PW oxidation step
140 and as a pre-treatment for MD desalination. The study gives insight into the relationship
141 between the content of organics and the performance of the membrane distillation step by
142 comparing a raw feed stream with the feed subject to Fenton oxidation. Furthermore, the
143 performance of traditional Fenton is compared with that of modified Fenton systems. To this
144 purpose, non-toxic and biodegradable organic ligands, namely, citrate and EDDS, are added
145 in PW at unadjusted pH to assess the ability of iron-ligand complexes to act as effective
146 oxidation catalysts (Van Devivere et al. 2001; Tandy et al. 2006; Chen et al. 2019). The

147 safety of the final desalinated effluent from the coupled system is then fully evaluated
148 through acute toxicity measurements.

149 **2. Materials and Methods**

150 *2.1. Chemicals, membrane, and produced water preparation*

151 All the organic contaminants, the iron ligands, i.e., sodium citrate and EDDS, ferrous
152 sulfate (FeSO₄), hydrogen peroxide (30% w/w), HCl, and NaOH, were purchased from
153 Sigma-Aldrich (Milan, Italy). Sodium chloride, sodium sulfate, and sodium bicarbonate were
154 acquired from Carlo Erba (Milan, Italy). All the solutions needed for the acute toxicity
155 analysis, namely, the reconstitution, the diluent, and the osmotic solutions were purchased
156 from Modern Water (London, UK). The freeze-dried *Vibrio fischeri* culture was purchased
157 from Ecotox LDS (Cornaredo (MI), Italy). Type I ultrapure water was used for the
158 experiments. A commercially available polytetrafluoroethylene (PTFE) membrane (Aquastill,
159 Sittard, Netherlands) was deployed in MD filtration tests.

160 The composition of the synthetic PW was based on published values of real wastewaters
161 and is listed in **Table 1**, together with the resulting total organic carbon (TOC) and total
162 dissolved solids (TDS) values (Olsson et al. 2013; Estrada and Bhamidimarri 2016; Coha et
163 al. 2021). Humic acids and a liquid petroleum jelly consisting of paraffins were used as
164 representative compounds for natural dissolved organic matter and oil & grease, respectively
165 (Lester et al. 2015). Xylenes, benzene, toluene, and methyl *tert*-butyl ether (MTBE) were
166 selected as representative volatile organic compounds (VOCs) (Coha et al. 2021).
167 Cyclohexane was added as representative of the <C₁₀ hydrocarbon fraction (Lester et al.
168 2015; Estrada and Bhamidimarri 2016). Phenol was added as representative substance for the
169 common phenols content in PW. The TDS included sodium, calcium, and magnesium

170 chlorides. All the components were added into water and the matrix was sonicated at room
 171 temperature for 1 h to enhance solubilization and mixing.

172

173 **Table 1** Composition of the synthetic produced water, compared with the reference real
 174 streams. The matrix includes representative pollutants to mimic typical TOC and TDS values.

Parameter	Component	Synthetic produced water		Real produced water
		Concentration (ppm)	Equivalent TOC (ppm)	Concentration (ppm)
TOC	Paraffins	200	Not dissolved	Maximum ~500
	Humic acids	200	60	
	Cyclohexane	2	1.8	
	Phenol	2.5	1.9	
	Xylenes	1	1	
	Benzene	12	11.3	
	Toluene	4	3.4	
	MTBE	260	178	
	TOT	681.5	257.4	
TDS	Sodium chloride	100,000		Typically, 35,000 – 240,000
	Calcium chloride	2,500		
	Magnesium chloride	4,000		
	TOT	106,500		
pH	5.5			Average ~100,000

175

176 2.2. Oxidation conditions

177 All the oxidation reactions were performed at room temperature under gentle stirring for a
 178 total duration of 1 h, using different ratios of hydrogen peroxide and catalyst, intended as
 179 iron(II) in the case of the traditional Fenton process and as the complex ligand-iron(II) in the
 180 case of modified Fenton. To promote organics oxidation, three additions of hydrogen
 181 peroxide (0, 20, 40 min) were carried out, each one of a 1/3 aliquot of the desired total
 182 amount. The dosages of iron sulfate and hydrogen peroxide are listed in **Table 2** for the
 183 traditional Fenton system; in these cases, the pH of the synthetic PW was adjusted to ~3
 184 (HCl). At the end of the reaction, the pH was increased to ~10 by addition of NaOH and this

185 step caused the precipitation of $\text{Fe}(\text{OH})_3$. After the sedimentation of the precipitate at 4 °C
186 overnight, the supernatant was collected and used for analysis and as a feed matrix for the
187 following MD filtration tests.

188

189 **Table 2** Traditional Fenton dosages in tests operated at different oxidation conditions.

Entry	FeSO_4 (mM)	H_2O_2 (mM)
MQ H_2O	-	-
Produced water	-	-
Ox 1	0.5	5
Ox 2	5	50
Ox 3	5	25
Ox 4	1	25
Ox 5	1	50
Ox 6	0.1	5
Ox 7	5	100

190

191 To perform the modified Fenton oxidations, the iron-ligand complexes, namely, Fe-
192 EDDS and Fe-citrate were spiked in the synthetic PW from a stock solution of 0.1 M of
193 iron(II) and 0.1 M of individual ligand. The pH was not adjusted and was equivalent to ~4
194 upon addition of Fe-EDDS and to ~5 upon addition of Fe-citrate. No precipitate formation
195 was observed in these systems. The resulting samples were used for analysis and as feed for
196 the following MD filtration tests without further processing.

197 2.3. Membrane distillation tests

198 The MD tests were performed in direct contact configuration using a lab-scale batch
199 system (Ricceri et al. 2019). The feed and distillate streams were circulated counter-currently
200 on their respective sides of the membrane. A constant crossflow rate of 1.66 L/min (0.278
201 m/s crossflow velocity) was maintained during the tests. The housing cell comprised a 250-
202 mm long, 50-mm wide, and 2-mm deep rectangular channel for a total active membrane area
203 of 125 cm^2 . The flux across the membrane was computed by recording the change in weight

204 of the distillate tank in time through a computer-interfaced balance. Initial volumes of ~1.9 L
205 and 1 L were used for the feed and distillate streams, respectively, unless otherwise stated.
206 Water was used in the distillate side, with specific conductivity always below 20 $\mu\text{S}/\text{cm}$. The
207 specific conductivity in the distillate tank was measured continuously during each test by a
208 conductivity meter (COND 7+, XS Instruments, Italy). The temperature of the feed and
209 distillate tanks were maintained constant throughout the experiments, at respective values of
210 50 ± 2 and 25 ± 1 $^{\circ}\text{C}$, by means of a thermostatic water bath and a chiller.

211 2.4. Analytical methods

212 The TOC of the matrices was measured using a Shimadzu TOC-L analyzer (catalytic
213 oxidation on Pt at 680 $^{\circ}\text{C}$). The calibration was performed using standards of potassium
214 phthalate, $\text{NaHCO}_3/\text{Na}_2\text{CO}_3$. The headspace, solid phase microextraction technique (HS-
215 SPME) was chosen as extraction method before carrying out the GC-MS analysis. Following
216 each reaction experiment, the vials were left in a thermostatic bath at 50 $^{\circ}\text{C}$ for 10 min to
217 promote the transfer of all the relevant compounds into the gas-phase headspace. Then, a
218 SPME fiber (df 75 μm , fiber assembly carboxen/polydimethylsiloxane, Supelco) was inserted
219 through the septum of the cap and was left in the headspace for 10 min, before withdrawing it
220 for the subsequent GC-MS analysis. Samples were analyzed on an Agilent 6890 GC system
221 coupled with an Agilent 5973 mass selective detector (MSD). For the chromatographic
222 separation, a Zebron-5MS capillary column (30 m \times 0.25 mm \times 0.25 μm) was used. The
223 injection port temperature was 270 $^{\circ}\text{C}$, and the oven temperature program was set as follows:
224 35 $^{\circ}\text{C}$ for 5 min, followed by an increase to 260 $^{\circ}\text{C}$ at a rate of 15 $^{\circ}\text{C}/\text{min}$ (total run time 25.33
225 min). Helium was used as carrier gas at a constant flow of 1 mL/min, and the injector was
226 held in splitless mode. The interface temperature was 270 $^{\circ}\text{C}$ and the ionization energy was
227 70 eV. The molecular structures of the by-products were identified by means of mass
228 spectrum library.

