

Investigation of pre-treatments improving low-temperature anaerobic digestion of waste activated sludge

*Original*

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### Abstract

This work analyzed the feasibility of pre-treatments to improve the anaerobic digestion (AD) of waste activated sludge (WAS) at 20 °C. We investigated different physicochemical pre-treatments (thermal at 115 °C, thermo-alkaline at pH 10 and 70 °C and ozonation at 190 mg-O<sub>3</sub> L<sup>-1</sup>) by comparing their performances about COD solubilization and sludge disintegration rate. Best performances were obtained by thermo-alkaline pre-treatment, followed by thermal and ozonation; results were consistent with literature. Pre-treated WAS was fed to 12 1-L anaerobic digesters operated in semi-continuous mode. Thermal and thermo-alkaline reactors produced biogas yields (0.30-0.36 m<sup>3</sup> kg<sup>-1</sup> VS in standard conditions, 65-70 % methane) analogous to mesophilic conditions. The economic assessment of the scale-up of the whole process demonstrated that thermo-alkaline pre-treatment made AD at 20 °C economically profitable for WAS generated by a 20,000 PE WWTP.

**Keywords** anaerobic digestion; biogas; low-temperature; pre-treatment; semi-continuous; waste activated sludge

**Taxonomy** Water Treatment, Treatment, Sustainable Economy, Energy Engineering

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**Corresponding Author** Silvia Fiore

**Corresponding Author's Institution** Politecnico di Torino

**Order of Authors** Marco Chiappero, Francesca Demichelis, Xuan Lin, Chenxiao Liu, Dominic Frigon, Silvia Fiore

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Turin, August 28, 2019

Dear Prof. Stefano Dionisi, Associate Editor of Process Safety and Environmental Protection,

We would like to submit the revised version of the manuscript PSEP\_2019\_1124 "**Investigation of pre-treatments aimed at improving low-temperature anaerobic digestion of waste activated sludge**" by Marco Chiappero, Francesca Demichelis, Xuan Lin, Chenxiao Liu, Dominic Frigon and Silvia Fiore.

The manuscript was revised and improved according to the Reviewers' comments.

Thank you very much for your time and consideration

Sincerely yours,

Silvia Fiore (corresponding author, on behalf of all authors)

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**Prof. Dr. Silvia Fiore**

Department of Engineering for Environment, Land and Infrastructures (DIATI),

Politecnico di Torino, Corso Duca degli Abruzzi 24, 10129 Torino, Italy

E-mail: [silvia.fiore@polito.it](mailto:silvia.fiore@polito.it)



- Semi-continuous anaerobic digestion was performed at 20 °C on waste activated sludge
- Physicochemical pre-treatments were investigated to improve biogas yields
- The assessment was based on COD solubilization and on disintegration rate
- Thermo-alkaline pre-treatment (0.09 g NaOH/g TS, 70 °C, 60 min) gave best results
- Biogas yields (0.30-0.36 m<sup>3</sup>/kg VS, 65-70 % CH<sub>4</sub>) were analogous to 35 °C conditions

1 **Investigation of pre-treatments improving low-temperature anaerobic**  
2 **digestion of waste activated sludge**

3

4 Marco Chiappero<sup>a</sup>, Francesca Demichelis<sup>a</sup>, Xuan Lin<sup>b</sup>, Chenxiao Liu<sup>b</sup>, Dominic Frigon<sup>b</sup>,  
5 Silvia Fiore<sup>a,\*</sup>

6 <sup>a</sup>DIATI (Department of Environment, Land and Infrastructure Engineering), Politecnico  
7 di Torino, Corso Duca degli Abruzzi 24, 10129, Torino, Italy

8 <sup>b</sup>Department of Civil Engineering and Applied Mechanics, McGill University,  
9 Sherbrooke St. West 817, H3A 0C3, Montreal, Quebec, Canada

10

11 **\*Corresponding author:** Prof Silvia Fiore, [silvia.fiore@polito.it](mailto:silvia.fiore@polito.it)

12

13 **Abstract**

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16 physicochemical pre-treatments (thermal at 115 °C, thermo-alkaline at pH 10 and 70 °C  
17 and ozonation at 190 mg-O<sub>3</sub> L<sup>-1</sup>) by comparing their performances about COD  
18 solubilization and sludge disintegration rate. Best performances were obtained by  
19 thermo-alkaline pre-treatment, followed by thermal and ozonation; results were  
20 consistent with literature. Pre-treated WAS was fed to 12 1-L anaerobic digesters  
21 operated in semi-continuous mode. Thermal and thermo-alkaline reactors produced  
22 biogas yields (0.30-0.36 m<sup>3</sup> kg<sup>-1</sup> VS in standard conditions, 65-70 % methane) analogous  
23 to mesophilic conditions. The economic assessment of the scale-up of the whole process

24 demonstrated that thermo-alkaline pre-treatment made AD at 20 °C economically  
25 profitable for WAS generated by a 20,000 PE WWTP.

26 **Keywords:** anaerobic digestion; biogas; low-temperature; pre-treatment; semi-  
27 continuous; waste activated sludge.

28 **Abbreviations:** AD, anaerobic digestion; COD, chemical oxygen demand; DR,  
29 disintegration rate; EU, European Union; HRT, hydraulic retention time, OLR, organic  
30 retention time; OZ, ozone pre-treatment; PE, person equivalent;  $R_x$ , removal of X;  
31  $S_{COD}$ , solubilization ratio; SBP, specific biogas production; SMP, specific methane  
32 production; SRT, solids retention time; TA, thermo-alkaline pre-treatment; TH, thermal  
33 pre-treatment; TS, Total Solids; VS, Volatile Solids; WAS, waste activated sludge;  
34 WWTP, wastewater treatment plant

## 35 **1. Introduction**

36 Annual waste activated sludge (WAS) production in EU is expected to reach 13 Mt of  
37 dry solids by 2020 (Milieu Ltd et al., 2008), due to Urban Wastewater Treatment  
38 Directive 91/271/EC and to the restrictive limits on nutrients removal imposed by  
39 current legislations (Panepinto et al., 2016). In addition, WAS management could be  
40 responsible of up to 50 % of the operating costs in a wastewater treatment plant  
41 (WWTP) (Appels et al., 2008). Consequently the implementation of environmental and  
42 economic sustainable WAS management technologies is crucial for any WWTP. WAS  
43 disposal strategies in EU-27 between 2006 and 2010 were mostly based on agricultural

44 reuse (44 %), incineration (22 %), composting (15 %), landfilling (11 %) and others (8  
45 %) (Eurostat, 2018).

46 In EU-27 aerobic and anaerobic digestion (AD) are the most common WAS  
47 stabilization approaches (Kelessidis and Stasinakis, 2012). Anaerobic stabilization  
48 processes are usually preferred over aerobic ones for medium-sized WWTPs and larger  
49 because biogas can partially cover the energy requirements of the plant. AD is a  
50 complex degradation process involving four main phases: hydrolysis, acidogenesis,  
51 acetogenesis and methanogenesis (Van Lier et al., 2008). Hydrolysis, consisting in the  
52 disintegration of cells and complex organic structures into polymers, followed by their  
53 hydrolysis to simpler monomers, is regarded as the rate-limiting step (Bakhshi et al.,  
54 2018). The flocs structure and the presence of extra polymeric substances around the  
55 cells make WAS disintegration particularly critical (Zhen et al., 2017). This limitation is  
56 generally overcome by increasing the operating temperature to enhance the microbial  
57 activity (Appels et al., 2008). Mesophilic (35 °C) anaerobic digesters usually serve  
58 medium and large scale WWTPs and they can be energy self-sufficient for WWTPs  
59 sized over 50,000 person equivalents (PE). However, 70 % of Italian WWTPs are below  
60 20,000 PE (Istat, 2018), a situation rather common throughout the EU. Moreover, AD  
61 processes can be limited by high requirements of thermal energy in colder climate  
62 countries (Rajagopal et al., 2017). In this framework, developing psychrophilic (below  
63 20 °C) AD of WAS could be strategic; it has lower energy demand and has as main  
64 challenges the lower rate of fermentation as a consequence of decreased temperature  
65 and the low biodegradability of WAS during the initial phase of hydrolysis (Dev et al.,  
66 2019). Psychrophilic AD has been previously investigated for wastewater (Gomec,  
67 2010), animal manure (Saady and Massé, 2016) and food waste (Rajagopal et al., 2017).

68 To our knowledge very few studies are available about low-temperature AD processes  
69 implemented on WAS (Bakhshi et al., 2018; Dev et al., 2019), however the key role of  
70 an adapted inoculum was already demonstrated (Dolejs et al., 2018).

71 Biological, mechanical, thermal, chemical processes and their combinations have been  
72 extensively investigated as pre-treatments within mesophilic AD processes applied on  
73 WAS deriving from urban (Carrère et al., 2010) and industrial wastewater (Demichelis  
74 et al., 2018a). Thermal pre-treatment is well-established at full-scale (Zhen et al., 2017).  
75 The application of heat in a wide temperature range (60-180 °C) can disintegrate cell  
76 walls and membranes of the active biomass in WAS, leading to partial solubilisation of  
77 intracellular components (Tyagi and Lo, 2011). Alkaline pre-treatments were reported  
78 to induce the disruption of cells due to high pH values and reactions between the alkali  
79 agent and cell walls (Tyagi and Lo, 2011). However, an excess of alkali may inhibit AD  
80 (Carrère et al., 2010). For this reason, alkaline processes have been often combined with  
81 thermal treatment, with the aim of reducing both alkali dose (Ruffino et al., 2016) and  
82 process temperature (Uma Rani et al., 2012). A recent study (Bakhshi et al., 2018)  
83 comparing AD of WAS at 35 °C with AD at 20 °C after pre-treatment with ozone,  
84 revealed the latter to produce more energy. However, additional research about pre-  
85 treatments implemented on low-temperature AD of WAS is strongly needed, with a  
86 specific focus on enhancing the produced biogas compared to the energy spent in the  
87 process. The present work was aimed at assessing the technical feasibility of low-  
88 temperature (20 °C) AD of raw and pre-treated WAS to investigate whether its  
89 efficiency could be comparable with a mesophilic process. Compared to (Bakhshi et al.,  
90 2018), the adopted approach involved the optimization of process parameters and the  
91 comparison of three physicochemical pre-treatments (thermal, thermo-alkaline,

92 ozonation), assessing the increase in WAS solubilization and biodegradability in terms  
93 of solids removal and biogas production. An assessment of the economic profitability of  
94 the scale-up of the overall process chain concluded the research.

## 95 **2. Material and methods**

### 96 *2.1. Waste activated sludge*

97 12 WAS samples were collected once per week from Régie d'Assainissement des Eaux  
98 du Bassin La Prairie (RAEBL) WWTP (240,000 PE) in Saint Catherine, Quebec,  
99 Canada. The treatment outline was made of preliminary processes, biological process  
100 and settling. WAS samples were diluted from an initial total solids (TS) content of 4 - 5  
101 %-wt to approximately 3 %-wt TS prior pre-treatments to achieve a constant TS amount  
102 during the tests.

### 103 *2.2. Pre-treatments*

104 Thermal (TH), Thermo-alkaline (TA) and Ozone (OZ) pre-treatments were selected  
105 according to previously discussed literature, optimized (see Supplementary Material)  
106 and compared about the increased solubilization of the sludge, expressed as chemical  
107 oxygen demand (COD). Pre-treatments were performed just after WAS sampling, then  
108 the sludge was stored at 4 °C until use. Two assessment parameters were adopted:  
109 solubilization ratio ( $S_{\text{COD}}$ ), defined as the ratio between soluble and total COD  
110 (respectively sCOD and tCOD), and disintegration rate (DR), which is the ratio between  
111 the increase in sCOD due to the pre-treatment and the maximum possible variation in  
112 sCOD (Kim et al., 2013; Li et al., 2012) (Eq.1).

$$113 \quad DR(\%) = \frac{sCOD_T - sCOD_o}{tCOD_0 - sCOD_0} \cdot 100 \quad (1)$$

114 where  $sCOD_T$  is the soluble COD after treatment;  $sCOD_0$  and  $tCOD_0$  are respectively  
 115 the soluble and the total COD before treatment.