229 The determination of the residual iron in solution was evaluated by a spectrophotometric
230 procedure adapted from previous literature (Harvey et al. 1955; Goncalves et al. 2020). The
231 total iron was determined by reducing the Fe(III) to Fe(II) with ascorbic acid (4×10^{-4} M) and
232 complexing the Fe(II) with o-phenanthroline (4×10^{-3} M) under acidic conditions (buffer pH =
233 3: H₃PO₄ 1 mM, NaH₂PO₄ 3 mM). The Fe(II) was determined without performing the
234 reduction step, and Fe(III) was obtained as the difference between total iron and Fe(II). The
235 calibration was obtained using a commercial standard solution of Fe(III) (1000 mg^{Fe}/L,
236 Sigma-Aldrich). The spectrophotometric analyses were performed using a Varian CARY 100
237 Scan double-beam UV–vis spectrophotometer, using quartz cuvettes with 10 mm path length
238 and working at a wavelength of 510 nm.

239 2.5. Toxicity analysis

240 All the toxicity experiments were performed with a Microtox Model 500 analyzer (Milan,
241 Italy). The analysis was performed by evaluating the bioluminescence inhibition assay using
242 the marine bacterium *Vibrio fischeri*. Samples were tested in a medium containing 2%
243 sodium chloride, and the luminescence was recorded after 5, 15, and 30 min of incubation at
244 15 °C. The luminescence inhibition percentage was determined by comparison with a non-
245 toxic control. The toxicity curves and the values of EC₅₀ were obtained from the software
246 (MicrotoxOmni). The pH of all the samples was adjusted in the range 6-8 before the analysis.
247 The method used for the toxicity analysis is the method APAT-IRSA 8030 (APAT 2003). All
248 the samples obtained at the end of the oxidation experiments were quenched with catalase in
249 order to avoid the detrimental effect of the residual hydrogen peroxide on the toxicity
250 measurements; see Figure S1 of the Supplementary Material (SM). Moreover, the samples
251 after the oxidation experiments were quantified for residual iron content and EDDS was
252 added in EDDS:Fe molar ratio of 1:1, to prevent the toxic effect of the residual iron in
253 solution (see Figure S2 in SM).

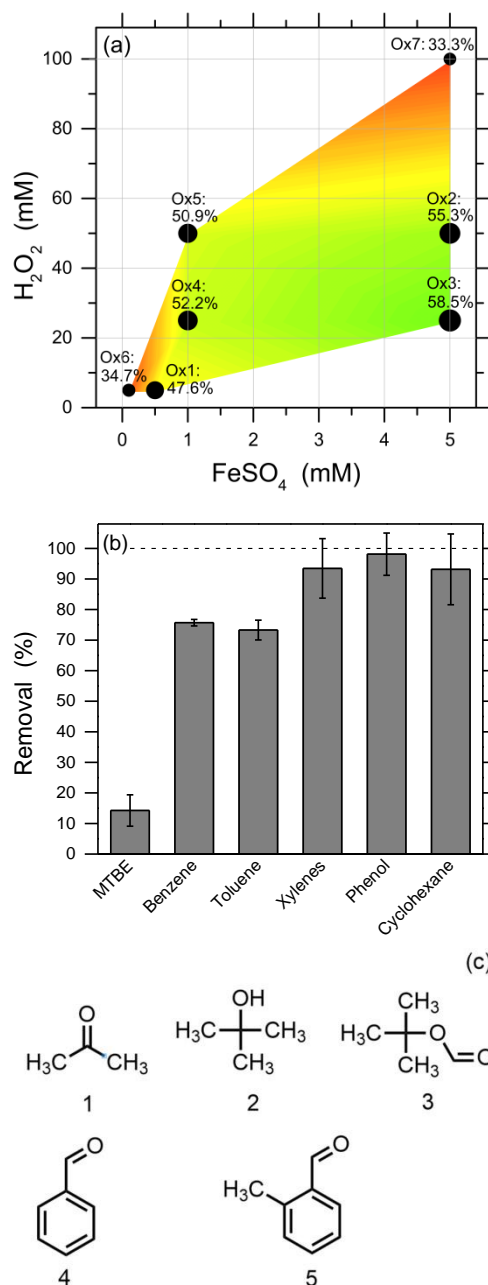
254 3. Results and Discussion

255 3.1. Efficacy of thermal Fenton oxidation on organics removal

256 **Table 3** and **Figure 1a** summarize the results of the Fenton oxidations in terms of TOC
257 removal and surface tension (ST) values, as a function of the relative addition of iron(II) and
258 H₂O₂. The highest TOC removal rates were obtained with Fe/H₂O₂ ratios between 0.02 and
259 0.2. At low iron dosage (≤ 0.5 mM, Ox1, Ox6), insufficient catalyst was available in solution,
260 while at high reagent concentrations (Ox7), the reaction was possibly self-inhibited. Previous
261 studies highlighted that a Fe/H₂O₂ molar ratio around 0.02 should avoid self-inhibition
262 reactions while providing high efficiency of oxidation (Voelker and Sulzberger 1996; De
263 Laat and Gallard 1999). Oxidation 3 (Ox3) reached the highest percentage of TOC removal,
264 coinciding with the largest dosage of FeSO₄ (5 mM) and a Fe/H₂O₂ ratio of 0.2. However,
265 Ox4 also achieved a high percentage of TOC removal, but with a substantially lower amount
266 of iron(II), namely, 1 mM, corresponding to a Fe/H₂O₂ ratio of 0.04. These conditions also
267 allowed reaching the highest value of ST (69 ± 3.1 dyn/cm), close to the ST measured for
268 pure water (72 ± 1.8 dyn/cm) and substantially higher than that of the PW (50.8 ± 2.7
269 dyn/cm). Ox4 was thus identified as the most promising oxidation and further tests were
270 conducted using the matrix oxidized under this condition.

271 **Table 3** Resulting TOC removal rates, and surface tension values of the oxidized matrix in
272 tests operated at different oxidation conditions.

Entry	TOC removal (%)	Surface tension (dyn/cm)
MQ H ₂ O	-	72 ± 1.8
Produced water	-	50.8 ± 2.7
Ox 1	47.6 ± 0.5	67.8 ± 1.6
Ox 2	55.3 ± 1.3	63.5 ± 1.8
Ox 3	58.5 ± 2.2	60.1 ± 2.3
Ox 4	52.2 ± 2.4	69.5 ± 3.1
Ox 5	50.9 ± 1.7	54.2 ± 2.1
Ox 6	34.7 ± 2.1	68.6 ± 2.0
Ox 7	33.3 ± 2.3	59.3 ± 1.7



273

274 **Figure 1.** (a) Graphical representation of the relationship between TOC removal and reagent
 275 dosage. The green color represents higher values of TOC removal. (b) Percentage of
 276 degradation of the parent organic contaminants in the synthetic produced water after Fenton
 277 reaction (Ox4, 60 min). The percentage of degradation was obtained by computing the target
 278 peaks area detected by GC-MS. (c) Chemical structures of the residual by-products
 279 preconcentrated onto the fiber during the SPME extraction and detected by GC-MS at the end
 280 of the Fenton reaction (Ox4, 60 min).