116 *2.2.1. Thermal and Ozone pre-treatment*

117 Thermal pre-treatment (TH) was carried out in a pressure cooker (Instant Pot, IP-  
 118 DUO80) on 0.75 L WAS samples at 115 - 118 °C and 0.8 bar for 30 min. Temperature  
 119 value was chosen according to previous studies (Tyagi and Lo, 2011; Carrere et al.,  
 120 2010) while the 30 min extent was defined after three COD solubilization tests (T1, T2  
 121 and T3) (Table 1 and Supplementary Material) performed on 0.3 L WAS samples. The  
 122 effect of the cooling phase at the end of TH was investigated by testing two cooling  
 123 modes: at room temperature and in ice-bath, the latter aiming to abruptly interrupt the  
 124 effect of the temperature at the end of pre-treatment.

125 Ozone pre-treatment (OZ) was performed as in Bakhshi et al. (2018), adopting an  
 126 average dose of 190 mg O<sub>3</sub> L<sup>-1</sup>.

127

128 **Table 1.** Operating conditions of COD solubilisation tests for the optimization of  
 129 thermal and thermo-alkaline pre-treatments.

<b>Thermal tests</b>	operating conditions		
	pre-treatment time (min)	cooling time (min)	cooling mode
T1	30, 60, 90, 120	30 – 40	room temperature
T2	10, 20, 30, 45, 60, 90, 120	15 - 20	ice bath
T3	30, 60, 90, 120	30 - 40	ice bath
	30, 60, 90, 120	15 - 20	room temperature
<b>Thermo-alkaline tests</b>	operating conditions		
	target pH	pre-treatment time (min)	
TA1, TA2, TA3	9	0, 30, 60, 90, 120	

---

10	0, 30, 60, 90, 120
11	0, 30, 60, 90, 120

---

130

131 *2.2.2. Thermo-alkaline pre-treatment*

132 Thermo-alkaline pre-treatment (TA) was operated at 70 °C for 60 min with 0.09 g  
 133 NaOH g<sup>-1</sup> TS (defined after three COD solubilisation tests TA1, TA2 and TA3) (Table 1  
 134 and Supplementary Material). Different doses of 5 N NaOH were added to 0.6 L WAS  
 135 samples until pH 9, 10 and 11, then each sample was split into five 0.1 L subsamples  
 136 further treated at 70 °C in a water bath for increasing time intervals (measured after  
 137 target temperature value was reached). Sludge samples were then cooled and pH was  
 138 adjusted to 7.0 - 7.5 with 10 N HCl.

139 *2.3. Anaerobic Digestion tests*

140 Twelve AD reactors (3 for each pre-treatment and 3 fed with raw WAS) were operated  
 141 at 20 °C in semi-continuous mode for 80 days with solids retention time (SRT) and  
 142 hydraulic retention time (HRT) equal to 15 days (Uma Rani et al., 2012). Each reactor  
 143 consisted of a 1-L Pyrex glass bottle, equipped with a polypropylene screw thread cap,  
 144 wrapped in aluminium foil and mixed through a magnetic stirrer (model 801, Apera  
 145 Instruments). Two holes in the cap allowed feeding and biogas collection in a 1-L gas  
 146 bag (30226-U, Supelco, Sigma-Aldrich). The experimental procedure started with a  
 147 start-up phase (30 days), in which the reactors were filled up to 0.8 L with digestate  
 148 from the mesophilic digester of the RAEBL WWTP as inoculum (Table 1). During the  
 149 start-up phase, lasted two SRTs, three times per week the reactors were fed by OZ-  
 150 WAS. Afterwards, the test phase lasted 50 days, corresponding to 3.3 SRTs, as it may  
 151 be assumed that steady state was reached after 3 SRTs (Bakhshi et al., 2018; Liao et al.,

152 2006). During the test phase, the four types of feeds (raw WAS, TH, OZ, TA) and the  
153 digestate from the reactors were characterized once per week evaluating soluble and  
154 total COD, total solids (TS) and volatile solids (VS). Biogas production was measured  
155 every 2 - 4 days and biogas was characterized at the end of the test phase. Digestate pH  
156 was checked at each feed (3 times per week).

#### 157 *2.4. Analytical procedures*

158 Chemical oxygen demand (COD) was measured through colorimetric method 5220D  
159 (APHA-AWWA-WEF, 2005). Prior sCOD analysis the samples were centrifuged at 20  
160  $\times 10^3$  g (Legend Micro 21, Sorvall™, Thermo Fisher Scientific centrifuge) and the  
161 supernatant was filtered through a 0.45  $\mu$ m membrane. TS and VS were analyzed by  
162 gravimetric methods 2540B and 2540E (APHA-AWWA-WEF, 2005). pH was  
163 measured with a Thermo Fisher Scientific 710A Orion pH/ISE meter. Daily specific  
164 biogas production (SBP) was measured through water displacement (Bakhshi et al.,  
165 2018) and referred to standard conditions. Methane content in biogas was analyzed by  
166 means of an Agilent 7820A gas chromatograph equipped with a PoraPLOT Q capillary  
167 column (25 m  $\times$  0.32 mm  $\times$  10  $\mu$ m, Agilent) and a TCD detector.

#### 168 *2.5. Sensitivity analysis*

169 All analyses were carried out in triplicates and average values are reported in the study  
170 together with standard deviation. Statistical tests on experimental data were carried out  
171 using data analysis extension of Microsoft Excel 2016. A correlation test investigated  
172 the presence of linear correlation between pairs of variables, considering significant  
173 those having  $p < 0.05$ .

174 *2.6. Scale up evaluation*

175 Energy and economic assessments were performed and simulated using SuperPro  
176 Designer® 8.0 software considering three scenarios: S0 - AD at 35 °C of WAS coming  
177 from secondary settling; S1 - AD at 35 °C of WAS coming from the same WWTP  
178 considered in this study; S2A - TH pre-treatment + AD at 20 °C; S2B - THA pre-  
179 treatment + AD at 20 °C; S2C - OZ pre-treatment + AD at 20 °C. In S0 and S1  
180 scenarios, sludge characteristics and biogas yields were respectively based on (Ruffino  
181 et al., 2016) and (Bakhshi et al., 2018). S2 scenarios were simulated considering the  
182 experimental data gathered in this work.

183 *2.6.1 Energy assessment*

184 The energy assessment was carried out under thermodynamic equilibrium and steady  
185 state conditions, considering atmospheric air (79 % <sub>v/v</sub> N<sub>2</sub> and 21 % <sub>v/v</sub> O<sub>2</sub>), assuming  
186 valid the ideal gas law and negligible gas leaks from connecting pipes (Mehr et al.,  
187 2017). The net energy load (Q<sub>n</sub>), expressed in MJ/d, was calculated considering the  
188 seasonal temperature average variations in Europe (IPCC, 2017) and it was expressed as  
189 the sum of energy consumed (Q<sub>c</sub>) and energy produced (Q<sub>p</sub>) (Eq. 2).

190 
$$Q_n = Q_c + Q_p \quad (2)$$

191 Q<sub>c</sub> was the sum of: energy required (Q<sub>req</sub>) to heat the pre-treatment units (TH at 118 °C,  
192 TA at 70 °C, O at 20 °C) and AD reactor (at 20 °C) (Eq. 3); energy to mix (Q<sub>mix</sub>) the  
193 pre-treatment units and AD reactor (Eq.4); energy losses (Q<sub>loss</sub>) from external and  
194 ground walls of AD reactor (Eq. 5); energy to transfer ozone (Q<sub>O3</sub>) to the inoculum and  
195 to perform OZ pre-treatments (Eq. 6).

196 
$$Q_{req} = m_{sludge} \cdot c_{sludge} \cdot (T_{reac} - T_{in}) \quad (3)$$

197 where  $m_{\text{sludge}}$  is the sludge mass flow rate [kg/d], while  $T_{\text{reac}}$  and  $T_{\text{in}}$  are respectively the  
 198 reactor and inlet temperatures, and  $c_p$  is the specific heat capacity ( $4200 \frac{\text{J}}{\text{kg} \cdot ^\circ\text{C}}$ )

$$199 \quad Q_{\text{mix}} = P_{\text{mix}} \cdot t_{\text{mix}} \quad (4)$$

200 where  $P_{\text{mix}}$  is the mixing power [J/h] and  $t_{\text{mix}}$  the required time to mix the sludge [h/d]

$$201 \quad Q_{\text{loss}} = U_{\text{ug}} \cdot A_{\text{ug}} \cdot (T_{\text{reac}} - T_{\text{gr}}) + U_{\text{ext}} \cdot A_{\text{ext}} \cdot (T_{\text{reac}} - T_{\text{ext}}) \quad (5)$$

202 where according to (Mehr et al., 2017)  $U_{\text{ug}}$  and  $U_{\text{ext}}$  are respectively the coefficients of  
 203 heat transfer for underground walls ( $2.33 \frac{\text{W}}{\text{m}^2 \cdot ^\circ\text{C}}$ ) and external walls ( $0.93 \frac{\text{W}}{\text{m}^2 \cdot ^\circ\text{C}}$ );  $A_{\text{ug}}$  and  
 204  $A_{\text{ext}}$  are respectively the areas of underground walls and external walls;  $T_{\text{gr}}$  and  $T_{\text{ext}}$  are  
 205 respectively the temperatures of underground walls and partial walls.

$$206 \quad Q_{\text{O}_3} = O_{3 \text{ dose}} \cdot m_{\text{sludge}} \cdot \text{Elec}_{\text{O}_3} \quad (6)$$

207 where  $\text{Elec}_{\text{O}_3}$  is the energy required to perform OZ and according to (Bakhshi et al.,  
 208 2018), equal to  $12.5 \frac{\text{kWh}}{\text{kgO}_3}$ .

209  $Q_p$  was the sum of two items: energy from methane production ( $Q_{\text{CH}_4}$ ) (Eq. 7) and  
 210 energy from heat recovery ( $Q_r$ ) (Eq. 8)

$$211 \quad Q_{\text{CH}_4} = V_{\text{CH}_4} \cdot \eta_{\text{el}} \cdot 39.4 \frac{\text{MJ}}{\text{m}^3} \quad (7)$$

212 where  $\eta_{\text{el}}$  is assumed to be 0.35.

$$213 \quad Q_r = \eta \cdot (T_{\text{ex-hot}} - T_{\text{ex-cold}}) \cdot m_{\text{sludge}} \cdot c_{\text{sludge}} \quad (8)$$

214 where  $\eta$  is the heat exchanger efficiency equal to 70% according to (Ruggeri et al.,  
 215 2015),  $T_{\text{ex-hot}}$  is the temperature of the AD reactor ( $20 \text{ }^\circ\text{C}$ ) and  $T_{\text{ex-cold}}$  depends on the  
 216 season.

217 The energy sustainability is achieved if the energy sustainability index (ESI) (Eq. 9) is  
 218 higher than 1.

219 
$$ESI = \frac{Q_p}{Q_c} \quad (9)$$

220

### 221 *2.6.2 Economic assessment*

222 The economic analysis was aimed to define the minimum plant size able to be  
223 economically profitable considering 365 working days per year. The assessment was  
224 based on the experimental data presented in this work and related to existing AD plants  
225 (Table 2), while costs evaluation was consistent with Chemical Engineering Plant Cost  
226 Index (Peters and Timmerhaus, 2003). Economic analysis considered capital and  
227 operational costs and revenues. Capital costs were made of fixed capital investment  
228 (FCI, consisting in equipment purchase for plant construction and working capital cost,  
229 which is 6.5 % of FCI) (Pommerat et al, 2017). The cost of land wasn't taken into  
230 account since the AD plant was hypothesized in the WWTP area. A 5-years  
231 amortization with a 2 % interest was assumed for the capital costs (Eq. 10):

232 
$$A[Euro] = C_0 \cdot \frac{i \cdot (1 + i)^n}{(1 + i)^n - 1} \quad (10)$$

233 where  $A$  is the amortization cost,  $C_0$  is the initial capital cost  $i$  is the interest and  $n$  the  
234 number of years considered for amortization. Operational costs included utilities,  
235 digestate disposal and labor costs (Table 2). Sludge collection and transport were not  
236 accounted since the AD plant was hypothesized in the WWTP area. This assumption  
237 was the core of the further assessment of the scale-up of the overall process. Our idea  
238 was to optimize WAS management in WWTPs through an on-site process, with two  
239 positive consequences: 1. Biogas/methane production, which is needed to heat the  
240 digester and could eventually, if in excess, be valorized to fulfil the energy needs of the

241 WWTP; 2. Decreasing the costs of the final disposal of the digestate (whose volume is  
 242 inferior compared to WAS).  
 243 Labour cost is considered an addition to the current staff of the WWTP; consequently 2,  
 244 3, 4 and 5 workers were hypothesized respectively for WWTPs serving 5,000 to 20,000  
 245 PE; 50,000 to 200,000 PE; 500,000 PE and 1,000,000 PE.  
 246 The revenues came from the surplus of energy produced in the plant from AD and heat  
 247 recovery. The annual income was calculated as the difference between the revenue and  
 248 the amortization for the first 5 years and operational costs. The profitability was  
 249 evaluated through: return of interest (ROI) (Eq. 11), net present value (NPV) (Eq. 12)  
 250 assuming 20 years plant lifetime with 5 % discount on the future cash flows to the  
 251 present value, according to (Demichelis et al. 2018b).

$$252 \quad ROI [\%] = \frac{\text{Annual net profit}}{\text{Initial total investment}} \cdot 100 \quad (11)$$

253 NPV represents the scenario profitability for the plant lifetime (20 years) considering a  
 254 5 % discount on the future cash flows to the present value.  $NPV > 0$  means that the  
 255 process is profitable.

$$256 \quad NPV [Euro] = \sum_{t=1}^T \frac{C_t}{(1+d)^t} - C_0 \quad (12)$$

257 where  $t$  is the plant lifetime,  $C_t$  is the net cash flow during period  $t$ ,  $C_0$  is the initial  
 258 capital investment and  $d$  is the discount rate. To conclude the economic profitability  
 259 assessment, Payback time is the time required to regain the investment cost.