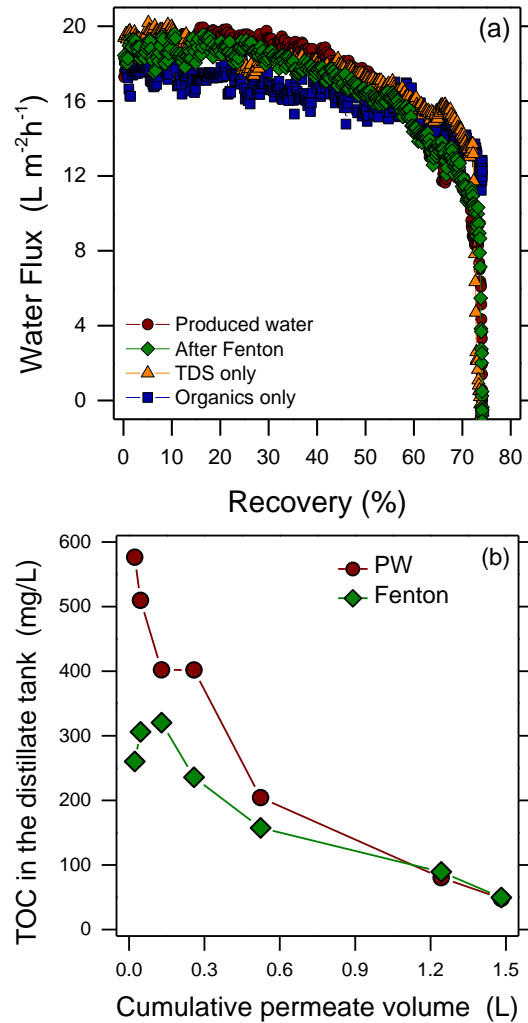
281 The solution obtained after Fenton oxidation Ox4 was further characterized through GC-
282 MS analysis. Figure S3 in the SM presents the chromatograms and the profile of target
283 substance degradation observed after 20, 40, and 60 min of oxidation. **Figure 1b** summarizes
284 the degradation efficiency after 60 min toward various organic contaminants. A near total
285 degradation of phenol, xylenes, and cyclohexane was observed. The oxidation process
286 degraded benzene and toluene with a yield around 75%. However, only a portion of MTBE
287 was degraded, and this result may also explain the residual TOC after Ox4 (see **Table 3**).
288 Indeed, MTBE contributes to a large part of the total TOC of the synthetic PW (see **Table 1**).
289 The major challenge for the high efficacy of the Fenton reaction in a hypersaline PW is
290 arguably the ability of humic acids and chloride to scavenge the hydroxyl radical (Kiwi et al.
291 2000; Goldstone et al. 2002), although humic acids may also favor the Fe(III)-Fe(II)
292 recycling (Vione et al. 2004). Nevertheless, the observed TOC removal rates and the yield of
293 degradation of target parent substances suggest the high potential of the Fenton reaction to
294 reach suitable levels of decontamination, also in the presence of a significant amount of
295 scavengers.

296 The GC-MS also allowed detection of the main volatile by-products of the Fenton process
297 (**Figure 1c**). Based on the molecular structure, the compounds labeled as 1, 2 and 3 in Figure
298 S3 reasonably derived from MTBE, while 4 and 5 likely derived from toluene and o-xylene,
299 respectively. Note that the identified by-products were more hydrophilic than the starting
300 contaminants, thus explaining the observed increase in ST. The relatively high percentage of
301 TOC removal together with the formation of more hydrophilic by-products are promising
302 conditions to obtain an improved feed solution of an MD step.

303

304 3.2. *Evaluation of thermal Fenton oxidation as a pre-treatment for membrane distillation*

305 Organic compounds in the feed matrix may affect the MD step by fouling and by wetting
306 the membrane, hence lowering the desalination efficiency, or by freely passing through the
307 hydrophobic membrane material (PTFE) (Vestervik et al. 2012; Pasternak and Kolwzan
308 2013). Wetting is theoretically described by eq. 1 and experimentally observed with an
309 increase of conductivity in the distillate stream of the MD step (Donaldson et al. 1969; Rezaei
310 et al. 2017). The water flux values presented in **Figure 2a** suggest that the Fenton pre-
311 treatment did not have an effect on the productivity or on the achievable recovery of the MD
312 step under laboratory filtration conditions: the recovery was roughly 75%, upon which the
313 water flux went to zero due to scaling (namely, the deposition of crystals on the membrane or
314 within its pores) and pore blockage. Note that similar productivity was also observed when
315 synthetic PW matrices containing solely salts or solely organics were used as the feed
316 solutions. In the latter case the flux did not go to zero, due to the absence of salt precipitation,
317 but it steadily decreased during the test possibly due to fouling phenomena. Increased
318 conductivity of the distillate solution was measured beginning roughly at 50% recovery for
319 the various feed streams, except that containing only organic compounds; see Figure S4 of
320 the SM. This result suggests that in our study salt passage and possibly wetting were mostly
321 imputable to high salt concentrations, with organics only associated with fouling
322 mechanisms. Salts can crystallize within the pores of the membrane, enlarging them, hence
323 lowering the LEP.



324 **Figure 2** (a) Results of MD filtration tests with different feed solutions: (red circles) synthetic
 325 produced water, (blue squares) only the organic content of the synthetic produced water,
 326 (orange triangles) only the TDS content of the synthetic produced water, and (green
 327 diamonds) the resulting feed after the thermal Fenton oxidation. (b) TOC concentration in the
 328 distillate tank (initial volume 1 L) as a function of cumulative permeated volume; the lines
 329 connecting the data points are only intended as guides for the eye.

330

331 The only organic compounds in the synthetic PW which may induce wetting were humic
 332 acids (HA) and the water-miscible compounds (WMC), namely, BTX, MTBE, phenol, and
 333 cyclohexane. HA are amphiphilic and may partly act as surfactants; however, not presenting

334 a clear separation between the hydrophobic and hydrophilic portion of the molecule, they can
335 virtually maintain a repulsive behavior for liquid water after the interaction with the
336 hydrophobic membrane, thus resulting in fouling, but not necessarily in wetting (Klavins and
337 Purmalis 2010; Wang et al. 2018; Horseman et al. 2021). On the other hand, WMC may
338 induce wetting by lowering the ST of the solution. However, since no wetting from organic
339 compounds was detected in this study, it is reasonable to assume that the ST threshold needed
340 to observe wetting under the condition of this study was lower than the ST value of the
341 synthetic PW (50.8 ± 2.7 dyn/cm). The fact that organic fouling did not cause wetting may be
342 rationalized with the short duration of the lab experiments, which were run for approximately
343 8 h before the observed drop in water flux due to scaling. The slow kinetics of fouling
344 phenomena in such a system may require longer filtration times to show wetting effects and
345 should become important at real scale during operation. Note that paraffins form a different
346 phase, hence they cannot lower the ST of the solution or create water bridges within the
347 membrane pores, but only freely pass through the PTFE membrane by virtue of their
348 hydrophobic nature.

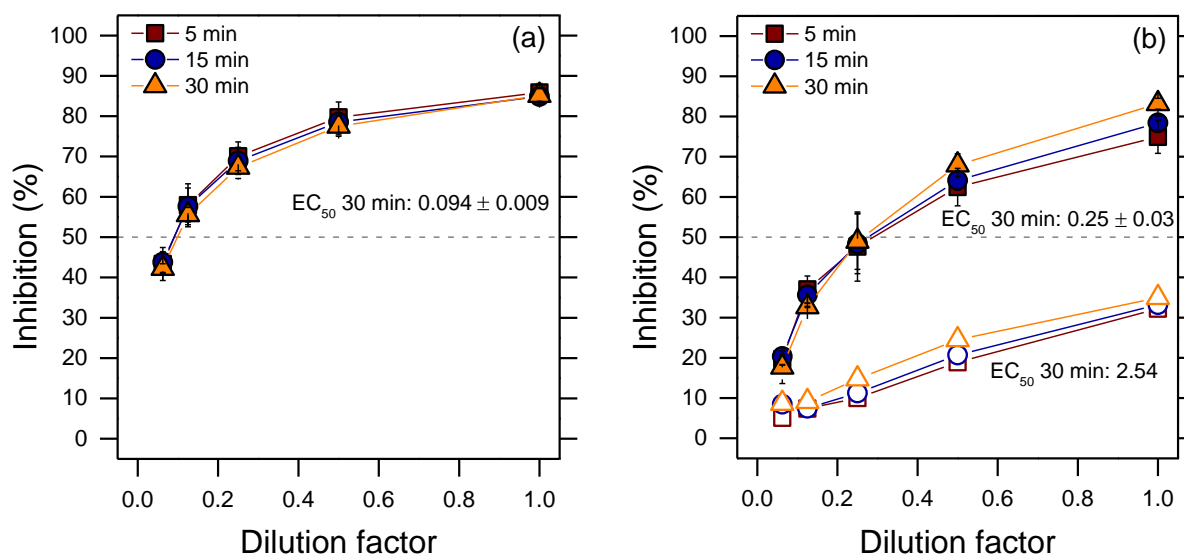
349 While the Fenton pre-treatment did not provide specific advantages in terms of
350 productivity, which was governed by salt concentration and partly affected by fouling, or of
351 prevention of wetting, which was not observed even for the use of untreated PW as feed
352 stream, oxidation had beneficial effect in terms of MD effluent quality and toxicity. The
353 results in **Figure 2b** suggest a clear reduction of the TOC in the distillate when desalinating
354 the feed matrix subject to Fenton reaction; see Figure S5 of the SM for the chromatograms of
355 the final effluent treated only in MD and by the coupled Fenton-MD system. As expected, the
356 MD process did not separate water from volatile WMC or non-aqueous oils. The first data
357 points in **Figure 2b** showed a high TOC passage, justifiable considering an instantaneous
358 passage of a fraction of the organic content, specifically, paraffins and WMC. Subsequently,

359 the TOC in the distillate tank decreased steadily by dilution with the nearly pure water vapor
360 permeating the membrane. The lower amount of TOC measured upon oxidation of the feed
361 stream with Fenton is imputable to both the mineralization of a fraction of toxic compounds
362 (~52% of mineralization, **Table 3**) and the transformation of organic substances to more
363 hydrophilic compounds, which are less prone to pass through the hydrophobic membrane.
364 Fenton oxidation may also provide beneficial effects in terms of MD performance at real
365 scale by thwarting fouling phenomena that would occur at longer time scales, but this effect
366 could not be observed in this study.

367 3.3. *Effect of the coupled system on the toxicity of the final effluent*

368 A general index of the safety of an effluent is its toxicity. Toxicity is also a legislated
369 parameter, allowing or denying the discharge of an effluent in the sewage system. According
370 to the Italian regulations (D.Lgs. 152/2006), the acute toxicity limit to discharge an effluent
371 in the sewage system is 80% of the inhibition of the target microorganism (in this case,
372 *Vibrio fischeri*). The synthetic PW of this study presented an acute toxicity around 100%
373 (Figure S6a of the SM), a value also expected for most of the real PW due to the wide variety
374 and large concentrations of contaminants typically present. Both organics and concentrated
375 salts may present large toxic effects, hence they both require a specific treatment. **Figure 3**
376 presents the residual acute toxicity of the effluent treated with only MD and with the coupled
377 Fenton-MD system expressed in term of dilution factor of the original samples. The EC_{50}
378 value was used as a comparative parameter of the quality of different effluents, and it is
379 defined as the half maximal effective concentration, namely, the concentration required to
380 obtain 50% of microorganism inhibition. This parameter increased significantly (from a
381 dilution factor of 0.094 ± 0.009 to 0.25 ± 0.03) when the effluent was oxidized and then
382 desalinated, compared to a stream that was not pre-treated, i.e., the coupled system gave

383 lower toxicity. However, when tested as is, the residual toxicity of the effluent was still high
 384 and over the limit of 80% after 30 min of contact time with the bacteria.



385 **Figure 3.** Residual toxicity of the effluent expresses in term of dilution factor of the original
 386 sample (a) after MD treatment only, and (b) upon treatment by the coupled Fenton-MD
 387 system (solid data points) with and (empty data points) without Fe sequestration. The toxicity
 388 was measured after 5, 15, and 30 minutes of contact (in red, blue, and orange, respectively)
 389 with the *Vibrio fischeri* culture. The dash line indicates the point where 50% of acute toxicity
 390 is reached.

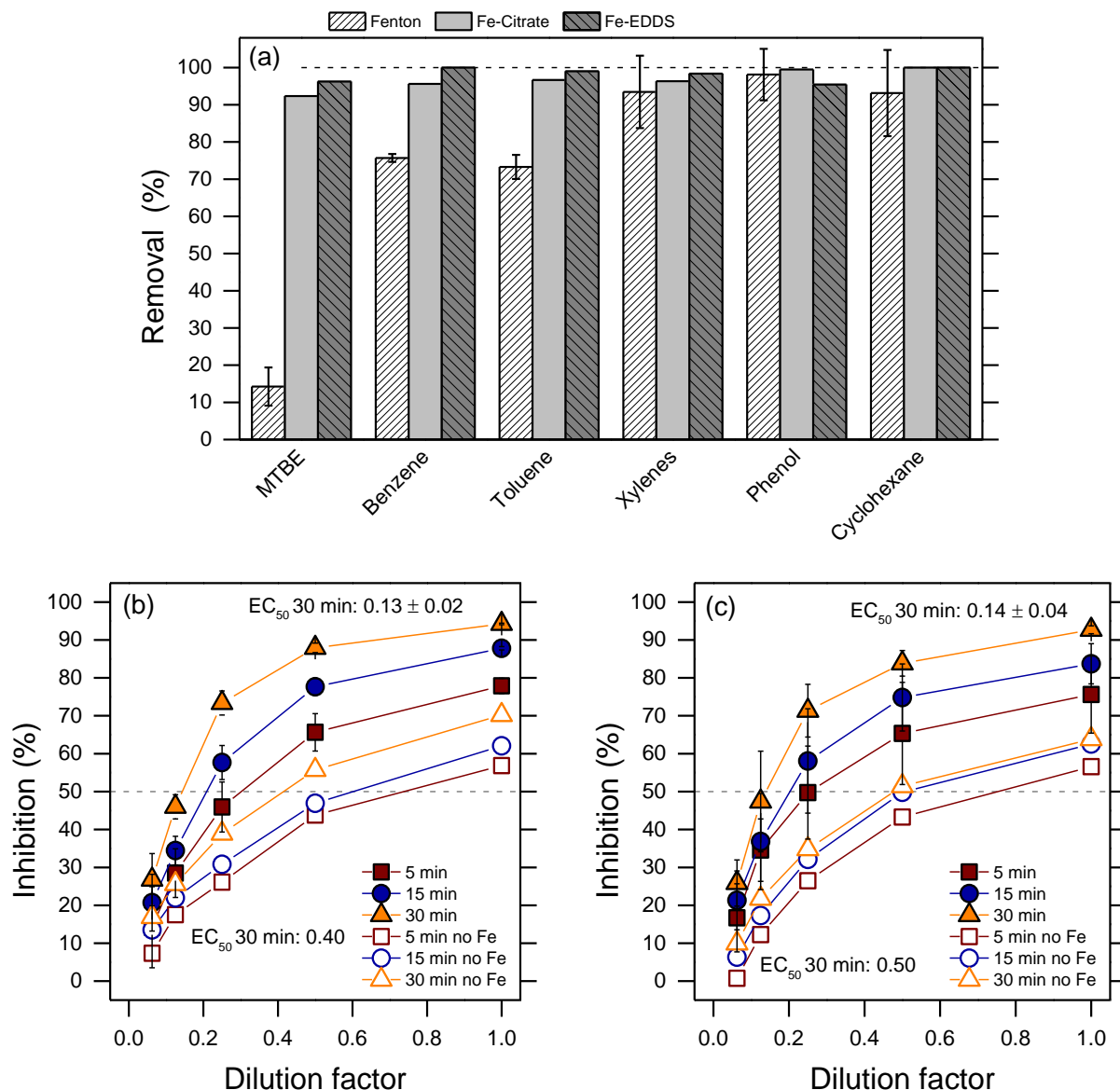
391

392 To understand the relative contribution to toxicity of the different contaminants, toxicity
 393 experiments were carried out with individual components (Figures S6 and S7 of the SM).
 394 Low toxicity was associated with HA and paraffins, while higher toxic effects were related to
 395 parent WMC, which however were degraded effectively by the Fenton oxidation. Calcium
 396 and magnesium chloride also showed negligible toxicity, while sodium chloride showed
 397 some toxicity only at concentrations >50 g/L, that is, far above the concentration measured
 398 after the MD desalination. On the other hand, iron (II) sulfate showed a significant toxicity:
 399 residual iron may thus be responsible for the toxicity observed for the effluent from Fenton