260

261 **Table 2.** Details of economic analysis: capital and operational costs and energy values

Investment costs		
Equipment	Unit	reference

Reactor	€/m <sup>3</sup>	2514.7	Dahiya et al., 2018
Stirrer	€/kW	46465.3	Akeberg and Zacchi, 2000
<b>Operational costs</b>			
Inoculum	€/m <sup>3</sup>	4.1	Wingren et al. 2003
NaOH	€/kg	0.27	Sigma-Aldrich, 2018
Digestate disposal	Euro/t	0.55	Arpa, 2017
Labour	€/year	44978	Eurostat, 2018
<b>Revenue</b>			
Energy value	€/kWh	0.22	SNAM, 2018

262

### 263 3. Results and discussion

#### 264 3.1. Pre-treatments

265 The characteristics of inoculum, raw WAS and pre-treated WAS are shown in Table 3.

266

267 **Table 3.** Physico-chemical characteristics of inoculum, raw WAS and pre-treated WAS  
 268 (TH: thermal pre-treatment; OZ: ozone pre-treatment; TA: thermo-alkaline pre-  
 269 treatment)

Parameter	inoculum	raw WAS	TH	OZ	TA
Dose				189 ± 53mg O <sub>3</sub> g <sup>-1</sup> TS	0.08 ± 0.01g NaOH g <sup>-1</sup> TS
pH	-	6.2 ± 0.3	6.0 ± 0.2	6.1 ± 0.3	7.3* ± 0.3
TS (g L <sup>-1</sup> )	25.5 ± 0.5	33.1 ± 1.9	33.2 ± 2.1	33.2 ± 2.0	36.0 ± 2.4
VS (g L <sup>-1</sup> )	17.2 ± 0.4	25.3 ± 1.4	25.3 ± 1.6	25.4 ± 1.5	25.3 ± 1.5
VS / TS (%)	-	76.9 ± 1.7	76.2 ± 1.8	76.5 ± 1.8	70.5 ± 1.9
tCOD (g O <sub>2</sub> L <sup>-1</sup> )	17.2 ± 0.6	37.8 ± 2.9	37.8 ± 2.0	39.2 ± 3.3	38.0 ± 2.7

sCOD (g O <sub>2</sub> L <sup>-1</sup> )	1.3 ± 0.1	2.7 ± 1.0	9.9 ± 1.3	4.0 ± 1.1	14.3 ± 0.8
sCOD/tCOD (%)	-	7.3 ± 3.2	26.2 ± 3.6	10.2 ± 2.9	37.8 ± 4.0
DR (%)	-	-	20.5 ± 4.2	3.7 ± 0.9	33.3 ± 3.3

\*after pH conditioning

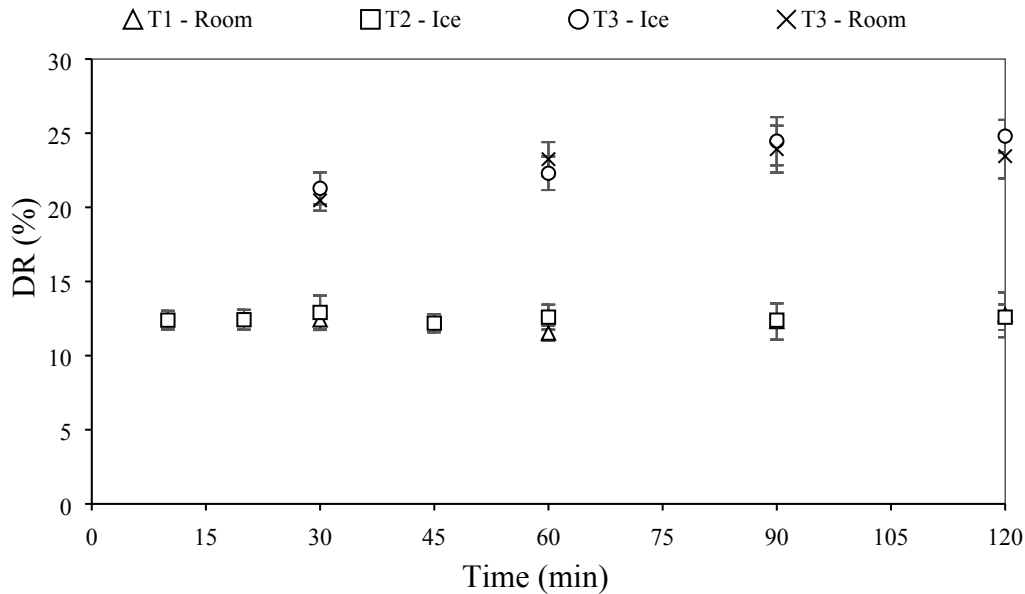
### 270 3.1.1. Optimization of operating conditions

271 TH was operated at 115 – 118 °C. Higher temperatures (160 - 180 °C) were shown to be  
272 more efficient (Bougrier et al., 2008; Carrère et al., 2010) but high energy demanding  
273 (Appels et al., 2010) Low-temperature (< 100 °C) pre-treatments need longer durations  
274 (from hours to days) (Ferrer et al., 2008; Xu et al., 2014). Therefore an intermediate  
275 temperature value was preferred (Ennouri et al., 2016; Jeong et al., 2007). The influence  
276 of the pre-treatment time on COD solubilization was investigated within three tests (T1,  
277 T2 and T3) (Table 1). The obtained DR values were: 11 – 13 % in T1 and T2; 21 – 25  
278 % in T3 (Figure 1), while the starting ratio between sCOD and tCOD before treatment  
279 was comparable (around 8 %). Full details about T1, T2 and T3 tests are in  
280 Supplementary Material. The gathered results are consistent with literature: Kim et al.  
281 (2003) obtained 10.4 % DR by autoclaving WAS (38.0 g L<sup>-1</sup> TS) at 121°C and 1.5 atm  
282 for 30 min; a thermal pre-treatment on WAS at 120 °C for 30 min led to 22 – 23 % DR  
283 (Jeong et al., 2007); heating WAS (14.26 ± 2.18 g L<sup>-1</sup>) at 121 °C and 1 bar for 15 min  
284 produced 15.7 % DR (Salsabil et al., 2010). A slight influence of time on solubilization  
285 of sludge during a thermal pre-treatment of WAS at 130 °C was already observed (Valo  
286 et al., 2004). Our research did not find a significant influence of pre-treatment time in  
287 improving COD solubilization, as DR was already stable after 30 min. The cooling  
288 mode at room temperature or in ice-bath did not determine significant influences on  
289 COD solubilization (see Figure 1).

290

291 **Figure 1.** Optimization of thermal pre-treatments through COD solubilisation tests (T1,

292 T2, T3) with two cooling modes (ice bath and at room temperature)



293

294 TA was based on sodium hydroxide, which was found to determine better solubilisation

295 than other alkali agents (Kim et al., 2003). The alkali doses corresponding to pH 9, 10

296 and 11 were selected from literature (Uma Rani et al., 2012; Xu et al., 2014). The

297 temperature value of 70 °C was chosen as a compromise between pre-treatment

298 performance and energy costs (Kim et al., 2013; Ruffino et al., 2016). The effect on

299 COD solubilization of the alkali dose and the thermal pre-treatment time was assessed

300 by three tests: TA1, TA2 and TA3 (see section 2.2.1 and Figure 2). Full details are

301 reported in Supplementary Material. The doses of NaOH needed for reaching pH 9, 10

302 and 11 were recorded during each test (Figure 2a). It was found a significant linear

303 positive correlation between the alkali dose and the pH increase ( $r(7) = 0.954$ ,  $p <$

304  $0.05$ ). Figure 2b shows the DR trends for different pH values (and doses of NaOH) with

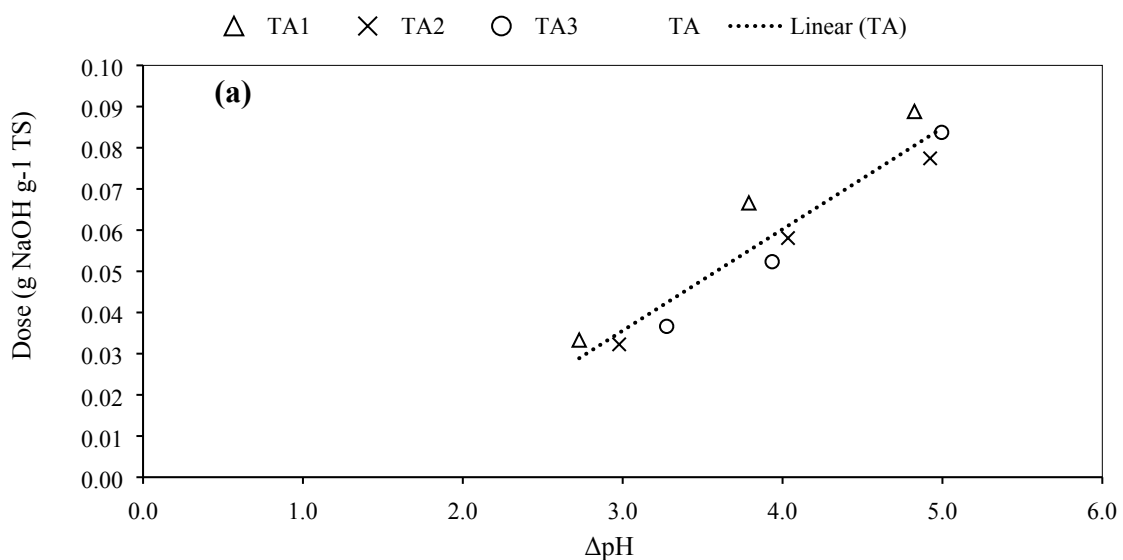
305 and without any thermal pre-treatment at 70 °C for increasing times. An enhancement of