400 and MD treatments. Indeed, by addition of EDDS in molar ratio 1:1 to the residual iron (0.26
401 mM), the acute toxicity of iron was shut down (see Figure S2 and S7 in SM for the toxicity of
402 Fe-EDDS and iron, respectively) and that of the effluent markedly decreased (from ~80% to
403 ~40%), accompanied with a substantial increase in EC₅₀ (from a dilution factor of 0.25 ± 0.03
404 to ~2.5). The remaining toxicity after this post-treatment step may be reasonably attributed to
405 the residual WMC and to the by-products of Fenton oxidation. Note that in this study EDDS
406 was added to mask the toxic behavior of the residual iron, which confirms the fact that
407 residual metal concentration causes toxicity in the effluent. However, different strategies to
408 remove the residual iron may be implemented in real plants, for example, its precipitation
409 under basic pH. Cleaner and more novel steps may involve enhanced ion exchange resins and
410 the use of adsorbents, such as magnetic nanoparticles (Khatri et al. 2017).

411 3.4. *Comparison between traditional Fenton and Fenton process in the presence of iron* 412 *ligands*

413 Iron ligands keep iron in solution without the need for the pH adjustment to 3 and limit
414 the production of sludge, thus potentially allowing for a more streamlined operation
415 compared to the traditional Fenton process. Citrate and EDDS were chosen as non-toxic
416 (Figure S2 of the SM) ligands and the Fe-citrate and Fe-EDDS systems were applied as
417 catalysts (Zhang et al. 2016). The optimized condition in terms of Fe-ligand to H₂O₂ molar
418 ratio corresponded to the dosages relative to Ox7 in **Table 2**; each ligand was dosed
419 equimolarly with iron. The results summarized in **Figure 4a** suggest higher degradation
420 efficiency of the modified Fenton systems toward WMC with respect to the traditional
421 thermal Fenton. Specifically, the modified Fenton processes achieved near complete
422 degradation of all WMC, including MTBE that was instead not removed by the classic
423 Fenton reaction; see also the chromatograms and percentage of substrates removal at different
424 oxidation times in Figures S8, S9, and S10 (SM).



426 **Figure 4** (a) Comparison of the organics removal rate by traditional Fenton and modified
 427 Fenton reactions with citrate and EDDS as iron ligands. Residual toxicity of the final effluent
 428 treated by the coupled system comprising modified-Fenton and MD using (b) Fe-citrate and
 429 (c) Fe-EDDS as catalyst. The open symbols are related to the effluent upon iron sequestration
 430 by EDDS. The toxicity was measured after 5, 15, and 30 minutes of contact with the *Vibrio*
 431 *fischeri* culture. The dash horizontal line indicates the point where 50% of acute toxicity is
 432 reached; the lines connecting the data points are only intended as guides for the eye.

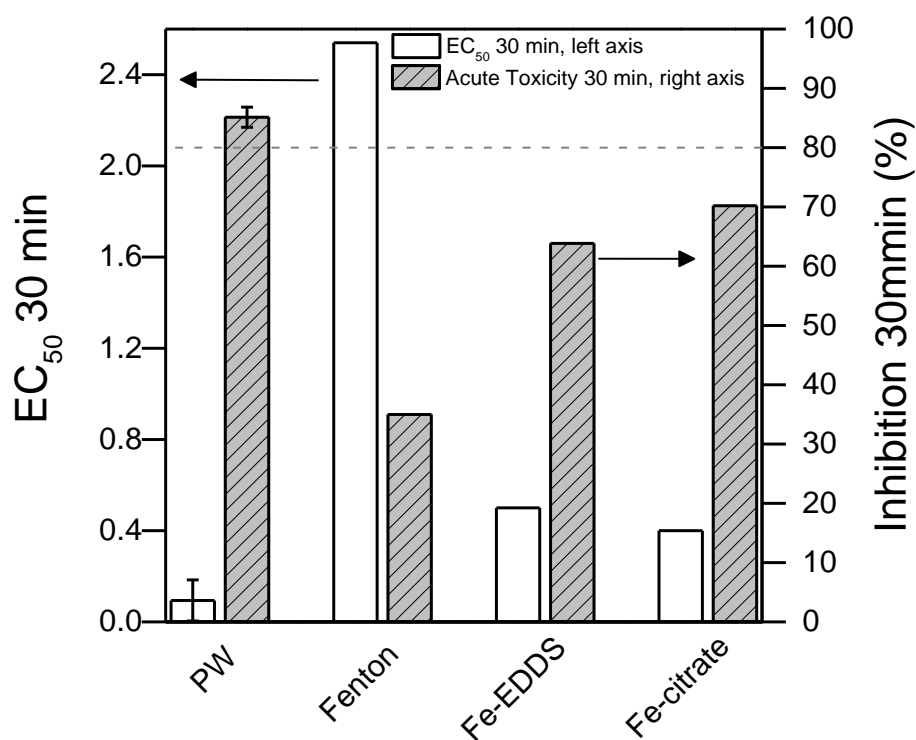
434 Despite the higher percentage of degradation of the target contaminants, the modified
435 Fenton process did not show further beneficial effect on the MD process in terms of
436 productivity and effluent quality compared to traditional Fenton; see Figure S11 of the SM.
437 With regard to EC_{50} , the quality of the effluent treated with the coupled system including
438 modified Fenton and MD clearly increased compared to the PW treated only with MD,
439 specifically, from a dilution factor of 0.094 ± 0.009 to 0.13 ± 0.02 and to 0.14 ± 0.04 in the
440 case of Fe-citrate and Fe-EDDS, respectively (see **Figure 4b and c**). However, the EC_{50} was
441 lower compared with the residual toxicity observed in the system that included the traditional
442 Fenton oxidation (0.25 ± 0.03), possibly due to the formation of more toxic by-products
443 (Zhang et al. 2016). When the contribution to toxicity of the residual iron (residual iron was
444 0.27 and 1.30 mM for Fe-citrate and Fe-EDDS, respectively) was prevented by EDDS
445 addition, the acute toxicity of the effluent clearly decreased (from ~90% to ~70% in both
446 cases) and the EC_{50} further increased (to 0.4 and 0.5 for Fe-citrate and Fe-EDDS,
447 respectively). At the end of the process, the effluent treated with the Fe-EDDS-based Fenton
448 system presented a slightly better quality in terms of toxicity than that treated with the Fe-
449 citrate system, consistently with the slightly higher degradation efficiency of the former
450 compared to the latter.

451

452 **4. Concluding remarks, challenges, and implications**

453 This work evaluated the impact of traditional and modified Fenton oxidations for the
454 abatement of highly toxic organic contaminants and as pre-treatment options for the
455 subsequent desalination of hypersaline produced waters by membrane distillation. Fe-citrate
456 and Fe-EDDS were used as inexpensive, easy to handle, environmentally friendly and
457 biodegradable systems in the modified Fenton processes. All the oxidative processes

458 provided relatively high degradation efficiency toward target contaminants also in presence
 459 of typical scavengers of the Fenton reaction, namely, chloride and humic acids. The
 460 observable beneficial effects of an oxidative pre-treatment were not evident in terms of MD
 461 productivity, fouling, or wetting, but directly translated into lower permeation of organics
 462 during distillation and in a significantly lower toxicity of the desalinated effluent.
 463 Specifically, the EC₅₀ and the acute toxicity (inhibition % of target organisms) were used as
 464 indexes for the evaluation of the quality of the final effluent. The target values of acute
 465 toxicity should be below the regulated limit of 80%. **Figure 5** offers a final evaluation of the
 466 toxicity parameters from the various treatments. The traditional Fenton coupled with the MD
 467 desalination was the best in terms of toxicity. However, the modified Fenton-MD coupled
 468 systems were able to overcome some of the practical limitations of the traditional Fenton
 469 while still providing an effluent with suitable quality for safe discharge as sewage (toxicity
 470 <80%). The modified Fenton oxidation may be more advisable for applications whereby
 471 easier operational tasks and lower sludge production are important, such as offshore.



472

473 **Figure 5** Summary of the residual toxicity and EC₅₀ values of the various effluents treated
474 with both MD and oxidation processes after 30 minutes of contact with the *Vibrio fischeri*
475 culture. The dash line is relative to the value of 80% of acute toxicity, namely, the regulated
476 legislative limit for a safe discharge in the sewage system in Italy.

477

478 In conclusion, the coupled oxidation-MD systems to treat PW allow a less toxic effluent
479 compared to the initial PW toxicity. All the final effluents obtained in this study may be
480 safely discharged in the sewage system and treated within the civil wastewater treatment
481 trains, according to the Italian legislation. The oxidation processes are promising for PW
482 treatments since they are able to degrade the toxic initial target contaminants almost
483 completely. Moreover, the modified Fenton process is able to effectively treat PW while
484 overcoming the practical limitations of traditional Fenton (sludge production and acidic pH).
485 The Fenton processes add iron in the effluent environment, this metal being associated with
486 intrinsic toxicity; thus, the toxicity associated with the residual iron needs to be properly
487 addressed and managed in real plants. A wide range of options are available to remove iron
488 and the best-fitting one should be selected case-by-case. Moreover, the oxidations struggle in
489 achieving the complete mineralization of all the organics in PW, thus some potentially toxic
490 by-products may be formed. Therefore, an accurate monitoring of the by-products may be
491 necessary, possibly also enforcing some control on the reaction pathway, to evaluate in each
492 case the possibility to safely discharge the final effluent to a civil wastewater treatment plant
493 and in case to specifically target the most troublesome by-products in a tertiary treatment
494 step.

495

496 **CRedit authorship contribution statement**

497 **Giulio Farinelli:** Conceptualization, Data curation, Formal analysis, Investigation,
498 Methodology, Visualization, Writing - original draft. **Marco Coha:** Data Curation, Formal
499 analysis, Investigation, Methodology, Visualization, Writing - review & editing. **Marco**
500 **Minella:** Data curation, Validation, Methodology, Writing - review & editing. **Debora**
501 **Fabbri:** Resources, Supervision, Writing - review & editing. **Marco Pazzi:** Formal analysis,
502 Methodology. **Davide Vione:** Resources, Supervision, Writing - review & editing. **Alberto**
503 **Tiraferri:** Funding acquisition, Project administration, Resources, Supervision,
504 Visualization, Writing - review & editing.

505

506 **Declaration of Competing Interest**

507 The authors declare no competing financial interest.

508

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512

513 **Appendix A. Supplementary data.**

514 Supplementary material related to this article can be found, in the online version.

515

516

517 **References**

- 518 Adewumi, M. A., Erb, J. E., Watson, R. W. (1992). Initial Design Considerations for a Cost-Effective
519 Treatment of Stripper Oil-Well Produced Water. *Produced Water*. R. J. P. and E. F. R.
520 Botsom, MA, Springer. **46**: 511-522.
- 521 Ahmadun, F. R., Pendashteh, A., Abdullah, L. C., Biak, D. R. A., Madaeni, S. S., Abidin, Z. Z., 2009.
522 Review of technologies for oil and gas produced water treatment. *J. Hazard. Mater.* 170, 530-
523 551. <https://doi.org/10.1016/j.jhazmat.2009.05.044>
- 524 Al-Ghouti, M. A., Al-Kaabi, M. A., Ashfaq, M. Y., Da'na, D. A., 2019. Produced water
525 characteristics, treatment and reuse: A review. *J. Water Process Eng.* 28, 222-239.
526 <https://doi.org/10.1016/j.jwpe.2019.02.001>
- 527 APAT, I., CNR (2003). Metodi Ecotossicologici. *Manuali e linee guida 29/2003 - Serie 8000*.
- 528 Aquilina, P. (2012). Impairment of Gas Well Productivity by Salt Plugging: A Review of
529 Mechanisms, Modeling, Monitoring Methods, and Remediation Techniques. *SPE Annual*
530 *Technical Conference and Exhibition*. San Antonio, Texas, USA.
- 531 Ariono, D., Purwasmita, M., Wenten, I. G., 2016. Brine Effluents: Characteristics, Environmental
532 Impacts, and Their Handling. *J. Eng. Technol. Sci.* 48, 367-387.
533 <https://doi.org/10.5614/j.eng.technol.sci.2016.48.4.1>
- 534 Ayed, L., Asses, N., Chammem, N., Ben Othman, N., Hamdi, M., 2017. Advanced oxidation process
535 and biological treatments for table olive processing wastewaters: constraints and a novel
536 approach to integrated recycling process: a review. *Biodegradation.* 28, 125-138.
537 <https://doi.org/10.1007/s10532-017-9782-0>
- 538 Bessa, E., Sant Anna, G. L., Dezotti, M., 2001. Photocatalytic/H₂O₂ treatment of oil field produced
539 waters. *Appl. Catal. B-Environ.* 29, 125-134. [https://doi.org/10.1016/S0926-3373\(00\)00199-5](https://doi.org/10.1016/S0926-3373(00)00199-5)
- 540 Canedo-Arguelles, M., Kefford, B., Schafer, R., 2019. Salt in freshwaters: causes, effects and
541 prospects - introduction to the theme issue. *Philos. T. R. Soc. B.* 374,
542 <https://doi.org/10.1098/Rstb.2018.0002>
- 543 Chahbane, N., Popescu, D. L., Mitchell, D. A., Chanda, A., Lenoir, D., Ryabov, A. D., Schramm, K.
544 W., Collins, T. J., 2007. Fe-III-TAML-catalyzed green oxidative degradation of the azo dye
545 Orange II by H₂O₂ and organic peroxides: products, toxicity, kinetics, and mechanisms.
546 *Green Chem.* 9, 49-57. <https://doi.org/10.1039/b604990g>
- 547 Chang, H. Q., Li, T., Liu, B. C., Vidic, R. D., Elimelech, M., Crittenden, J. C., 2019. Potential and
548 implemented membrane-based technologies for the treatment and reuse of flowback and
549 produced water from shale gas and oil plays: A review. *Desalination.* 455, 34-57.
550 <https://doi.org/10.1016/j.desal.2019.01.001>
- 551 Chang, H. Q., Liu, B. C., Wang, H. Z., Zhang, S. Y., Chen, S., Tiraferri, A., Tang, Y. Q., 2019.
552 Evaluating the performance of gravity-driven membrane filtration as desalination
553 pretreatment of shale gas flowback and produced water. *J. Membr. Sci.* 587,
554 <https://doi.org/10.1016/J.Memsci.2019.117187>
- 555 Chen, L., Wang, D., Long, C., Cui, Z. X., 2019. Effect of biodegradable chelators on induced
556 phytoextraction of uranium- and cadmium-contaminated soil by *Zebrina pendula* Schnizl. *Sci.*
557 *Rep.* 9, <https://doi.org/10.1038/S41598-019-56262-9>
- 558 Chen, W., Zhang, F. F., Hong, J. L., Shi, W. X., Feng, S. T., Tan, X. F., Geng, Y., 2016. Life cycle
559 toxicity assessment on deep-brine well drilling. *J. Clean. Prod.* 112, 326-332.
560 <https://doi.org/10.1016/j.jclepro.2015.07.062>
- 561 Chen, Y. M. L., Wang, Z. X., Jennings, G. K., Lin, S. H., 2017. Probing Pore Wetting in Membrane
562 Distillation Using Impedance: Early Detection and Mechanism of Surfactant-Induced
563 Wetting. *Environ. Sci. Tech. Lett.* 4, 505-510. <https://doi.org/10.1021/acs.estlett.7b00372>