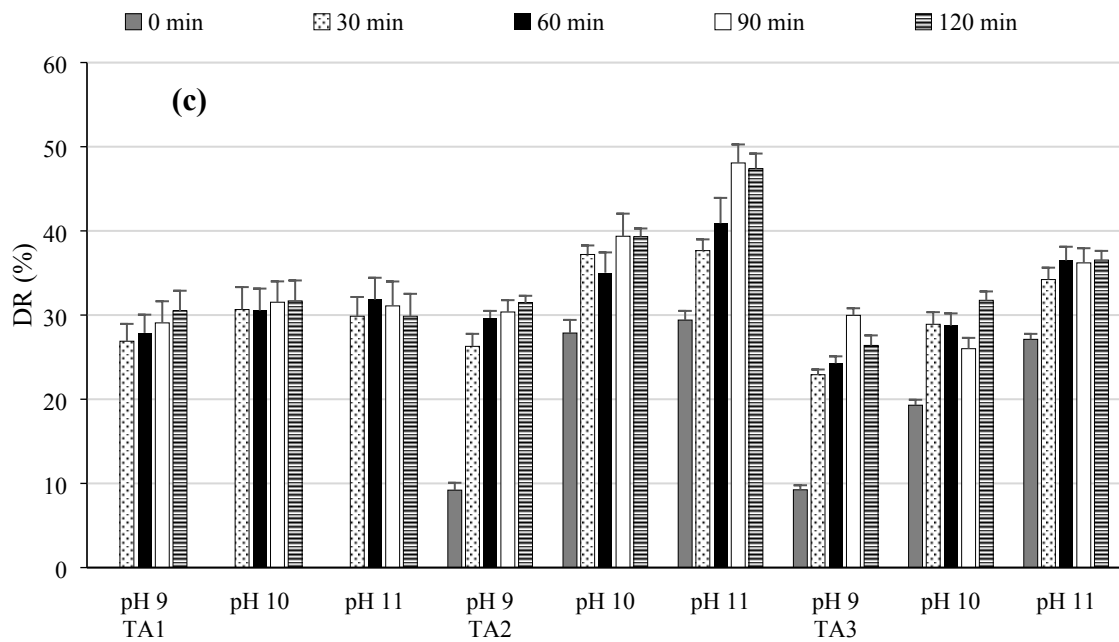
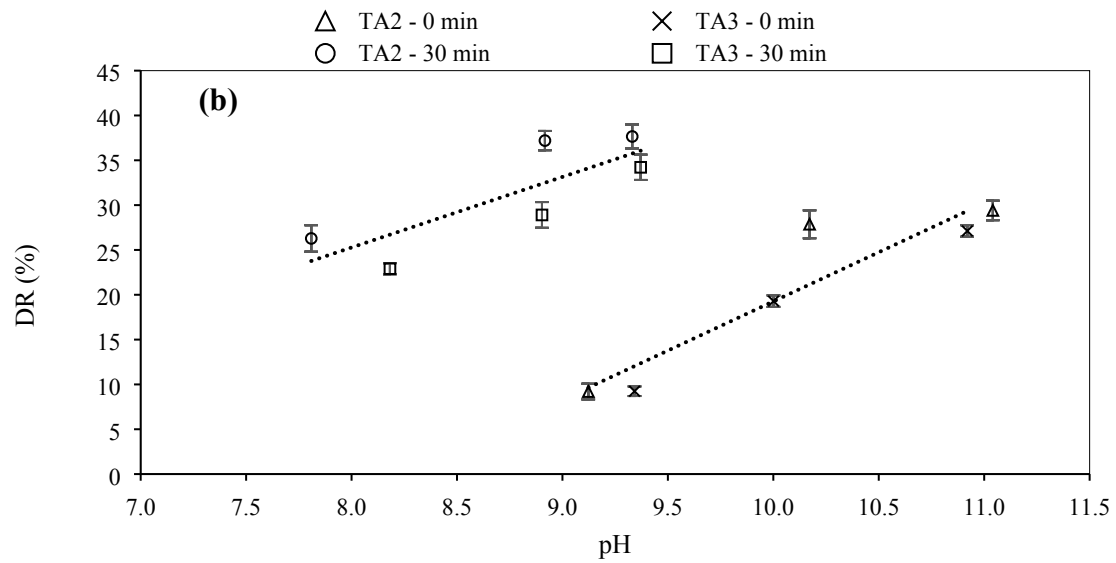
306 COD solubilisation linked to pH variations was observed: an increase of alkali dose  
307 from pH 9 to 11 at room temperature determined a DR increase from 10 % to almost 30  
308 % revealing a significant positive linear correlation between the initial pH and DR ( $r(4)$   
309 = 0.928,  $p < 0.05$ ). These DR values are in good agreement with previous studies  
310 (Ruffino et al., 2016; Li et al. 2012). As for tests T1, T2 and T3, COD solubilisation  
311 was linked to pH variations showing a significant linear correlation between pH and DR  
312 ( $r(4) = 0.815$ ,  $p < 0.05$ ). However, comparing DR values obtained by adding NaOH  
313 with or without thermal pre-treatment, it seemed that the lower duration of the pre-  
314 treatment emphasised the effect of pH increase (Figure 2b). Further confirmations of  
315 these patterns are in Supplementary Material. Figure 2c shows DR values obtained for  
316 each combination of pH and thermal pre-treatment time in TA1, TA2 and TA3. The  
317 results of each test were grouped into pH 9, 10 and 11 and each dose presented a group  
318 of bars corresponding to increasing heating times from left to right. Overall, DR values  
319 ranged from almost 10 % after the lower dose of NaOH to close to 50 % after the  
320 thermal pre-treatment. As already observed, the effect of an increased dose of NaOH (0  
321 min) is evident from TA2 and TA3, as well as the increase of DR due the thermal pre-  
322 treatment (0 min versus 30 min). However, a variation of the pre-treatment time did not  
323 seem to enhance COD solubilisation at 70 °C. For instance, in TA1 at pH 9 DR values  
324 varied from 27 % to 30 % for pH 10 and pH 11 were respectively 31 – 32 % and 30 –  
325 32 %. In TA3 DR values after the thermal pre-treatment at pH 9, 10 and 11 varied  
326 respectively from 23 % to 30 %, from 29 % to 32 % and from 34 % to 37 %, Only TA2  
327 seemed to suggest a slight effect of the duration on COD solubilisation. These results  
328 are in agreement with Appels et al. (2010): a moderate increase of sCOD was observed,  
329 if compared to higher temperatures, when heating WAS at 70 °C for 15 - 60 min;

330 furthermore, their results were in line with the outcomes of our thermal tests T1, T2 and  
 331 T3. A significant enhancement of COD solubilization, reaching DR values of 32 % in  
 332 TA1, 41 % in TA2 and 36 % in TA3 was observed, thanks to synergic effects of TA  
 333 pre-treatment, while avoiding the use of a high NaOH dose and its possible inhibition  
 334 problems (Li et al., 2012; Penaud et al., 1999) as well as limiting the energy  
 335 expenditures due to higher temperatures and pre-treatment durations. On the grounds of  
 336 the gathered results, we adopted the combination of pH 11 (0.089 g NaOH g<sup>-1</sup> TS) and  
 337 60 min pre-treatment time as optimum for the subsequent AD tests.

338  
 339

340 **Figure 2.** Optimization of thermo-alkaline pre-treatments through COD solubilisation  
 341 tests (TA1, TA2, TA3): (a) dose of NaOH as a function of the pH increase; (b)  
 342 Disintegration Rate as a function of pH after 0 min and 30 min of TH; (c) Disintegration  
 343 Rate for different combinations of pH and treatment times





344

345

346 *3.1.2. Effect of pre-treatments on solubilization and characteristics of waste activated*

347 *sludge*

348 The optimal operating conditions for TH (115 – 118 °C for 30 min), OZ (190 mg O<sub>3</sub> L<sup>-</sup>

349 <sup>1</sup>) and TA (0.09 g NaOH g<sup>-1</sup> TS at 70 °C for 60 min) pre-treatments were selected as

350 described in section 3.1.1. Table 1 reports the mean physico-chemical characteristics of

351 raw and pre-treated WAS according to the optimal operating conditions adopted during  
352 the AD tests. The pH of raw WAS, equal to 6.2, was close to the values measured on  
353 TH and OZ samples. However, pH 9.0 - 9.5 of WAS after TA was adjusted to  $7.3 \pm 0.3$   
354 before AD tests, to reach the optimum pH range for methanogenic bacteria, equal to 6.5  
355 - 7.2 (Appels et al., 2008). TS were around  $32 - 33 \text{ g L}^{-1}$  for raw WAS, TH and OZ,  
356 while TA showed an increase up to  $36 \text{ g L}^{-1}$  as a result of the use of NaOH. Higher TS  
357 compared to raw WAS after alkali addition were previously observed (Valo et al.,  
358 2004), investigating a thermo-alkaline pre-treatment on WAS at  $130 \text{ }^\circ\text{C}$  and pH 10. VS  
359 were stable for all samples around  $25 \text{ g L}^{-1}$ . Accordingly, the organic content of TA  
360 sludge was affected by TS variation down to 70 % compared to the 76 – 77 % of other  
361 samples.

362 Overall, the optimized pre-treatments seemed to enhance the sludge solubilisation. The  
363 sCOD of pre-treated samples increased compared to raw WAS: the sCOD of TA sludge  
364 raised by 4.3 times, sCOD of TH by 2.6 times while sCOD of OZ by 0.5 times.  
365 Moreover, the disintegration rate values after different pre-treatments were:  $\text{DR}_{\text{TA}} >$   
366  $\text{DR}_{\text{TH}} > \text{DR}_{\text{OZ}}$ .  $\text{DR}_{\text{TA}}$  value of 33 % was consistent with the results of TA1, TA2 and  
367 TA3 tests (see Supplementary Material) for pH 11 and 60 min. This value can be  
368 compared with other studies: Ruffino et al. (2016) and Campo et al. (2018) obtained  
369 DR values of 25 - 30% on WAS after thermo-alkaline pre-treatment at  $70 \text{ }^\circ\text{C}$  for 90 min  
370 dosing  $0.04 - 0.08 \text{ g NaOH g}^{-1} \text{ TS}$ . DR values of 64.8 % and 68.7 % were found by Kim  
371 et al. (2013) treating WAS with 0.1 M (about  $0.24 \text{ g NaOH g}^{-1} \text{ TS}$ ) and 0.2 M of NaOH  
372 at  $75 \text{ }^\circ\text{C}$  for 6 hours. Demichelis et al. (2018a) achieved 39 % DR after a thermo-  
373 alkaline treatment ( $0.08 \text{ g NaOH g}^{-1} \text{ TS}$ ) for 15 min at  $50 \text{ }^\circ\text{C}$  on industrial WAS. In  
374 addition, 21 %  $\text{DR}_{\text{TH}}$  (achieved in October-December 2017) was close to the results of

375 test T3 (November 2017) but significantly different from those of T1 and T2  
 376 (September 2017), in accordance with the previous hypothesis. These results were in  
 377 agreement with literature: a thermal treatment on WAS at 121 °C for 30 min gave a DR  
 378 of 10.5 % (Kim et al., 2003); at 121 °C under 1 bar for 15 min a led to a DR of 15.7 %  
 379 (Salsabil et al., 2010). DR<sub>OZ</sub> around 4 % was significantly lower than the values  
 380 achieved from other pre-treatments. The dose of 190 mg O<sub>3</sub> L<sup>-1</sup> (corresponding to 0.01 g  
 381 O<sub>3</sub> g<sup>-1</sup> TS) adopted in the present study seemed to be too low to determine a significant  
 382 COD solubilisation. The reported optimum dose of O<sub>3</sub> ranged between 0.05 and 0.5 g  
 383 O<sub>3</sub> g<sup>-1</sup> TS (Zhen et al., 2017). Bougrier et al. (2006) adopted 0.16 g O<sub>3</sub> g<sup>-1</sup> TS obtaining  
 384 a DR value around 22 % for WAS. In conclusion, TH and TA pre-treatments were able  
 385 to enhance WAS solubilisation, while the adopted ozone dose was too low.

386

387 **Table 4.** Physico-chemical characteristics of digested sludge of each group of reactors  
 388 after three SRTs (SBP: specific biogas production; SMP: specific methane production)

<b>Parameter</b>	<b>Raw WAS</b>	<b>TH</b>	<b>OZ</b>	<b>TA</b>
pH	7.00 ± 0.10	7.09 ± 0.08	7.06 ± 0.05	7.28 ± 0.05
TS (g L <sup>-1</sup> )	32.3 ± 0.04	24.0 ± 2.3	29.6 ± 0.3	28.1 ± 0.3
TS removal (%)	1.0 ± 0.1	26.5 ± 7.0	9.4 ± 0.8	14.1 ± 1.1
VS (g L <sup>-1</sup> )	23.5 ± 0.3	16.6 ± 1.2	20.9 ± 0.4	16.7 ± 0.2
VS removal	6.5 ± 1.2	33.8 ± 4.8	16.7 ± 1.7	33.2 ± 0.7
VS/TS (%)	72.5 ± 0.9	69.2 ± 1.4	70.6 ± 0.8	59.6 ± 1.4
VS/TS removal (%)	5.5 ± 1.2	9.8 ± 1.8	8.0 ± 1.0	22.3 ± 1.8
tCOD (g O <sub>2</sub> L <sup>-1</sup> )	37.4 ± 0.8	27.9 ± 3.8	31.7 ± 0.9	28.1 ± 2.8
sCOD (g O <sub>2</sub> L <sup>-1</sup> )	1.8 ± 0.02	3.6 ± 0.2	2.4 ± 0.1	3.7 ± 0.2
SBP (m <sup>3</sup> kg <sup>-1</sup> VS <sub>IN</sub> )	0.21 ± 0.03	0.30 ± 0.03	0.24 ± 0.02	0.36 ± 0.001
CH <sub>4</sub> content (%)	69.3 ± 1.2	66.3 ± 1.6	69.2 ± 0.9	70.5 ± 0.2
SMP (m <sup>3</sup> kg <sup>-1</sup> VS <sub>IN</sub> )	0.15 ± 0.02	0.20 ± 0.02	0.17 ± 0.02	0.25 ± 0.001

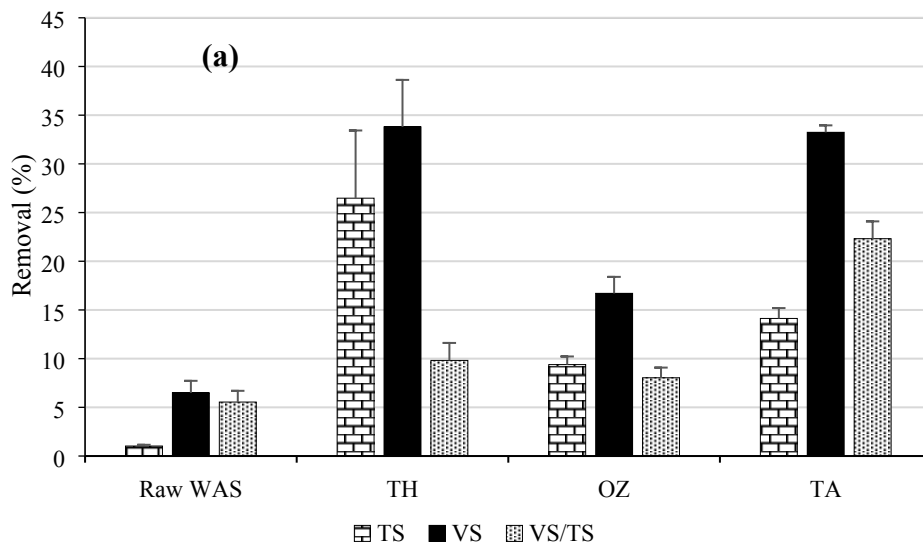
389

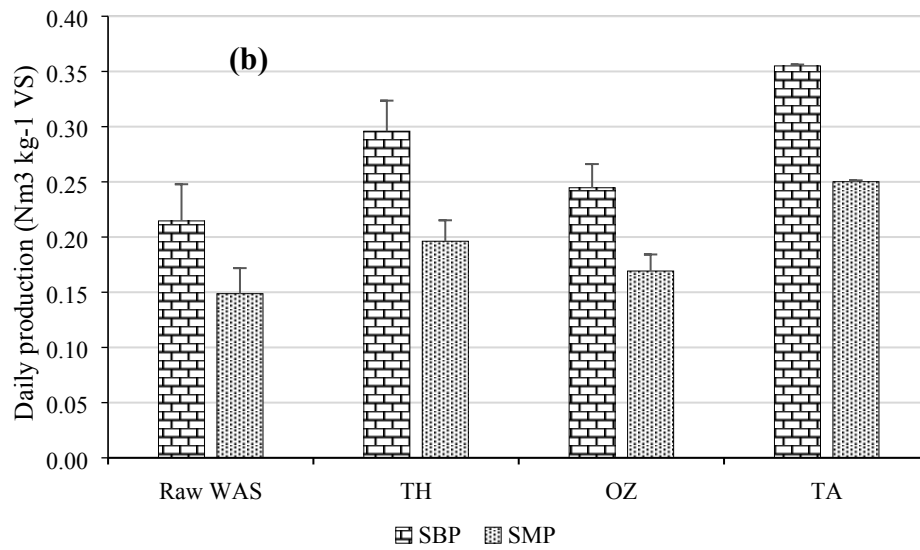
390 3.2. Anaerobic digestion tests

391 The performances of the different pre-treatments in improving AD were compared in  
392 terms of solids removal and biogas production. Table 4 summarizes the mean physic-  
393 chemical characteristics of digested sludge of each group of triplicate reactors at steady  
394 state. Overall, after three SRTs the pH values of digestate were close to neutrality. The  
395 daily organic loading rate (OLR) was evaluated for all AD tests at each feeding  
396 operation to prevent any overload problems (data not shown). The average daily OLR  
397 was 1.7 g VS<sub>IN</sub> L<sup>-1</sup>d<sup>-1</sup> for all reactors.

398

399 **Figure 3.** Results of anaerobic digestion test at 20 °C: (a) solids removal (TS, VS and  
400 VS/TS); (b) specific biogas production (SBP) and specific methane production (SMP)





401

402 *3.2.1. Solids removals*

403 The first two main objectives of AD of any substrate are the reduction of the solids  
 404 content, assessed by the removal of TS, and its stabilisation (evaluated through the  
 405 removal of VS and VS/TS ratio). Figure 3a shows TS removal ( $R_{TS}$ ) for the different  
 406 samples after three SRTs. It could be pointed out that TS removal was almost negligible  
 407 in the case of raw WAS. However,  $R_{TS}$  exceeded 25 % for TH, was around 15 % for TA  
 408 and about 10 % for OZ. The difference between removals of TH and TA could have  
 409 been affected by NaOH addition, as previously mentioned. Hence there was not  
 410 evidence of a better performance of TH compared to TA in terms of TS removal. VS  
 411 removal showed a clear difference between TH and TA. The higher mineral content  
 412 after TA helped to reduce VS/TS ratio of the digested sludge. Further confirmation  
 413 comes from the fact that VS removal for TH and TA were around 33 – 34 %. In the case  
 414 of OZ,  $R_{VS}$  was about 17 % and lower than 7 % for raw WAS. These results seemed to  
 415 be partially consistent with the increased COD solubilization of WAS by pre-treatments.  
 416 On one hand, the low COD solubilization induced by OZ corresponded to low removals  
 417 of TS and VS. On the other hand, the solubilization occurred for TA appeared to be

418 significantly larger than that of TH, however this was not followed by higher solids  
419 removals during subsequent digestion process.

420 The gathered results in terms of solids removals can be compared with other studies  
421 carried out in mesophilic conditions. The results of TA were lower than those of Xu et  
422 al. (2014): in their study a thermo-alkaline treatment at pH 11 with NaOH for 10 h at 90  
423 °C (DR = 43.7 %) and a thermal treatment at 70 °C for 9 h (DR = 27.9 %) led to VS  
424 removal respectively equal to 46.2 % and 43.7 %, while 38.9 % was recorded for raw  
425 WAS from batch AD tests. It can be noticed that both the duration and temperature of  
426 thermo-alkaline pre-treatment were higher in Xu et al. (2014) than in the present study  
427 and the removal of VS was significant also for the untreated WAS. However, our results  
428 are consistent with the ones achieved by (Uma Rani et al., 2012) from AD tests in semi-  
429 continuous mode after a thermo-alkaline pre-treatment on WAS with NaOH at pH 12  
430 and 60 °C for 60 min, which resulted in a removal of TS and VS concentrations of 25.1  
431 % and 33 %, in comparison with 9.6 % and 17 % for raw WAS. In addition, VS  
432 removal gathered from TH was comparable to the results of Kim et al. (2003) related to  
433 a thermo-alkaline pre-treatment with NaOH (7 g L<sup>-1</sup>) at 121 °C for 30 min and a thermal  
434 pre-treatment at 121 °C for 30 min, which determined VS removals respectively about  
435 45 % and 30 %, in comparison with less than 15 % for raw WAS. On the contrary,  
436 higher solids removals were found by Ennouri et al. (2016) who performed a thermal  
437 pre-treatment at 120 °C and 1.5 atm for 30 min on urban and industrial WAS  
438 determining respectively 74 % and 71 % VS removals in comparison with 48 % and 56  
439 % for raw WAS. It was not possible to find literature data concerning TS and VS  
440 removal during AD of WAS at 20 °C. In conclusion, it could be stated that WAS

441 biodegradability at 20 °C could be enhanced by TH and TA pre-treatments, while the  
442 biodegradability of raw WAS seemed very low.

443

### 444 3.2.2. *Biogas and methane production*

445 Figure 3b shows SBP and SMP values achieved from AD at 20 °C at steady state. SBP  
446 for raw WAS was around 0.2 m<sup>3</sup> kg<sup>-1</sup> VS<sub>IN</sub> with 70 % methane. SBP values were  
447 enhanced by pre-treatments: biogas production increased, compared to raw WAS, of 65  
448 % for TA, 38 % for TH and 14 % for OZ. In addition, SBP values are consistent with  
449 enhanced COD solubilization due to pre-treatments. TH and TA pre-treatments resulted  
450 in similar VS removal values but TA led to higher biogas production. Methane  
451 concentration appeared promising for all samples, ranging between 65 and 70 %-vv  
452 (Table 4). Different pre-treatments did not seem to affect methane percentage in  
453 comparison with raw WAS. As a result, SMP revealed an increase, compared to raw  
454 WAS, of 68 % for TA, 32 % for TH and 14 % for OZ. SMP values are consistent with  
455 the increased solubilization of organic matter produced by different pre-treatments, even  
456 though TH and TA seemed to be equivalent in terms of solids removals. Overall, SBP  
457 values varied from 0.21 (raw WAS) to 0.36 Nm<sup>3</sup> kg<sup>-1</sup> VS<sub>IN</sub> (TA), while SMP values  
458 from 0.15 (raw WAS) to 0.25 m<sup>3</sup> kg<sup>-1</sup> VS<sub>IN</sub> (TA), in agreement with (Dolejs et al.,  
459 2018), who observed SMP values equal to 0.22 m<sup>3</sup> kg<sup>-1</sup> of COD added.

460 The results of this work are consistent with those of other studies carried out in  
461 mesophilic conditions. Considering TH, a thermal pre-treatment on WAS at 121 °C for  
462 60 min enhanced SBP from 0.35 to 0.42 L g<sup>-1</sup> VSS<sub>IN</sub> (Barjenbruch and Kopplow, 2003),  
463 while at 90 °C for 60 min increased SBP from 0.035 to 0.377 L g<sup>-1</sup> VS (Appels et al.,  
464 2010). The increase of methane production due to TA was comparable with the results

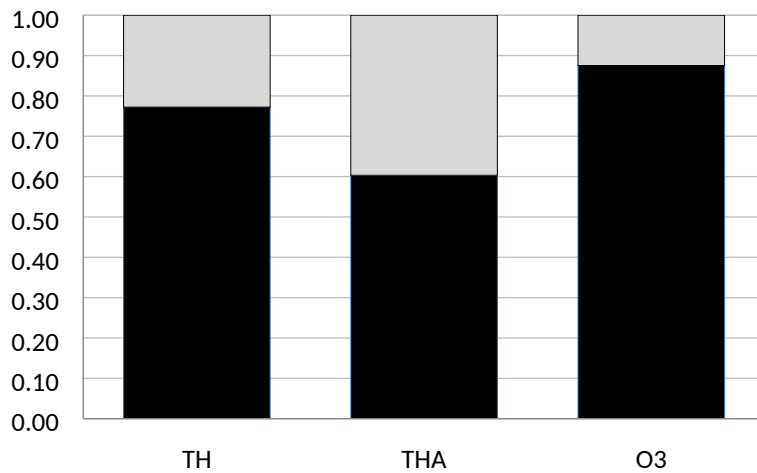
465 of (Kim et al., 2013): a thermo-alkaline pre-treatment of WAS (12 g TS L<sup>-1</sup>) with NaOH  
466 at 75 °C for 6 h achieved 70 % increase of methane production. As already mentioned, a  
467 thermo-alkaline pre-treatment of WAS (3 % TS) with NaOH (pH 11) at 90 °C for 10 h  
468 to an increase of biogas production from 0.396 to 0.605 L g<sup>-1</sup> VS (Xu et al., 2014). Our  
469 TA results are also comparable with SBP values obtained by Ruffino et al. (2016), who  
470 performed a thermo-alkaline pre-treatment with 0.04 g NaOH/g TS at 70 °C for 90 min  
471 on WAS (5-6 % TS), obtaining an increase of the biogas production from 0.236 to 0.299  
472 m<sup>3</sup> kg<sup>-1</sup> VS (+26.8 %) through mesophilic AD in batch mode.

### 473 3.3. Scale-up: energy assessment

474 The energy assessment was performed for three scenarios (see section 2.6). The  
475 calculated ESI values were higher than 1 for all scenarios: from 200,000 PE for S0;  
476 from 100,000 PE for S1 and from 2,000 PE for S2 A, B and C. In details, ESI for S0  
477 was 1.03 ± 0.10; for S1 1.04±0.08 and S2 A (TH + AD), B (TA +AD) and C (OZ +AD)  
478 were 2.09 ± 0.11, 2.63 ± 0.09 and 1.03±0.11 respectively. The standard deviation  
479 considered the temperature variation of WAS from January to December. The achieved  
480 ESI values are in agreement with (Ruggeri et al., 2015; Bakhshi et al., 2018). Among  
481 the three configurations of S2, S2-B (TA +AD) reached the highest ESI value, since the  
482 energy consumption due to the pre-treatment was the lowest (Figure 4). The percentage  
483 contribution in energy consumption of TH, TA and OZ were 77.22 %, 60.33 % and  
484 87.53 % respectively and the detailed energy for S2 A, B, C are reported in Table 5.