564 Coha, M., Farinelli, G., Tiraferri, A., Minella, M., Vione, D., 2021. Advanced oxidation processes in
565 the removal of organic substances from produced water: Potential, configurations, and
566 research needs. *Chem. Eng. J.* 414, <https://doi.org/10.1016/j.cej.2021.128668>

567 Dalmacija, B., Karlovic, E., Tamas, Z., Miskovic, D., 1996. Purification of high-salinity wastewater
568 by activated sludge process. *Water Res.* 30, 295-298. [https://doi.org/10.1016/0043-
569 1354\(95\)00170-0](https://doi.org/10.1016/0043-1354(95)00170-0)

570 De Laat, J., Gallard, H., 1999. Catalytic decomposition of hydrogen peroxide by Fe(III) in
571 homogeneous aqueous solution: Mechanism and kinetic modeling. *Environ. Sci. Technol.* 33,
572 2726-2732. <https://doi.org/10.1021/Es981171v>

573 Diya'uddeen, B. H., Aziz, A. R. A., Daud, W. M. A. W., 2012. On the Limitation of Fenton Oxidation
574 Operational Parameters: A Review. *Int. J. Chem. React. Eng.* 10,
575 <https://doi.org/10.1515/1542-6580.R2>

576 Donaldson, E. C., Thomas, R. D., Lorenz, P. B., 1969. Wettability Determination and Its Effect on
577 Recovery Efficiency. *Soc. Pet. Eng. J.* 9, 13-20. <https://doi.org/10.2118/2338-PA>

578 Estrada, J. M., Bhamidimarri, R., 2016. A review of the issues and treatment options for wastewater
579 from shale gas extraction by hydraulic fracturing. *Fuel.* 182, 292-303.
580 <https://doi.org/10.1016/j.fuel.2016.05.051>

581 Farinelli, G., Minella, M., Pazzi, M., Giannakis, S., Pulgarin, C., Vione, D., Tiraferri, A., 2020.
582 Natural iron ligands promote a metal-based oxidation mechanism for the Fenton reaction in
583 water environments. *J. Hazard. Mater.* 393, <https://doi.org/10.1016/j.jhazmat.2020.122413>

584 Farinelli, G., Minella, M., Sordello, F., Vione, D., Tiraferri, A., 2019. Metabisulfite as an
585 Unconventional Reagent for Green Oxidation of Emerging Contaminants Using an Iron-
586 Based Catalyst. *ACS Omega.* 4, 20732-20741. <https://doi.org/10.1021/acsomega.9b03088>

587 Franken, A. C. M., Nolten, J. A. M., Mulder, M. H. V., Bargeman, D., Smolders, C. A., 1987. Wetting
588 Criteria for the Applicability of Membrane Distillation. *J. Membr. Sci.* 33, 315-328.
589 [https://doi.org/10.1016/S0376-7388\(00\)80288-4](https://doi.org/10.1016/S0376-7388(00)80288-4)

590 Goldstone, J. V., Pullin, M. J., Bertilsson, S., Voelker, B. M., 2002. Reactions of hydroxyl radical
591 with humic substances: Bleaching, mineralization, and production of bioavailable carbon
592 substrates. *Environ. Sci. Technol.* 36, 364-372. <https://doi.org/10.1021/es0109646>

593 Goncalves, N. P. F., Minella, M., Fabbri, D., Calza, P., Malitesta, C., Mazzotta, E., Prevot, A. B.,
594 2020. Humic acid coated magnetic particles as highly efficient heterogeneous photo-Fenton
595 materials for wastewater treatments. *Chem. Eng. J.* 390,
596 <https://doi.org/10.1016/J.Cej.2020.124619>

597 Gonzalez, D., Amigo, J., Suarez, F., 2017. Membrane distillation: Perspectives for sustainable and
598 improved desalination. *Renew. Sust. Energ. Rev.* 80, 238-259.
599 <https://doi.org/10.1016/j.rser.2017.05.078>

600 Haber, F., Weiss, J., Pope, W. J., 1934. The catalytic decomposition of hydrogen peroxide by iron
601 salts. *Proc. R. Soc. Lond. A - Math. Phys. Sci.* 147, 332-351.
602 <https://doi.org/10.1098/rspa.1934.0221>

603 Han, L., Tan, Y. Z., Netke, T., Fane, A. G., Chew, J. W., 2017. Understanding oily wastewater
604 treatment via membrane distillation. *J. Membr. Sci.* 539, 284-294.
605 <https://doi.org/10.1016/j.memsci.2017.06.012>

606 Harvey, A. E., Smart, J. A., Amis, E. S., 1955. Simultaneous Spectrophotometric Determination of
607 Iron(II) and Total Iron with 1,10-Phenanthroline. *Anal. Chem.* 27, 26-29.
608 <https://doi.org/10.1021/ac60097a009>

609 Horseman, T., Yin, Y., Christie, K. S., Wang, Z., Tong, T., Lin, S., 2021. Wetting, Scaling, and
610 Fouling in Membrane Distillation: State-of-the-Art Insights on Fundamental Mechanisms and
611 Mitigation Strategies. *ACS ES&T Eng.* 1, 117-140.
612 <https://doi.org/10.1021/acsestengg.0c00025>

613 Howell, J. A., 2004. Future of membranes and membrane reactors in green technologies and for water
614 reuse. *Desalination.* 162, 1-11. [https://doi.org/10.1016/S0011-9164\(04\)00021-9](https://doi.org/10.1016/S0011-9164(04)00021-9)

615 Igunnu, E. T., Chen, G. Z., 2014. Produced water treatment technologies. *Int. J. Low-Carbon Technol.*
616 9, 157-177. <https://doi.org/10.1093/ijlct/cts049>

617 Jimenez, S., Mico, M. M., Arnaldos, M., Medina, F., Contreras, S., 2018. State of the art of produced
618 water treatment. *Chemosphere.* 192, 186-208.
619 <https://doi.org/10.1016/j.chemosphere.2017.10.139>

620 Kabyal, A., Yang, M., Abbassi, R., Li, S. H., 2020. A risk -based approach to produced water
621 management in offshore oil and gas operations. *Process Saf. Environ.* 139, 341-361.
622 <https://doi.org/10.1016/j.psep.2020.04.021>

623 Kargbo, D. M., Wilhelm, R. G., Campbell, D. J., 2010. Natural Gas Plays in the Marcellus Shale:
624 Challenges and Potential Opportunities. *Environ. Sci. Technol.* 44, 5679-5684.
625 <https://doi.org/10.1021/es903811p>

626 Khatri, N., Tyagi, S., Rawtani, D., 2017. Recent strategies for the removal of iron from water: A
627 review. *J. Water. Process Eng.* 19, 291-304. <https://doi.org/10.1016/j.jwpe.2017.08.015>

628 Kiwi, J., Lopez, A., Nadtochenko, V., 2000. Mechanism and kinetics of the OH-radical intervention
629 during fenton oxidation in the presence of a significant amount of radical scavenger (Cl-).
630 *Environ. Sci. Technol.* 34, 2162-2168. <https://doi.org/10.1021/Es991406i>

631 Klavins, M., Purmalis, O., 2010. Humic substances as surfactants. *Environ. Chem. Lett.* 8, 349-354.
632 <https://doi.org/10.1007/s10311-009-0232-z>

633 Kleinitz, W., Koehler, M., Dietzsch, G. (2001). The Precipitation of Salt in Gas Producing Wells. *SPE*
634 *European Formation Damage Conference*. The Hague, Netherlands.