485

486 **Figure 4.** Relative amounts of energy consumed for pre-treatments (black) and AD  
487 (grey) processes



488

489

490

**Table 5.** Details of energy assessment

	TH	THA	O
<b>Q c[MJ/d]</b>	104.57	83.76	209.30
Qreq for treatment [%]	58.11	40.87	2.73
Qmix for treatment [%]	0.57	1.42	0.57
Q O <sub>3</sub> for treatment [%]	0.00	0.00	73.61
Q req for AD [%]	5.46	6.81	2.59
Q mix for AD[%]	27.27	40.16	13.62
Q loss for AD[%]	4.50	10.73	4.53
Q O <sub>3</sub> for inoculum of AD [%]	4.09	0.01	2.35
<b>Q p [MJ/d]</b>	218.25	220.31	216.32
Q CH <sub>4</sub> [%]	12.50	15.61	10.64
Q r [%]	87.50	84.39	89.36
<b>ESI</b>	2.09	2.63	1.03

491

#### 492 3.4. Scale-up: economic assessment

493 The economic assessment was aimed to detect the minimum plant size that could be

494 profitable comparing the three different scenarios S0, S1 and S2 A, B, C (Figure 5).

495 According to (Eurostat, 2018), 24.99 kg/y PE of sludge were produced in EU28. The  
496 detailed economic assessment (Table 6) proved for S0 a partial economic profitability at  
497 500,000 PE and complete profitability after 1,000,000 PE with NPV equal to 19.05 M€,  
498 ROI 67.69 % and payback time of 5 years. According to (Arnò et al, 2017), the  
499 economic profitability should be reached at 50,000 PE combining AD of WAS and  
500 organic fraction of municipal solid waste. For S1 a partial economic profitability was  
501 achieved after 5-years amortisation after 1,000,000 PE, but  $ROI < 0$ ,  $NPV < 0$  and  
502 payback time  $>20$  y were obtained. Hence, for S1 the economic sustainability was not  
503 reached.

504 For S2, the minimum plant size to reach the economic sustainability was equal to  
505 50,000 PE for S2-A and 20,000 PE for S2-B, whereas S2-C didn't reach the economic  
506 sustainability. Considering S2 proposed configurations, S2-B reached the best  
507 performances with ROI equal to 45.16 %, NPV 0.21 M€ and payback time after 3 years.

508 No data are available for economic assessment of low temperature AD in literature,  
509 nevertheless the economic profitability for 50,000 PE for S2-B exhibited the same order  
510 of magnitude of mesophilic AD of WAS (Rosa et al., 2018; Zhang et al, 2019).

511 **Table 6.** Economic assessment of the three investigated scenarios (wV = working volume, V = volume)

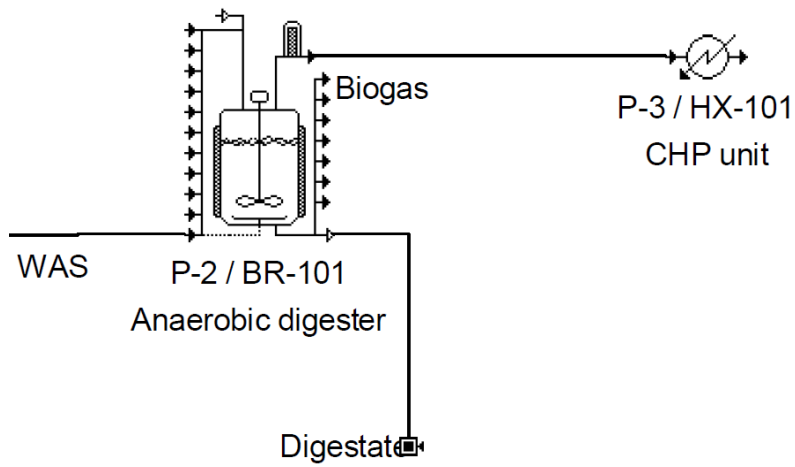
		PE	5.000	10.000	20.000	50.000	100.000	200.000	500.000	1.000.000
		wV unit pre-treatment [m <sup>3</sup> ]	0.34	0.68	1.36	3.40	6.79	13.59	33.96	679.29
		V unit pre-treatment [m <sup>3</sup> ]	0.42	0.85	1.70	4.25	8.49	16.98	42.46	849.11
		AD reactor wV [m <sup>3</sup> ]	5.09	10.19	20.38	50.95	101.89	203.79	509.47	10189.32
		AD reactor V [m <sup>3</sup> ]	6.37	12.74	25.47	63.68	127.37	254.73	636.83	12736.64
<b>S0</b>		Investment cost [M€]	0.03	0.033	0.05	0.09	0.17	0.09	0.59	3.8
		Operational cost [M€]	0.05	0.07	0.07	0.15	0.17	0.08	0.08	0.81
		Revenues [M€]	<0	<0	<0	<0	0.003	0.03	0.1	2.62
		Profitability first 5 years	<0	<0	<0	<0	-0.21	-0.74	-0.11	0.99
		Profitability after 5 years	<0	<0	<0	<0	-0.17	-0.05	0.012	1.81
		ROI [%]	<0	<0	<0	<0	1.74	29.32	16.44	67.68
		NPV [M€]	<0	<0	<0	<0	<0	<0	<0	19.05
		Payback time [y]	>20	>20	>20	>20	>20	>20	>20	5
<b>S1</b>		Investment cost [M€]	0.03	0.03	0.05	0.095	0.17	0.095	0.59	3.88
		Operational cost [M€]	0.05	0.07	0.07	0.15	0.17	0.082	0.09	0.81
		Revenues [M€]	-0.003	-0.004	-0.004	-0.003	-0.008	0.010	0.065	2.10
		Profitability first 5 years	-0.08	-0.11	-0.13	-0.25	-0.34	-0.16	-0.62	-2.59
		Profitability after 5 years	-0.05	-0.075	-0.079	-0.16	-0.17	-0.07	-0.021	1.29
		ROI [%]	<0	<0	<0	<0	<0	<0	<0	<0
		NPV [M€]	<0	<0	<0	<0	<0	<0	<0	<0
		Payback time [y]	>20	>20	>20	>20	>20	>20	>20	>20
<b>S2</b>	TH+AD	Investment cost [M€]	0.03	0.03	0.03	0.07	0.29			
		Operational cost [M€]	0.07	0.08	0.09	0.18	0.22			
		Revenues [M€]	0.02	0.04	0.09	0.22	0.42			

	Profitability first 5 years	-0.06	-0.04	-0.01	0.01	0.13
	Profitability after 5 years	-0.05	-0.03	0.00	0.03	0.19
	ROI [%]	<0	<0	<0	<0	43.53
	NPV [M€]	<0	<0	<0	0.32	2.06
	Payback time [y]	>20	>20	>20	3	8
	Investment cost [M€]	0.03	0.03	0.03	0.07	0.29
	Operational cost [M€]	0.07	0.08	0.09	0.18	0.22
	Revenues [M€]	0.02	0.05	0.11	0.27	0.53
TA+AD	Profitability first 5 years	-0.05	-0.03	0.01	0.07	0.25
	Profitability after 5 years	-0.05	-0.03	0.02	0.08	0.30
	ROI [%]	<0	<0	45.16	99.67	86.32
	NPV [M€]	<0	<0	0.21	0.99	3.55
	Payback time [y]	>20	>20	3	2	2
	Investment cost [M€]	0.03	0.03	0.08	0.16	0.29
	Operational cost [M€]	0.07	0.08	0.09	0.20	0.26
	Revenues [M€]	0.01	0.01	0.03	0.07	0.14
OZ+AD	Profitability first 5 years	-0.08	-0.08	-0.08	-0.16	-0.18
	Profitability after 5 years	-0.07	-0.07	-0.07	-0.13	-0.12
	ROI [%]	<0	<0	<0	<0	<0
	NPV [M€]	<0	<0	<0	<0	<0
	Payback time [y]	>20	>20	>20	>20	>20

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514 **Figure 5.** Process outlines of the considered scenarios: S0, S1, S2: (A) TH + AD; (B)  
 515 TA + AD; (C) OZ + AD



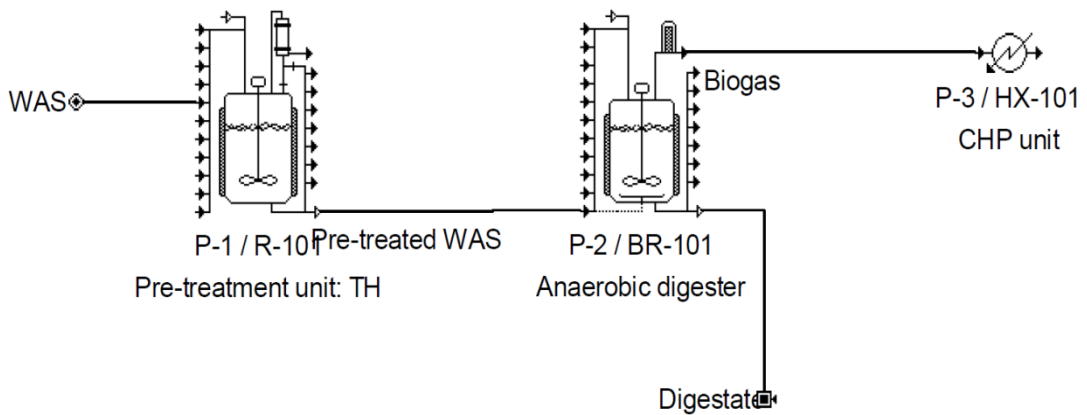
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S0, S1

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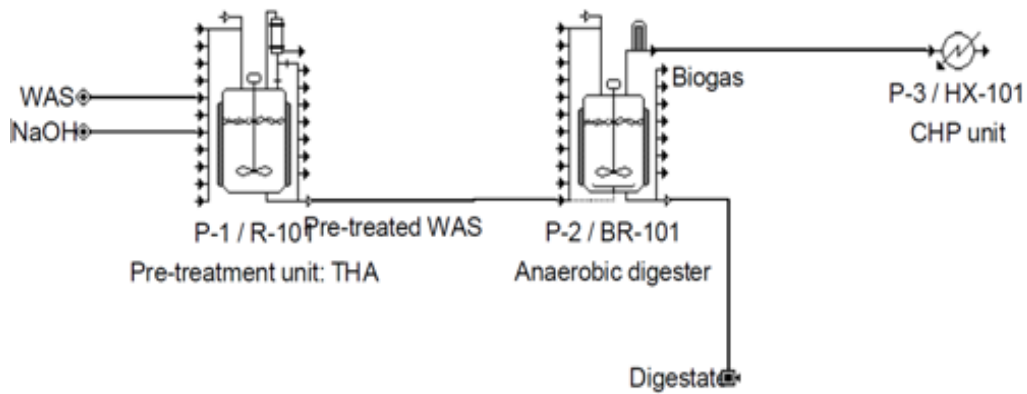
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S2-A

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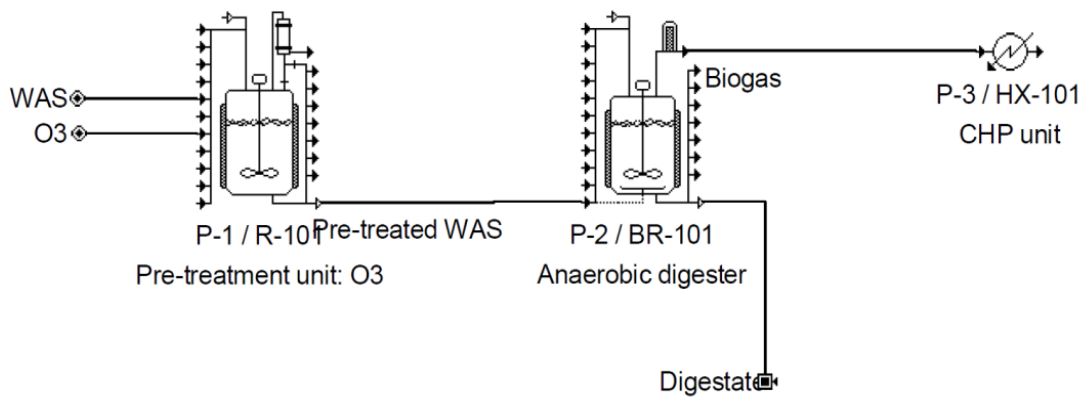
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S2-B



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S2-C

532 **4. Conclusions**

533 This work investigated physic-chemical pre-treatments to improve the anaerobic  
534 digestion of waste activated sludge at 20 °C. Thermo-alkaline pre-treatment, followed  
535 by thermal and ozonation, achieved the best performances, in agreement with literature.  
536 Biogas and methane yields obtained from semi-continuous reactors after thermal and  
537 thermo-alkaline pre-treatments (0.30 - 0.36 m<sup>3</sup> kg<sup>-1</sup> VS, 65 -70 % methane) were  
538 equivalent to literature data referred to mesophilic conditions. The economic assessment

539 of the scale-up of the whole process demonstrated that thermo-alkaline pre-treatment  
540 made AD at 20 °C was economically profitable for WAS generated by a 20,000 PE  
541 WWTP. Anaerobic digestion at 20 °C was demonstrated to have a promising potential  
542 and considering the almost complete lack of literature studies about psychrophilic  
543 processes, further research is urgently needed.