635 Lester, Y., Ferrer, I., Thurman, E. M., Sitterley, K. A., Korak, J. A., Aiken, G., Linden, K. G., 2015.
636 Characterization of hydraulic fracturing flowback water in Colorado: Implications for water
637 treatment. *Sci. Total Environ.* 512, 637-644. <https://doi.org/10.1016/j.scitotenv.2015.01.043>

638 Lin, S. H., Nejati, S., Boo, C., Hu, Y. X., Osuji, C. O., Ehmelech, M., 2014. Omniphobic Membrane
639 for Robust Membrane Distillation. *Environ. Sci. Tech. Let.* 1, 443-447.
640 <https://doi.org/10.1021/ez500267p>

641 Liu, Y., Lu, H., Li, Y., Xu, H., Pan, Z., Dai, P., Wang, H., Yang, Q., 2021. A review of treatment
642 technologies for produced water in offshore oil and gas fields. *Sci. Total Environ.* 775,
643 <https://doi.org/10.1016/j.scitotenv.2021.145485>

644 Ma, H. Z., Wang, B., 2006. Electrochemical pilot-scale plant for oil field produced wastewater by
645 M/C/Fe electrodes for injection. *J. Hazard. Mater.* 132, 237-243.
646 <https://doi.org/10.1016/j.jhazmat.2005.09.043>

647 McCormack, P., Jones, P., Hetheridge, M. J., Rowland, S. J., 2001. Analysis of oilfield produced
648 waters and production chemicals by electrospray ionisation multi-stage mass spectrometry
649 (ESI-MSn). *Water Res.* 35, 3567-3578. [https://doi.org/10.1016/S0043-1354\(01\)00070-7](https://doi.org/10.1016/S0043-1354(01)00070-7)

650 Messele, S. A., Bengoa, C., Stuber, F. E., Giralt, J., Fortuny, A., Fabregat, A., Font, J., 2019.
651 Enhanced Degradation of Phenol by a Fenton-Like System (Fe/EDTA/H2O2) at
652 Circumneutral pH. *Catalysts.* 9, <https://doi.org/10.3390/Catal9050474>

653 Miklos, D. B., Remy, C., Jekel, M., Linden, K. G., Drewes, J. E., Hubner, U., 2018. Evaluation of
654 advanced oxidation processes for water and wastewater treatment - A critical review. *Water*
655 *Res.* 139, 118-131. <https://doi.org/10.1016/j.watres.2018.03.042>

656 Mohammad-Pajoo, E., Weichgrebe, D., Cuff, G., Tosarkani, B. M., Rosenwinkel, K. H., 2018. On-
657 site treatment of flowback and produced water from shale gas hydraulic fracturing: A review
658 and economic evaluation. *Chemosphere.* 212, 898-914.
659 <https://doi.org/10.1016/j.chemosphere.2018.08.145>

660 Neff, J. M., Sauer, T. C., Maciolek, N. (1992). Composition, Fate and Effects of Produced Water
661 Discharges to Nearshore Marine Waters. *Produced Water*. R. J. P. and E. F. R. Boston, MA,
662 Springer. 46: 371-385.

663 Olsson, O., Weichgrebe, D., Rosenwinkel, K. H., 2013. Hydraulic fracturing wastewater in Germany:
664 composition, treatment, concerns. *Environ. Earth Sci.* 70, 3895-3906.
665 <https://doi.org/10.1007/s12665-013-2535-4>

- 666 Pasternak, G., Kolwzan, B., 2013. Surface tension and toxicity changes during biodegradation of
667 carbazole by newly isolated methylotrophic strain *Methylobacterium* sp GPE1. *Int. Biodeter.*
668 *Biodegr.* 84, 143-149. <https://doi.org/10.1016/j.ibiod.2012.07.021>
- 669 Rezaei, M., Warsinger, D. M., Lienhard, V. J. H., Samhaber, W. M., 2017. Wetting prevention in
670 membrane distillation through superhydrophobicity and recharging an air layer on the
671 membrane surface. *J. Membr. Sci.* 530, 42-52. <https://doi.org/10.1016/j.memsci.2017.02.013>
- 672 Ricceri, F., Giagnorio, M., Farinelli, G., Blandini, G., Minella, M., Vione, D., Tiraferri, A., 2019.
673 Desalination of Produced Water by Membrane Distillation: Effect of the Feed Components
674 and of a Pre-treatment by Fenton Oxidation. *Sci. Rep.* 9, 14964.
675 <https://doi.org/10.1038/s41598-019-51167-z>
- 676 Shaffer, D. L., Chavez, L. H. A., Ben-Sasson, M., Castrillon, S. R. V., Yip, N. Y., Elimelech, M.,
677 2013. Desalination and Reuse of High-Salinity Shale Gas Produced Water: Drivers,
678 Technologies, and Future Directions. *Environ. Sci. Technol.* 47, 9569-9583.
679 <https://doi.org/10.1021/es401966e>
- 680 Shang, W., Tiraferri, A., He, Q. P., Li, N. W., Chang, H. Q., Liu, C., Liu, B. C., 2019. Reuse of shale
681 gas flowback and produced water: Effects of coagulation and adsorption on ultrafiltration,
682 reverse osmosis combined process. *Sci. Total Environ.* 689, 47-56.
683 <https://doi.org/10.1016/j.scitotenv.2019.06.365>
- 684 Shokrollahzadeh, S., Golmohammad, F., Naseri, N., Shokouhi, H., Arman-mehr, M., 2012. Chemical
685 oxidation for removal of hydrocarbons from gas-field produced water. *Procedia Engineer.* 42,
686 942-947. <https://doi.org/10.1016/j.proeng.2012.07.487>
- 687 Tandy, S., Ammann, A., Schulin, R., Nowack, B., 2006. Biodegradation and speciation of residual
688 SS-ethylenediaminedisuccinic acid (EDDS) in soil solution left after soil washing. *Environ.*
689 *Pollut.* 142, 191-199. <https://doi.org/10.1016/j.envpol.2005.10.013>
- 690 Tang, P., Li, J. L., Li, T., Tian, L., Sun, Y., Xie, W. C., He, Q. P., Chang, H. Q., Tiraferri, A., Liu, B.
691 C., 2021. Efficient integrated module of gravity driven membrane filtration, solar aeration and
692 GAC adsorption for pretreatment of shale gas wastewater. *J. Hazard. Mater.* 405,
693 <https://doi.org/10.1016/j.jhazmat.2020.124166>
- 694 Van Devivere, P. C., Saveyn, H., Verstraete, W., Feijtel, T. C. J., Schowanek, D. R., 2001.
695 Biodegradation of metal-[S,S]-EDDS complexes. *Environ. Sci. Technol.* 35, 1765-1770.
696 <https://doi.org/10.1021/es0001153>
- 697 Vesterkvist, P. S. M., Misiorek, J. O., Spooft, L. E. M., Toivola, D. M., Meriluoto, J. A. O., 2012.
698 Comparative Cellular Toxicity of Hydrophilic and Hydrophobic Microcystins on Caco-2
699 Cells. *Toxins.* 4, 1008-1023. <https://doi.org/10.3390/toxins4111008>
- 700 Vione, D., Merlo, F., Maurino, V., Minero, C., 2004. Effect of humic acids on the Fenton degradation
701 of phenol. *Environ. Chem. Lett.* 2, 129-133. <https://doi.org/10.1007/s10311-004-0086-3>
- 702 Voelker, B. M., Sulzberger, B., 1996. Effects of fulvic acid on Fe(II) oxidation by hydrogen peroxide.
703 *Environ. Sci. Technol.* 30, 1106-1114. <https://doi.org/10.1021/Es9502132>
- 704 Wang, Z. X., Chen, Y. M. L., Sun, X. M., Duddu, R., Lin, S. H., 2018. Mechanism of pore wetting in
705 membrane distillation with alcohol vs. surfactant. *J Membrane Sci.* 559, 183-195.
706 <https://doi.org/10.1016/j.memsci.2018.04.045>
- 707 Zhang, Y., Klammerth, N., Chelme-Ayala, P., Gamal El-Din, M., 2016. Comparison of Nitrilotriacetic
708 Acid and [S,S]-Ethylenediamine-N,N'-disuccinic Acid in UV-Fenton for the Treatment of Oil
709 Sands Process-Affected Water at Natural pH. *Environ. Sci. Technol.* 50, 10535-10544.
710 <https://doi.org/10.1021/acs.est.6b03050>

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