544

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550 Chiappero performed the experimental activity, analyzed the results and wrote the draft  
551 of the manuscript; Francesca Demichelis supported data analysis and performed the  
552 energy and economic assessments; Xuan Lin and Chenxiao Liu contributed to the  
553 experimental activity; Dominic Frigon and Silvia Fiore planned and supervised the  
554 research and revised the manuscript.

#### 555 **References**

- 556 1. Akeberg C, Zacchi G. An economic evaluation of fermentative production of lactic  
557 acid from wheat flour. *Bioresour Technol* 2000;75:119-126.
- 558 2. APHA-AWWA-WEF, 2005. Standard methods for the examination of water and  
559 wastewater, 21st ed. Washington, DC, US.
- 560 3. Appels, L., Degrève, J., Van der Bruggen, B., Van Impe, J., Dewil, R., 2010.  
561 Influence of low temperature thermal pre-treatment on sludge solubilisation, heavy

- 562 metal release and anaerobic digestion. *Bioresour. Technol.* 101, 5743–5748.  
563 <https://doi.org/10.1016/j.biortech.2010.02.068>
- 564 4. Arnò P, Fiore S, Verda V. Assessment of anaerobic co-digestion in areas with  
565 heterogeneous waste production densities. *Energy* 2017;122:221-236.
- 566 5. Arpa, Agenzia Regionale per la Protezione dell’Ambiente.  
567 <http://www.arpa.piemonte.gov.it/approfondimenti/temi-ambientali/rifiuti>; 2017.  
568 Italian.
- 569 6. Bakhshi, Z., Jauffur, S., Frigon, D., 2018. Assessing energy benefits of operating  
570 anaerobic digesters at low temperature with solids pre-ozonation. *Renew. Energy*  
571 115, 1303–1311. <https://doi.org/10.1016/j.renene.2017.08.080>
- 572 7. Barjenbruch, M., Kopplow, O., 2003. Enzymatic, mechanical and thermal pre-  
573 treatment of surplus sludge. *Adv. Environ. Res.* 7, 715–720.  
574 [https://doi.org/10.1016/S1093-0191\(02\)00032-1](https://doi.org/10.1016/S1093-0191(02)00032-1)
- 575 8. Bougrier, C., Albasi, C., Delgenès, J.P., Carrère, H., 2006. Effect of ultrasonic,  
576 thermal and ozone pre-treatments on waste activated sludge solubilisation and  
577 anaerobic biodegradability. *Chem. Eng. Process. Process Intensif.* 45, 711–718.  
578 <https://doi.org/10.1016/j.cep.2006.02.005>
- 579 9. Bougrier, C., Delgenès, J.P., Carrère, H., 2008. Effects of thermal treatments on  
580 five different waste activated sludge samples solubilisation, physical properties and  
581 anaerobic digestion. *Chem. Eng. J.* 139, 236–244.  
582 <https://doi.org/10.1016/j.cej.2007.07.099>
- 583 10. Campo, G., Cerutti, A., Zanetti, M., Scibilia, G., Lorenzi, E., Ruffino, B., 2018.  
584 Enhancement of waste activated sludge (WAS) anaerobic digestion by means of  
585 pre- and intermediate treatments. Technical and economic analysis at a full-scale

- 586 WWTP. J. Environ. Manage. 216, 372–382.  
587 <https://doi.org/10.1016/j.jenvman.2017.05.025>
- 588 11. Carrère, H., Dumas, C., Battimelli, A., Batstone, D.J., Delgenès, J.P., Steyer, J.P.,  
589 Ferrer, I., 2010. Pretreatment methods to improve sludge anaerobic degradability: A  
590 review. *J. Hazard. Mater.* 183, 1–15. <https://doi.org/10.1016/j.jhazmat.2010.06.129>
- 591 12. Dahiya S, Kumar N, Sravan S, Chatterjee S, Sarkar O, Venkata Mohan S. Food  
592 waste biorefinery: Sustainable strategy for circular bioeconomy. *Bioresour Technol*  
593 2018; 248:2-12.
- 594 13. Demichelis, F., Fiore, S., Onofrio, M., 2018a. Pre-treatments aimed at increasing  
595 the biodegradability of cosmetic industrial waste. *Process Saf. Environ. Prot.* 118,  
596 245–253. <https://doi.org/10.1016/j.psep.2018.07.001>
- 597 14. Demichelis, F., Fiore, S., Pleissner, D., Venus, J. 2018b. Technical and economic  
598 assessment of food waste valorization through a biorefinery chain. *Ren. Sus.*  
599 *Energy Rev.*, 94, 38-48
- 600 15. Dev, S., Saha, S., Kurade, M.B., Salama, E., El-Dalatony, M.M., Ha, G., Chang,  
601 S.W., Jeon, B., 2019. Perspectives on anaerobic digestion for biomethanation in  
602 cold environments. *Ren. Sus. Energy Rev.*, 103, 85-89
- 603 16. Dolejs, P., El tayar, G., Vejmelkova, D., Pecenka, M., Polaskova, M., Bartacek, J.,  
604 2018. Psychrophilic anaerobic treatment of sewage: biomethane potential, kinetics  
605 and importance of inoculum selection. *J. Clean. Prod.*, 199, 93-100
- 606 17. Ennouri, H., Miladi, B., Diaz, S.Z., Güelfo, L.A.F., Solera, R., Hamdi, M.,  
607 Bouallagui, H., 2016. Effect of thermal pretreatment on the biogas production and  
608 microbial communities balance during anaerobic digestion of urban and industrial  
609 waste activated sludge. *Bioresour. Technol.* 214, 184–191.

- 610 <https://doi.org/10.1016/j.biortech.2016.04.076>
- 611 18. Eurostat, 2018. Sewage sludge production and disposal from urban wastewater (in  
612 dry substance (d.s)) [WWW Document]. URL  
613 <http://ec.europa.eu/eurostat/tgm/table.do?tab=table&plugin=1&language=en&pcod>  
614 [e=ten00030](http://ec.europa.eu/eurostat/tgm/table.do?tab=table&plugin=1&language=en&pcod) (accessed 5.21.18).
- 615 19. Ferrer, I., Ponsá, S., Vázquez, F., Font, X., 2008. Increasing biogas production by  
616 thermal (70 °C) sludge pre-treatment prior to thermophilic anaerobic digestion.  
617 *Biochem. Eng. J.* 42, 186–192. <https://doi.org/10.1016/j.bej.2008.06.020>
- 618 20. Gomec, C.Y., 2010. High-rate anaerobic treatment of domestic wastewater at  
619 ambient operating temperatures: A review on benefits and drawbacks. *J. Environ.*  
620 *Sci. Heal. - Part A Toxic/Hazardous Subst. Environ. Eng.* 45, 1169–1184.  
621 <https://doi.org/10.1080/10934529.2010.493774>
- 622 21. IPCC. Intergovernmental panel on climate change.  
623 [http://www.ipcc.ch/publications\\_and\\_data/publications\\_and\\_data\\_reports.shtml](http://www.ipcc.ch/publications_and_data/publications_and_data_reports.shtml);  
624 2017 [accessed 21/11/2018]
- 625 22. Istat, 2018. Istituto Nazionale di Statistica [WWW Document]. URL  
626 <https://www.istat.it/> (accessed 5.16.18).
- 627 23. Jeong, T.-Y., Cha, G.-C., Choi, S.S., Jeon, C., 2007. Evaluation of Methane  
628 Production by the Thermal Pretreatment of Waste Activated Sludge in an  
629 Anaerobic Digester. *J. Ind. Eng Chem* 13, 856–863.
- 630 24. Kelessidis, A., Stasinakis, A.S., 2012. Comparative study of the methods used for  
631 treatment and final disposal of sewage sludge in European countries. *Waste Manag.*  
632 32, 1186–1195. <https://doi.org/10.1016/j.wasman.2012.01.012>
- 633 25. Kim, J., Yu, Y., Lee, C., 2013. Thermo-alkaline pretreatment of waste activated

- 634 sludge at low-temperatures: Effects on sludge disintegration, methane production,  
635 and methanogen community structure. *Bioresour. Technol.* 144, 194–201.  
636 <https://doi.org/10.1016/j.biortech.2013.06.115>
- 637 26. Li, H., Li, C., Liu, W., Zou, S., 2012. Optimized alkaline pretreatment of sludge  
638 before anaerobic digestion. *Bioresour. Technol.* 123, 189–194.  
639 <https://doi.org/10.1016/j.biortech.2012.08.017>
- 640 27. Liao, B.Q., Droppo, I.G., Leppard, G.G., Liss, S.N., 2006. Effect of solids retention  
641 time on structure and characteristics of sludge flocs in sequencing batch reactors.  
642 *Water Res.* 40, 2583–2591. <https://doi.org/10.1016/j.watres.2006.04.043>
- 643 28. Mehr AG. 2017. Solar-assisted integrated biogas solid oxide fuel cell SOFC  
644 installation in wastewater treatment plant: Energy and economic analysis. *Applied*  
645 *Energy*;191, 620-638.
- 646 29. Milieu Ltd, WRc, RPA, 2008. Environmental, economic and social impacts of the  
647 use of sewage sludge on land Final Report Part I: Overview Report.
- 648 30. Panepinto, D., Fiore, S., Genon, G., Acri, M., 2016. Thermal valorization of sewer  
649 sludge: Perspectives for large wastewater treatment plants. *J. Clean. Prod.* 137,  
650 1323–1329. <https://doi.org/10.1016/j.jclepro.2016.08.014>
- 651 31. Peters MS, Timmerhaus KD. *Plant Design and Economic for Chemical engineers.*  
652 4<sup>th</sup> ed. Boston: McGraw-Hill, 2003.
- 653 32. Pommeret A., Yang X., Kwan T., Christoforou E., Fokaides P., Lin CSK. 2017.  
654 Techno-Economic Study and Environmental Assessment of Food Waste based  
655 Biorefinery. Chapter 12. Morone P, Papendiek F, Tartiu VE. *Food Waste Reduction*  
656 *and valorisation.* Cham: Springer, 121-146.
- 657 33. Rajagopal, R., Bellavance, D., Rahaman, M.S., 2017. Psychrophilic anaerobic

658 digestion of semi-dry mixed municipal food waste: For North American context.  
659 Process Saf. Environ. Prot. 105, 101–108.  
660 <https://doi.org/10.1016/j.psep.2016.10.014>

661 34. Rosa, A.P., Chernicharo, C.A.L, Lobato, L.C.S, Silva, R.V, Padilh, R.F.,Borges  
662 J.M. 2018. Assessing the potential of renewable energy sources (biogas and sludge)  
663 in a full-scale UASB-based treatment plant. *Renew. Energy*, 124, 21-26

664 35. Ruffino, B., Campo, G., Cerutti, A., Zanetti, M., Lorenzi, E., Scibilia, G., Genon,  
665 G., 2016. Preliminary Technical and Economic Analysis of Alkali and Low  
666 Temperature Thermo-alkali Pretreatments for the Anaerobic Digestion of Waste  
667 Activated Sludge. *Waste and Biomass Valorization* 7, 667–675.  
668 <https://doi.org/10.1007/s12649-016-9537-x>

669 36. Ruggeri, B., Battista, F., Bernardi M., Fino, D, Mancini, G. 2015 The selection of  
670 pretreatment options for anaerobic digestion (AD): A case study in olive oil waste  
671 production. *Chem. Eng. J.*, 259, 630-639

672 37. Saady, N.M.C., Massé, D.I., 2016. Starting-up low temperature dry anaerobic  
673 digestion of cow feces and wheat straw. *Renew. Energy* 88, 439–444.  
674 <https://doi.org/10.1016/j.renene.2015.11.066>

675 38. Salsabil, M.R., Laurent, J., Casellas, M., Dagot, C., 2010. Techno-economic  
676 evaluation of thermal treatment, ozonation and sonication for the reduction of  
677 wastewater biomass volume before aerobic or anaerobic digestion. *J. Hazard.*  
678 *Mater.* 174, 323–333. <https://doi.org/10.1016/j.jhazmat.2009.09.054>

679 39. Sigma-Aldrich.  
680 <https://www.sigmaaldrich.com/catalog/substance/sodiumhydroxide4000131073211>  
681 [?lang=it&region=IT](https://www.sigmaaldrich.com/catalog/substance/sodiumhydroxide4000131073211?lang=it&region=IT) [last access 22/09]

- 682 40. SNAM, 2018.  
683 <https://www.borsaitaliana.it/borsa/azioni/scheda/IT0003153415.html?lang=it>  
684 [accessed 21/11/2018]
- 685 41. Tyagi, V.K., Lo, S.L., 2011. Application of physico-chemical pretreatment methods  
686 to enhance the sludge disintegration and subsequent anaerobic digestion: An up to  
687 date review. *Rev. Environ. Sci. Biotechnol.* 10, 215–242.  
688 <https://doi.org/10.1007/s11157-011-9244-9>
- 689 42. Uma Rani, R., Adish Kumar, S., Kaliappan, S., Yeom, I.T., Rajesh Banu, J., 2012.  
690 Low temperature thermo-chemical pretreatment of dairy waste activated sludge for  
691 anaerobic digestion process. *Bioresour. Technol.* 103, 415–424.  
692 <https://doi.org/10.1016/j.biortech.2011.09.124>
- 693 43. Valo, A., Carrère, H., Delgenès, J.P., 2004. Thermal, chemical and thermo-  
694 chemical pre-treatment of waste activated sludge for anaerobic digestion. *J. Chem.*  
695 *Technol. Biotechnol.* 79, 1197–1203. <https://doi.org/10.1002/jctb.1106>
- 696 44. Van Lier, J.B., Mahmoud, N., Zeeman, G., 2008. *Anaerobic Wastewater Treatment,*  
697 *Biological Wastewater Treatment: Principles, Modelling and Design.*  
698 <https://doi.org/10.1021/es00154a002>
- 699 45. Wingren A, Galbe M, Zacchi G. Techno-Economic-Evaluation of Producing  
700 Ethanol from Softwood: Comparison of SSF and SHF and Identification of  
701 Bottlenecks. *Biotechnol* 2003;19:1109-1117.
- 702 46. Xu, J., Yuan, H., Lin, J., Yuan, wenxiang, 2014. Evaluation of thermal, thermal-  
703 alkaline, alkaline and electrochemical pretreatments on sludge to enhance anaerobic  
704 biogas production. *J. Taiwan Inst. Chem. Eng.* 45, 2531–2536.  
705 <https://doi.org/10.1016/j.jtice.2014.05.029>
- 706 47. Zhang, H. Lucia Rigamonti, L., Visigalli, S., Turolla, A., Gronchi, P., Canziani, R.

707 2019. Environmental and economic assessment of electro-dewatering application to  
708 sewage sludge: A case study of an Italian wastewater treatment plant. *J.Clean Prod.*  
709 *Accepted, In Pres.* <https://doi.org/10.1016/j.jclepro.2018.11.044>

710 48. Zhen, G., Lu, X., Kato, H., Zhao, Y., Li, Y.Y., 2017. Overview of pretreatment  
711 strategies for enhancing sewage sludge disintegration and subsequent anaerobic  
712 digestion: Current advances, full-scale application and future perspectives. *Renew.*  
713 *Sustain. Energy Rev.* 69, 559–577. <https://doi.org/10.1016/j.rser.2016.11.187>

714

**Table A1.** Results of COD solubilisation tests T1, T2 and T3: determination of the optimum treatment time for thermal pre-treatment. Standard deviation values are given.

Test	Operating conditions		tCOD (g O <sub>2</sub> L <sup>-1</sup> )	sCOD (g O <sub>2</sub> L <sup>-1</sup> )	sCOD <sub>0</sub> / tCOD <sub>0</sub> (%)	(sCOD <sub>t</sub> - sCOD <sub>0</sub> ) /sCOD <sub>0</sub> (%)	DR (%)
	Cooling mode	Treatment time (min)					
T1	Room temperature	0	49.26 ± 0.41	3.98 ± 0.10	8.1 ± 0.2	-	-
		30	-	9.60 ± 0.21	-	141 ± 8	12.4 ± 0.7
		60	-	9.20 ± 0.13	-	131 ± 6	11.5 ± 0.5
		90	-	9.55 ± 0.45	-	140 ± 14	12.3 ± 1.2
		120	-	9.76 ± 0.58	-	145 ± 17	12.8 ± 1.5
T2	Ice bath	0	94.27 ± 0.95	7.47 ± 0.13	7.9 ± 0.2	-	-
		10	-	18.23 ± 0.40	-	144 ± 8	12.4 ± 0.6
		20	-	18.25 ± 0.43	-	144 ± 8	12.4 ± 0.7
		30	-	18.68 ± 0.85	-	150 ± 13	12.9 ± 1.1
		45	-	18.05 ± 0.39	-	142 ± 7	12.2 ± 0.6
		60	-	18.40 ± 0.58	-	146 ± 10	12.6 ± 0.8
		90	-	18.23 ± 0.08	-	144 ± 4	12.4 ± 0.3
		120	-	18.40 ± 0.61	-	146 ± 10	12.6 ± 0.9
T3	Ice bath	0	42.01 ± 0.49	3.31 ± 0.06	7.9 ± 0.2	-	-
		30	-	11.54 ± 0.33	-	249 ± 13	21.3 ± 1.1
		60	-	11.94 ± 0.36	-	261 ± 14	22.3 ± 1.1
		90	-	12.77 ± 0.55	-	286 ± 19	24.5 ± 1.6
		120	-	12.91 ± 0.34	-	290 ± 13	24.8 ± 1.1
	Room temperature	0	42.01 ± 0.49	3.31 ± 0.06	7.9 ± 0.2	-	-
		30	-	11.23 ± 0.19	-	240 ± 9	20.5 ± 0.7
		60	-	12.30 ± 0.36	-	272 ± 14	23.2 ± 1.2
		90	-	12.56 ± 0.53	-	280 ± 19	23.9 ± 1.6
		120	-	12.38 ± 0.50	-	274 ± 18	23.4 ± 1.5

**Table A2.** Results of COD solubilisation tests TA1, TA2 and TA3: determination of the optimum dose and treatment time for thermo-alkaline pre-treatment. Standard deviation values are given.

Test	Operating conditions		pH	Dose (gNaOH 'TS)	g <sup>-</sup> tCOD (g O <sub>2</sub> L <sup>-1</sup> )	sCOD (g O <sub>2</sub> L <sup>-1</sup> )	(sCOD <sub>t</sub> sCOD <sub>0</sub> ) /sCOD <sub>0</sub> (%)	DR (%)
	pH target	Treatment time (min)						
TA1	9	Before NaOH	6.42	0	30.62 ± 2.02	1.65 ± 0.09	-	-
		0	9.14	0.033	-	-	-	-
		30	7.77	0.033	-	9.44 ± 0.09	471 ± 29	26.9 ± 2.1
		60	7.69	0.033	-	9.73 ± 0.12	488 ± 31	27.9 ± 2.2
		90	7.66	0.033	-	10.07 ± 0.33	509 ± 39	29.1 ± 2.6
		120	7.65	0.033	-	10.49 ± 0.16	534 ± 34	30.5 ± 2.4
	10	Before	6.41	0	30.62 ± 2.02	1.65 ± 0.09	-	-
		0	10.20	0.067	-	-	-	-
		30	8.98	0.067	-	10.53 ± 0.33	537 ± 40	30.7 ± 2.7
		60	8.93	0.067	-	10.50 ± 0.30	535 ± 39	30.5 ± 2.6
		90	8.88	0.067	-	10.78 ± 0.17	552 ± 35	31.5 ± 2.5
		120	8.85	0.067	-	10.83 ± 0.12	555 ± 34	31.7 ± 2.4
	11	Before	6.32	0	30.62 ± 2.02	1.65 ± 0.09	-	-
		0	11.15	0.089	-	-	-	-
		30	9.64	0.089	-	10.30 ± 0.11	523 ± 32	29.9 ± 2.3
		60	9.52	0.089	-	10.88 ± 0.24	558 ± 38	31.8 ± 2.6
		90	9.44	0.089	-	10.66 ± 0.43	545 ± 44	31.1 ± 2.9
		120	9.43	0.089	-	10.30 ± 0.35	523 ± 40	29.9 ± 2.7
TA2	9	Before	6.15	0	34.07 ± 0.56	2.49 ± 0.07	-	-
		0	9.12	0.032	-	5.39 ± 0.20	117 ± 11	9.2 ± 0.9
		30	7.81	0.032	-	10.79 ± 0.37	334 ± 20	26.3 ± 1.5
		60	7.80	0.032	-	11.85 ± 0.11	377 ± 13	29.7 ± 0.8
		90	7.72	0.032	-	12.08 ± 0.33	386 ± 20	30.4 ± 1.4
		120	7.66	0.032	-	12.42 ± 0.10	399 ± 14	31.4 ± 0.8
	10	Before	6.14	0	34.07 ± 0.56	2.49 ± 0.07	-	-
		0	10.17	0.058	-	11.29 ± 0.39	354 ± 21	27.9 ± 1.6
		30	8.92	0.058	-	14.23 ± 0.17	472 ± 17	37.2 ± 1.1
		60	8.94	0.058	-	13.55 ± 0.66	445 ± 32	35.0 ± 2.4
		90	8.87	0.058	-	14.92 ± 0.74	500 ± 36	39.4 ± 2.7
		120	8.88	0.058	-	14.90 ± 0.10	499 ± 16	39.3 ± 1.0
	11	Before	6.12	0	34.07 ± 0.56	2.49 ± 0.07	-	-
		0	11.04	0.077	-	11.77 ± 0.22	374 ± 16	29.4 ± 1.1
		30	9.33	0.077	-	14.38 ± 0.28	478 ± 20	37.7 ± 1.3
		60	9.24	0.077	-	15.40 ± 0.84	520 ± 40	40.9 ± 3.0
		90	9.26	0.077	-	17.67 ± 0.55	611 ± 31	48.1 ± 2.2
		120	9.17	0.077	-	17.46 ± 0.41	602 ± 26	47.4 ± 1.8
TA3	9	Before	6.07	0	39.58 ± 0.80	2.87 ± 0.04	-	-
		0	9.34	0.037	-	6.27 ± 0.14	118 ± 7	9.2 ± 0.5
		30	8.18	0.037	-	11.29 ± 0.08	293 ± 6	22.9 ± 0.6
		60	8.03	0.037	-	11.78 ± 0.19	310 ± 9	24.3 ± 0.8
		90	8.04	0.037	-	13.87 ± 0.14	383 ± 8	30.0 ± 0.8
		120	7.94	0.037	-	12.56 ± 0.34	337 ± 14	26.4 ± 1.2
	10	Before	6.07	0	39.58 ± 0.80	2.87 ± 0.04	-	--
		0	10.00	0.052	-	9.96 ± 0.13	246 ± 7	19.3 ± 0.6
		30	8.90	0.052	-	13.48 ± 0.43	369 ± 17	28.9 ± 1.4
		60	8.76	0.052	-	13.43 ± 0.43	367 ± 17	28.8 ± 1.4
		90	8.78	0.052	-	12.42 ± 0.38	332 ± 15	26.0 ± 1.3
		120	8.68	0.052	-	14.53 ± 0.24	406 ± 11	31.8 ± 1.1
	11	Before	5.93	0	39.58 ± 0.80	2.87 ± 0.04	-	-
		0	10.92	0.084	-	12.83 ± 0.03	346 ± 5	27.1 ± 0.6
		30	9.37	0.084	-	15.44 ± 0.39	437 ± 16	34.2 ± 1.4
		60	9.29	0.084	-	16.27 ± 0.47	466 ± 19	36.5 ± 1.6
		90	9.32	0.084	-	16.16 ± 0.52	462 ± 21	36.2 ± 1.7
		120	9.24	0.084	-	16.29 ± 0.21	467 ± 11	36.5 ± 1.1